

# Spatial and Temporal Variation Characteristics of Nitrogen and Phosphorus in the Mudong River Basin of Karst Wetlands in Southwest China

Han Runhan<sup>1</sup>, Pan Linyan<sup>2,3\*</sup>, Dai Junfeng<sup>1,4\*</sup>, Huang Kai<sup>5</sup>, Du Ming<sup>1</sup>, Hu Songli<sup>1</sup>, Zeng Derong<sup>1</sup>

<sup>1</sup>College of Environmental Science and Engineering, Guilin University of Technology, Guilin 541004, Guangxi, China

<sup>2</sup>College of Environment and Resources, Guangxi Normal University, Guilin 541004, China

<sup>3</sup>Guangxi Key Laboratory of Environmental Processes and Remediation in Ecologically Fragile Regions, Guilin China

<sup>4</sup>Guangxi Key Laboratory of Environmental Pollution Control Theory and Technology, Guilin China

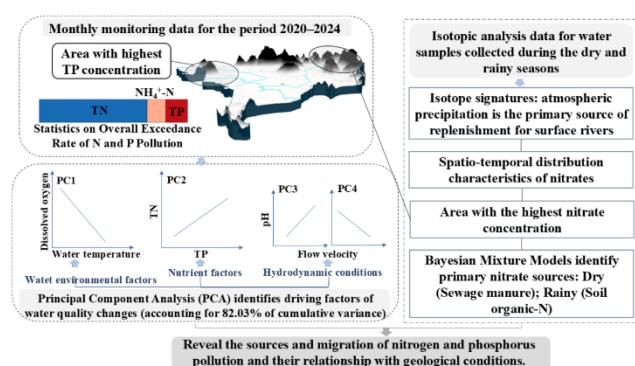
<sup>5</sup>Guangxi Institute of Water Resources Research, Nanning China

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\*to whom all correspondence should be addressed: e-mail: ann-fred@163.com; whudjf@163.com

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## Graphical abstract



## Abstract

The purpose of this study was to analyze the spatiotemporal distribution characteristics and driving factors of nitrogen (N) and phosphorus (P) in surface water of the Mudong River basin in Guilin, through hydrological monitoring, PCA (principal component analysis), and dual isotope tracing of nitrate. The results suggest that (1) Monitoring data show that total nitrogen (TN) is the main pollutant with seasonal variations. The exceedance rate reached a staggering 70.6%. The concentration is usually higher in the dry season. The concentrations of total phosphorus (TP) show a clear spatial, increasing trend from the upstream towards the downstream, with peak concentrations located at the western discharge. (2) As seen in the PCA results, water environmental factors (water temperature and dissolved oxygen), nutrient factors (TN and TP), and hydrodynamic conditions were the three input factors. (3) The Bayesian mixing model revealed a clear seasonal pattern in nitrate sources with a contribution from sewage manure as the

major source during the dry season (37.9%) whereas it turned to soil organic-N in the rainy season (63.1%).

**Keywords** : Nitrogen; Phosphorus; Isotopes; Source Apportionment; Karst

## 1. Introduction

The “kidney” of the earth, wetlands conserve species, the climate and purify water. Nevertheless, human activity threatens wetland ecosystems more than ever and pollution by nitrogen (N) and phosphorus (P) have become a global issue. Excessive nutrients N and P are essential, but they can cause serious consequences such as water eutrophication, algal bloom, ecosystem imbalances (Vitousek *et al.*, 1997; Smith *et al.*, 1999; Xue *et al.*, 2009; Xu *et al.*, 2020). In semi-arid zones and regions with high agricultural activity, this phenomenon has pronounced effects (Liang *et al.*, 2021; Pan *et al.*, 2021).

Nutrient pollution is especially serious in karst areas due to the geology of the region. As precipitation and surface runoff infiltrate underground through cracks and other pathways, attenuation of the nutrients is less (Reimann *et al.*, 2011). Because of this direct pathway, farm and domestic pollutants, such as fertilizer and manure, can rapidly move into groundwater and surface water with little attenuation (Crain, 2010). The karst landscapes in Southwest China, for example, already underwent severe nitrogen loss (Song *et al.*, 2017). Moreover, these dissolution channels can also reduce the hydraulic retention time of pollutants, thus weakening denitrification, organic matter degradation and the like (Heffernan *et al.*, 2012). The way of N and P migration and transformation in karst basins is quite different from non-karst basins because of the unique geohydrological traits

of these basins. There is a shortage of studies that centre on relevant research.

At present, water pollution source tracing is being done by researchers the world over using stable isotope techniques. The use of stable isotope analysis combined with hydrochemical data and mixing models (SIAR/MixSIAR) has become an essential tool for tracing nitrogen sources and identifying transformation processes. In a typical karst wetland in Guilin, China, Li *et al.* (2022) carried out a study, where they measured  $\delta\text{D-H}_2\text{O}$ ,  $\delta^{18}\text{O-H}_2\text{O}$ ,  $\delta^{15}\text{N-NO}_3^-$ , and  $\delta^{18}\text{O-NO}_3^-$  in conjunction with the MixSIAR model. Their work quantitatively apportioned nitrate contributions from chemical fertilizers (33.4%), livestock and domestic wastewater (39.8%), and soil organic nitrogen (26.8%), thereby revealing the quantitative structure of nitrogen sources in the karst area (Li *et al.*, 2022). Similarly, Guo *et al.* (2023), by integrating isotopic data with conventional hydrochemical ions and model analysis, effectively differentiated mixed pollution sources and concluded that nitrate pollution in the studied urban river was primarily derived from domestic wastewater and the nitrification of chemical fertilizers. Multi-isotope-based source analysis of nitrate nitrogen has also been carried out in the typical Huixian Karst Wetland of Guilin, Guangxi. Studies have indicated that artificial  $\text{NH}_4^+$  fertilizers, soil nitrogen, and manure/sewage are the dominant sources of  $\text{NO}_3^-$  in this wetland (Liao *et al.*, 2022). However, most existing research has focused on core wetland or groundwater zones, or has been limited by short monitoring periods. Significant gaps remain in understanding coupled basin-scale dynamics, seasonal variations between dry and rainy periods, the co-behavior of nitrogen and phosphorus, and the integrated analysis of karst hydrodynamics with isotopic tracers.

Based on the aforementioned background, this study, conducted at a typical karst wetland watershed scale, aims to: (1) incorporate monitoring across both dry and rainy seasons to elucidate the spatiotemporal variations in nitrogen and phosphorus concentrations and identify their primary influencing factors; and (2) apply stable isotope techniques to identify pollution sources and, by integrating seasonal variations, reveal shifts in the contributions of different sources between dry and rainy periods, thereby enriching the perspective on seasonal dynamics of non-point source pollution in karst regions.

## 2. Materials methods

### 2.1. Study area

The Mudong River Basin is located in Lingui District, Guilin City, Guangxi Zhuang Autonomous Region, China, and belongs to the irrigation area governed by the West Main Drainage Channel of the Qingshi Lake Reservoir, which is one of the core areas of the Huixian karst wetland. The geographical coordinates of the watershed are  $110^\circ 09'$  to  $110^\circ 14'$  east longitude and  $25^\circ 04'$  to  $25^\circ 08'$  north latitude, and the Mudong River Basin area is about  $31.09 \text{ km}^2$  (Figure 1). The river basin has a subtropical monsoon climate, with an average annual temperature of

about  $20^\circ\text{C}$  and an average annual precipitation of 1,800 mm, with uneven distribution of precipitation, with 75%-80% of the precipitation concentrated in the rainy season from April to September.

The topography of the Mudong River basin is mainly characterized by undulating peaks and valleys and basin plains, with isolated peaks scattered across the area. The northern part is mountainous, while the southern part is relatively flat, showing an overall topographic pattern of higher elevation in the north and lower in the south. The land use in the Mudong River basin is primarily agricultural, with rice cultivation occupying a significant proportion of the total basin area. In addition, residential areas, orchards, grasslands, and a small amount of water bodies are also distributed throughout the basin.

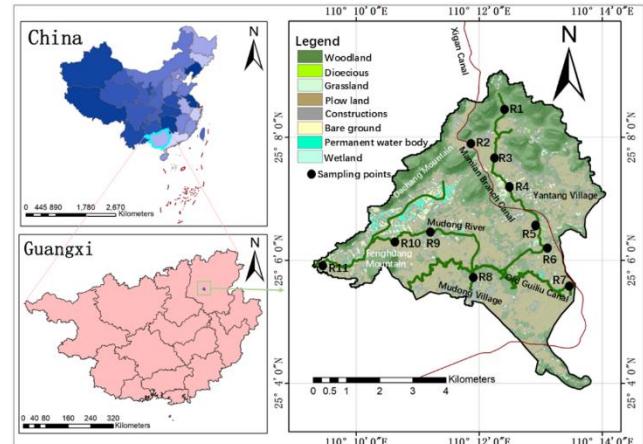


Figure 1. Location map of the study area

### 2.2. Sample collection and testing

Regarding the sampling and monitoring of N and P pollution concentrations, based on the representativeness of the point data and safety portability during sampling, a total of 11 sampling points in the study area were selected, and water samples from 10 of them (no R5 points) were chosen to be sent for testing of hydroxide isotopes and nitrogen and oxygen isotopes. The water samples were collected monthly from May 2020 to September 2024, a total of 28 times. Two groups of water samples were sent for hydroxide isotope and nitrate nitrogen, and oxygen isotope tests during the dry and rainy seasons from 2020 to 2021. The field sampling was operated according to the Technical Specification for Surface Water and Wastewater Monitoring (HJ/T91-2002) and combined with the actual situation on site. Table 1 shows the detection method of each index.

### 2.3. Data processing

In this study, data from both field monitoring and laboratory experiments were preliminarily organized using Excel 2019. Spatial distribution maps and nitrate migration overview maps were generated using ArcGIS 10.2 and its 3D extension module, ArcScene. All other figures were created in Origin 2024. To quantitatively analyze nitrate pollution sources, this study employed surface water nitrate nitrogen and oxygen dual isotope data ( $\delta^{15}\text{N-NO}_3^-$  and  $\delta^{18}\text{O-NO}_3^-$ ). Within the R

programming environment, the *simmr* package was utilized to construct Bayesian mixture models, estimating

the contribution ratios of potential pollution sources during both dry and rainy seasons.

**Table 1.** Table of detection methods for various water quality parameters

Testing program	Detection methods	Main instruments used
TN	Alkaline potassium persulfate digestion UV spectrophotometric method (HJ636-2012)	
NO <sub>3</sub> <sup>-</sup> -N	Ultraviolet spectrophotometric method (HJ / T 346-2007)	
NH <sub>4</sub> <sup>+</sup> -N	Nessler's reagent colorimetric method (HJ535-2009)	
TP	Ammonium molybdate spectrophotometric method (GB11893-89)	UV-6000 UV-Vis spectrophotometer
TDP	Ammonium molybdate spectrophotometric method (GB11893-89)	
δD-H <sub>2</sub> O and δ <sup>18</sup> O-H <sub>2</sub> O	Mass spectrometry	Picarro L1115-i Liquid water and water vapor isotope analyzer
δ <sup>15</sup> N-NO <sub>3</sub> <sup>-</sup> and δ <sup>18</sup> O-NO <sub>3</sub> <sup>-</sup>	The denitrifying bacteria method	Delta V-plus Stable isotope mass spectrometer

**Table 2.** Statistics on the exceedance rates of N and P pollution during the dry season and the rainy season

Indicator	GB3838-2002 III water standard	Dry season exceedance rate (%)	Rainy season exceedance rate (%)	Total exceedance rate (%)
TN	≤1.0	77.7	66.4	70.6
NO <sub>3</sub> <sup>-</sup> -N	≤10	0	0	0
NH <sub>4</sub> <sup>+</sup> -N	≤1.0	15.3	10.4	12.0
TP	≤0.2	10.4	16.5	14.4

### 3. Results and analysis

#### 3.1. Spatio-temporal characteristics of Nitrogen and Phosphorus and identification of driving factors

##### 3.1.1. Spatial and temporal variation characteristics of Nitrogen and Phosphorus

During the monitoring period (2020–2024), the catchment-averaged annual precipitation was 2,604 mm, with the maximum annual total of 3,177 mm recorded in 2024 and the minimum of 2,084 mm in 2023. In accordance with long-term rainfall statistics and the fertilization-irrigation schedule of the dominant crop (rice), the hydrologic year is partitioned into a rainy season (April–September) and a dry season (October–March).

In the study area, the flow velocity of surface water ranged from 0 to 0.88 m/s during the dry season and from 0 to 1.53 m/s during the rainy season. With the exception of site R8, the mean flow velocities at all monitoring points were lower in the dry season than in the rainy season. Minimum flow velocities were predominantly recorded in December, while maximum values generally occurred in May, indicating strong seasonal variability. During the dry season, water temperature and dissolved oxygen (DO) concentration fluctuated between 7.1°C and 33.6°C and 1.2 mg L<sup>-1</sup> and 44 mg L<sup>-1</sup>, respectively. In the rainy season, these parameters ranged from 15.2°C to 40.4°C and 0.3 mg L<sup>-1</sup> to 21.4 mg L<sup>-1</sup>, showing a clear seasonal variation trend. In contrast, pH values were 5.67–8.74 in the dry season and 6.51–8.52 in the rainy season, showing no significant seasonal effect. From June 2023 to September 2024, additional electrical conductivity (EC) measurements were taken across 11 sampling campaigns at the monitoring sections, comprising 2

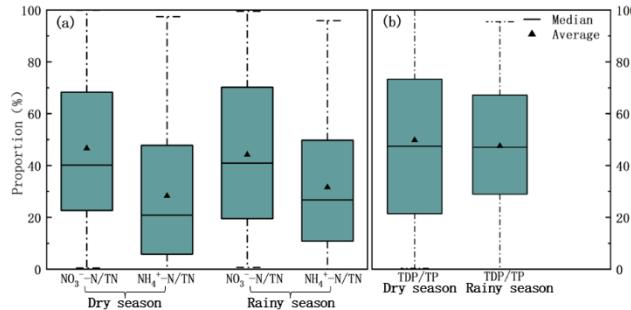
campaigns in the dry season and 9 in the rainy season. The EC values varied from 82.1 μS/cm to 461 μS/cm, with only site R2 exhibiting a notably wider range of variation.

During the monitoring interval, both TN and NO<sub>3</sub><sup>-</sup>-N exhibited pronounced temporal variability. Dry-season concentrations ranged from 0.111–9.496 mg L<sup>-1</sup> for TN and 0.014–6.691 mg L<sup>-1</sup> for NO<sub>3</sub><sup>-</sup>-N, whereas rainy-season values spanned 0.023–7.047 mg L<sup>-1</sup> and 0.005–5.012 mg L<sup>-1</sup>, respectively. Owing to the high dispersion of NH<sub>4</sub><sup>+</sup>-N and TDP data, arithmetic means were deemed non-representative; consequently, median values are reported to characterize the central tendency of these solutes. Median NH<sub>4</sub><sup>+</sup>-N and TDP concentrations were 0.339 mg L<sup>-1</sup> and 0.028 mg L<sup>-1</sup> in the dry season, and 0.360 mg L<sup>-1</sup> and 0.041 mg L<sup>-1</sup> during the rainy period.

In accordance with the Guilin Water Functional Area Planning and referencing the Environmental Quality Standards for Surface Water (GB3838-2002) (Ministry of Environmental Protection of the People's Republic of China, 2002), the concentrations of N and P in the surface water of the Mudong River Basin generally comply with the Class III water quality standards (**Table 2**). During the monitoring period, 70.6% of the surface water samples exhibited TN concentrations exceeding 1.0 mg/L. Notably, the exceedance rate for NH<sub>4</sub><sup>+</sup>-N was 12.0%, significantly higher than that for NO<sub>3</sub><sup>-</sup>-N, which showed no exceedance (0%). Additionally, 14.4% of samples had TP concentrations surpassing the 0.2 mg/L threshold. These results indicate that TN is the primary non-point source pollutant in the river basin.

During the dry season, the mean proportions of NO<sub>3</sub><sup>-</sup>-N and NH<sub>4</sub><sup>+</sup>-N in TN were 46.3% and 28.3%, respectively, while during the rainy season, these values were 44.2% and 31.6%, respectively (**Figure 2a**). NO<sub>3</sub><sup>-</sup>-N constituted

the dominant form of TN loss. The mean proportions of TDP in TP during the dry and rainy seasons were 49.8% and 47.5%, respectively (Figure 2b). TDP represented the predominant form of TP in surface river water, with no significant seasonal variation observed.



**Figure 2.** Ratios of  $\text{NO}_3^-$ -N/TN(a),  $\text{NH}_4^+$ -N/TN (a), and TDP/TP (b) during the dry and rainy seasons

Based on the influences of precipitation, topography, karst development, and Quaternary sediment thickness, the study area was classified into three hydrodynamic zones (Pan *et al.*, 2021): the northern recharge area (R1–R2), the central runoff area (R3–R8), and the western discharge area (R9–R11).

As shown in Figure 3, the mean TN concentration in the northern recharge area ranged from 1.174 mg/L to 2.636 mg/L. The mean  $\text{NO}_3^-$ -N/TN ratios at the two monitoring sites were 70.5% and 40.3%, respectively, which were higher than those at other sites. In the central runoff area, the mean TN concentration varied between 0.948 mg/L and 4.292 mg/L, with the mean  $\text{NO}_3^-$ -N/TN ratio ranging from 24.6% to 69.4%. Among these, site R7 exhibited the lowest ratio. In the western discharge area, the mean TN concentration ranged from 1.329 mg/L to 3.839 mg/L, while the mean  $\text{NO}_3^-$ -N/TN ratio varied between 28.9%

**Table 3.** Results of the Mann-Whitney U test for seasonal differences in water quality parameters

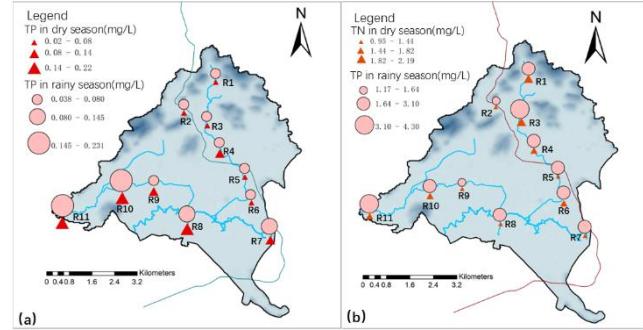
Parameter	U Statistics	Z value	P-value	Significance( $\alpha=0.05$ )	Median trend
Flow velocity	4763.5	-1.711	0.087	NO	No significant difference
Water temperature	1405.5	-15.316	<0.0001	YES	Dry<Rainy
pH	17109	1.792	0.073	NO	No significant difference
DO	17588.5	7.302	<0.0001	YES	Dry>Rainy
TN	21526.5	4.786	<0.0001	YES	Dry>Rainy
TP	15908.5	-0.865	0.387	NO	No significant difference

**Table 4.** Results of the Principal Component Analysis (PCA)

Principal Component	Eigenvalue	Variance Percentage (%)	Cumulative Variance
PC1	1.692	28.21%	28.21%
PC2	1.253	20.89%	49.09%
PC3	1.133	18.88%	67.97%
PC4	0.843	14.05%	82.03%
PC5	0.693	11.55%	93.57%
PC6	0.386	6.43%	100.00%

Calculation results indicate that significant seasonal differences ( $p<0.0001$ ) exist in three parameters: water temperature, DO, and TN. The median water temperature during the rainy season is higher than that in the dry season, directly reflecting significant seasonal temperature changes. In contrast, the DO concentration in the dry season is significantly higher than in the rainy

and 38.4%. The mean TP concentration in the northern recharge area ranged from 0.029 mg/L to 0.045 mg/L. In the central runoff area, the mean TP concentration varied between 0.047 mg/L and 0.185 mg/L, with site R8 showing the highest value. In the western discharge area, the mean TP concentration ranged from 0.079 mg/L to 0.231 mg/L, exhibiting an increasing trend along the flow path and reaching its maximum at the outflow site R11.



**Figure 3.** Spatial variation of TP (a) and TN (b) in surface water during the dry and rainy seasons

### 3.1.2. Statistical test for seasonal variation in water quality

To verify the statistical significance of differences between the dry and rainy seasons, this study employed the Mann-Whitney U test (also known as the Whitney-Wilcoxon test) for statistical analysis. Due to insufficient EC sample numbers, these were excluded from statistical parameters. The statistical measures included the test statistic U value, standardized Z value, and asymptotic significance P value, with the significance level set at  $\alpha=0.05$  (Table 3).

season. This pattern can be attributed to two main factors: higher water temperatures directly limit oxygen solubility, while microbial activity during the rainy season accelerates oxygen consumption. In addition, abundant rainfall and runoff in the rainy season dilute nitrogen-containing pollutants, resulting in significantly higher total nitrogen concentrations in the dry season compared to

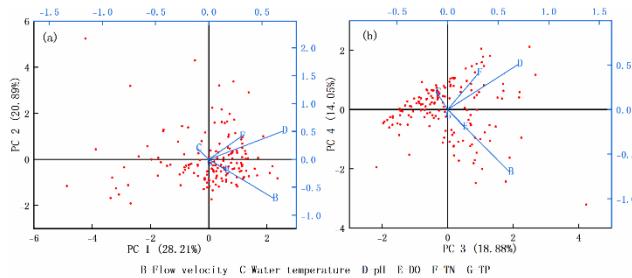
the rainy season. Although the median total phosphorus concentration does not show obvious seasonal variation, its coefficient of variation reaches 110.94% in the dry season, compared to only 89.04% in the rainy season, indicating significant differences. This pattern is generally observed in nutrient parameters (Luo *et al.*, 2024).

### 3.1.3. PCA: identifying driving factors

Based on monitoring and experimental data, this study used Principal Component Analysis (PCA) to explore the correlations among key water quality parameters. These parameters include flow velocity, water temperature, pH, dissolved oxygen (DO), total nitrogen (TN), and total phosphorus (TP).

The eigenvalues of the first three principal components (PC1, PC2, and PC3) are all have eigenvalues greater than 1, collectively explaining 67.79% of the cumulative variance. To more fully reflect the variability of the data, PC4 is included as one of the main driving conditions for water quality changes, contributing to a cumulative variance of 82.03% (**Table 4**).

Principal component (PC1) has high loadings on water temperature (0.647) and DO (-0.607), showing an inverse relationship between these two parameters. An increase in water temperature directly reduces the solubility of oxygen in water and accelerates microbial oxygen consumption. This shows that water temperature is a primary driver of water quality changes and indirectly reveals that water quality exhibits significant seasonal variation. PC2 is composed of nutrient factors TP (0.790) and TN (0.512), which are related to water eutrophication. The most probable causes are fertilization in orchards and farmland within the watershed and the discharge of domestic agricultural wastewater. PC3 and PC4 respectively show a positive correlation between flow velocity and pH value and a negative correlation. High-flow areas promote water re-oxygenation, reduce carbon dioxide concentration, and increase pH; conversely, low-flow areas, due to greater water stability, create a more suitable environment for plant growth. Photosynthesis consumes more carbon dioxide in the water, yet the pH continues to rise (Figure 4).



**Figure 4.** PCA Biplot: PC1 vs. PC2 (a), PC3 vs. PC4 (b)

### 3.2. Transport characteristics of Nitrogen and Phosphorus

### 3.2.1. Hydrogen and oxygen isotope signatures

During the dry season, the range of  $\delta D$ -H<sub>2</sub>O values in surface water was -35.98‰ to -22.97‰ and -37.96‰ to -17.55‰, respectively. The range of  $\delta^{18}\text{O}$ -H<sub>2</sub>O values was -5.72‰ to -1.55‰ and -6.41‰ to -3.18‰, respectively.

Craig (1961) developed the global meteoric waterline (GLMWL) relationship equation:

$$\delta D - H_2O = 8\delta^{18}O - H_2O + 10 \quad (1)$$

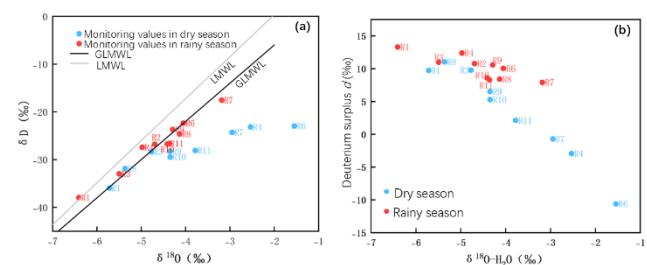
The local meteoric waterline (LMWL) equation for Guilin established by Wu *et al.* (2013) is :

$$\delta D - H_2O = 8.8 \delta^{18}O - H_2O + 17.96 \quad (2)$$

Based on the LMWL and combined with the  $\delta\text{D-H}_2\text{O}$  and  $\delta^{18}\text{O-H}_2\text{O}$  hydrogen and oxygen isotope values of surface water in the study area, the isotope points of both  $\delta\text{D-H}_2\text{O}$  and  $\delta^{18}\text{O-H}_2\text{O}$  are distributed near the LMWL. This indicates that atmospheric precipitation is the primary source of replenishment for surface water (**Figure 5a**).

The deuterium surplus  $d$ , first proposed by Dansgaard (1964), is an important isotopic parameter, and it is believed that there is a correspondence between  $d$  values and local atmospheric precipitation values. Based on this property, the source of recharge of water bodies can be effectively traced. During the water cycle, the  $d$  value usually shows a decreasing trend with the enrichment of  $\delta^{18}\text{O}-\text{H}_2\text{O}$  isotopes, and increases accordingly with the weakening of evaporation intensity. Therefore, the  $d$ -value can be used as a preliminary indicator to determine the degree of isotope exchange and the intensity of evaporation in water bodies.

During the dry and rainy seasons, surface water  $\delta^{18}\text{O}$  values ranged from  $-10.61\text{\textperthousand}$  to  $11.04\text{\textperthousand}$  and  $7.91\text{\textperthousand}$  to  $13.30\text{\textperthousand}$ , respectively. Variations in  $\delta^{18}\text{O-H}_2\text{O}$  values among sampling points resulted in corresponding changes in  $d$  values. As shown in **Figure 5b**, the lowest  $d$  values during the dry season were recorded at points R6 ( $-10.61\text{\textperthousand}$ ) and R7 ( $7.91\text{\textperthousand}$ ). This may be attributed to their location in the Mudong Lake of Huixian wetland, characterized by a large surface area, slow flow velocity, and high evaporation intensity. The highest  $\delta^{18}\text{O-H}_2\text{O}$  value during the dry season was recorded at R8 ( $11.04\text{\textperthousand}$ ), likely because water flow at this point is restricted, the surface is persistently covered by water hyacinth, and evaporation intensity is low.



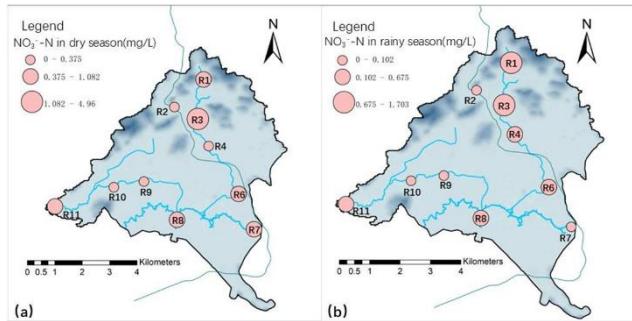
**Figure 5.** Relationship between  $\delta D$  and  $\delta^{18}\text{O}$  (a);  $d$  and  $\delta^{18}\text{O-H}_2\text{O}$  (b) in surface water

### 3.2.2. Spatio-temporal distribution characteristics of nitrates

Rainfall exerts a certain influence on nitrate concentrations (Carey *et al.*, 2014; Husic *et al.*, 2023). Based on the monthly precipitation data for the study

area in 2020, the monitoring period was divided into the rainy season (May–September 2020) and the dry season (October–December 2020). During the dry season, nitrate concentrations ranged from 0.040 mg/L to 5.5 mg/L, showing an overall upward trend as months progressed. During the rainy season, nitrate concentrations ranged from 0.009 mg/L to 3.731 mg/L. Over the course of the months, nitrate concentrations in the water generally increased. The increase in June was slight, likely due to higher rainfall diluting nitrate concentrations in surface water.

Spatially (**Figure 6**), during the dry season, nitrate concentrations at all sampling points except R3 and R11 averaged below 1 mg/L. Among these, R3 recorded the highest nitrate concentration, while R9 had the lowest. During the rainy season, nitrate concentrations at all sampling points except R1 and R3 averaged below 1 mg/L. Point R1 recorded the highest average nitrate concentration, while Point R9 had the lowest. Throughout the monitoring period, nitrate concentrations in upstream Mudong River samples were generally high, whereas those in the middle and lower reaches of Mudong River and in the ancient Guiliu Canal were relatively low.



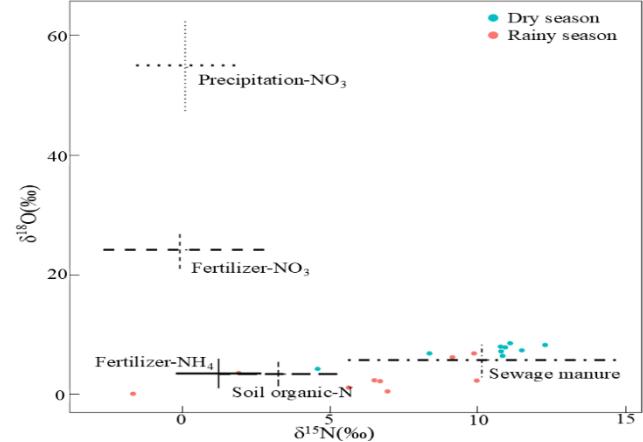
**Figure 6.** Spatial distribution of nitrate concentration during the dry season (a) and the rainy season (b)

### 3.2.3. Quantitative contribution of pollutant sources based on Bayesian Mixture Models

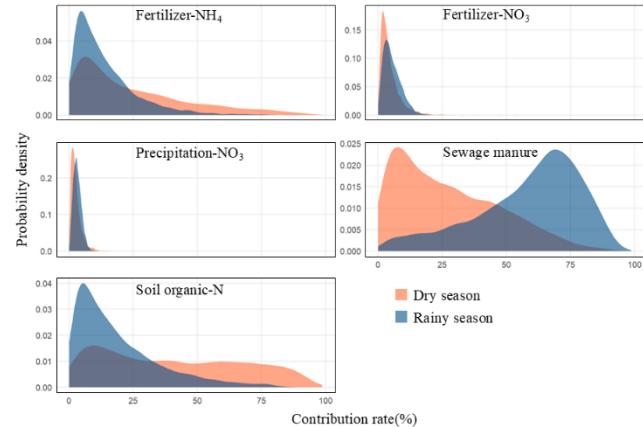
Based on  $\delta^{15}\text{N-NO}_3^-$  and  $\delta^{18}\text{O-NO}_3^-$  nitrogen-oxygen isotope data from surface water in the study area, the variation range of  $\delta^{18}\text{O-NO}_3^-$  is 0.037‰ to 6.796‰, while that of  $\delta^{15}\text{N-NO}_3^-$  ranges from 1.651‰ to 9.969‰. Based on the characteristics of potential pollution sources in the study area (Li *et al.*, 2022), five terminal pollution sources were established: precipitation-NO<sub>3</sub>, fertilizer-NO<sub>3</sub>, fertilizer-NH<sub>4</sub>, soil organic-N, and sewage manure. The model was run using a Markov Chain Monte Carlo

simulation. Convergence diagnostics ( $\text{R-hat} \approx 1.0$ ) confirmed the model's excellent stability.

**Figure 7** displays the distribution of various pollution sources and samples in the  $\delta^{15}\text{N}$ – $\delta^{18}\text{O}$  dual isotope space. The figure reveals that sample points exhibit clustering trends across different seasons: dry-season samples cluster more toward the sewage manure region, while rainy-season samples shift distinctly toward the soil organic-N region. This preliminary finding indicates seasonal variations in pollution sources.



**Figure 7.** Isotope spatial map



**Figure 8.** Posterior probability density distribution of pollution source contribution rate

According to the model output results (**Table 5**), during the dry season, sewage and manure are the main sources (37.9%), followed by soil organic-N (23.6%) and nitrate from precipitation (17.6%). However, in the rainy season, soil organic-N contributes the most (63.1%), far exceeding sewage and manure (13.9%) and nitrate from precipitation (10.4%).

**Table 5.** Comparison of contribution rates of pollution sources during the dry and rainy seasons (median %)

Pollution source	Dry season	Rainy season	Change quantity
Precipitation-NO <sub>3</sub>	17.6	10.4	-7.2
Fertilizer-NO <sub>3</sub>	3.4	4.8	+1.4
Fertilizer-NH <sub>4</sub>	2.2	3.0	+0.8
Soil organic-N	23.6	63.1	+39.5
Sewage manure	37.9	13.9	-24.0

As shown in **Figure 8**, the posterior distribution curves of soil organic-N and wastewater manure exhibit a clear

separation between the dry and rainy seasons, with very little overlap. This indicates a significant difference in their contributions across different seasons.

#### 4. Discussion

The velocity of flow of surface water in the study area was significantly high during the rainy season as compared to the dry season, which exhibit such typical seasonal pattern. Water temperatures were higher much during rainy seasons and concentrations of DO were lower. This was partly due to the fact that raising the temperature of the water decreases the solubility of the oxygen molecule in water. Moreover, the rainfall during the rainy season increased organic matter input to the surface, which in turn increased the dissolved oxygen consumption (Zhang *et al.*, 2024). Demars (2019) also reported this phenomenon.

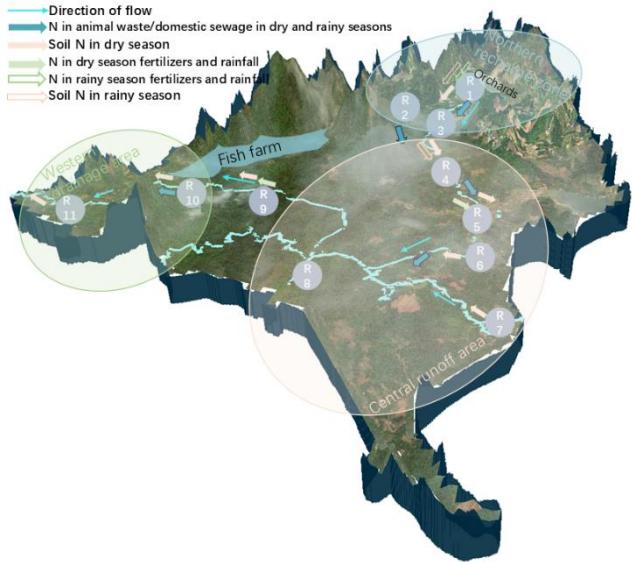
According to the result of study that the total nitrogen (TN) concentration during dry season is far more compared to rainy season. The weakening of dilution caused by heavy rainfalls and runoff in the monsoon season is a plausible reason for loss of nitrogen retention capacity. In terms of TP, there are no significant differences between dry and wet seasons in total phosphorus (TP). However, in dry season, total phosphorus (TP) has a CV of 110.94 percent as compared to 89.04 percent in rainy season. The control of local inputs and outputs is related to this variation. According to earlier studies, the variations of total phosphorus in watersheds are usually tied with a heavy rainfall, runoff, and sediment resuspension (Zhang *et al.*, 2022; Bender *et al.*, 2018). The considerable variability of total phosphorus (TP) in the dry season during this study was related to domestic sewage and agricultural runoff.

Isotope values in water, or  $\delta D\text{-H}_2\text{O}$  and  $\delta^{18}\text{O}\text{-H}_2\text{O}$ , are affected by soil type and vegetation cover at the point where atmospheric precipitation recharges surface rivers and shallow groundwater. As a result, these values vary seasonally and spatially. In the dry months, it is observed that lesser cooling, less rainfall and greater evaporation cause generally higher isotope values at various sampling sites. The Local meteoric waterline (LMWL) of Guilin is also further away from these values. The wet season that is hot, humid, and rainy consists of surface runoff, wetland seepage, and the replenishment of irrigation water. As a result, the abundance value of isotopes of hydrogen and oxygen tend to be closer to the LMWL of Guilin.

R1 and R3 are both located next to orchards with R1 and vegetable fields with R3 which had high average nitrate concentrations among other explanation including application of fertilizer. Water flow velocity near R6 tends to be very low, which results in capturing animal dung and domestic sewage. Thus, average nitrate concentration is high. **Figure 9** shows the movement of nitrate in this watershed.

Seasonal changes in pollution source contributions highlight the impact of surface conditions and hydrological processes on nitrogen movement. There was a remarkable jump in the soil organic-N from 23.6% in dry

season to 63.1% in rainy season giving a total of 39.4%, which shows that the leaching and eroding of nitrogen in the soil due to rainwater is the primary cause of nitrate pollution during the rainy season. Throughout the dry season, sewage manure contributed approximately 37.9% of total nitrogen in the rivers. This value decreased to 13.9% in the rainy season, showing a change of -24.0%. This indicates that non-point source pollution gets diluted during the rainy season. The fraction of precipitation- $\text{NO}_3^-$  was greater in the dry season (17.6%) than in the rainy season (10.4%). This dry season higher fraction might be due to the accumulation with atmospheric particulate matters deposition whereas the rainy season lower fraction may be due to the wet deposition dilution effect.



**Figure 9.** Schematic diagram of  $\text{NO}_3^-$  transport in the surface water of the river basin

Research findings show that nitrogen and phosphorus pollution sources within the study area exhibit seasonal migration characteristics. However, this paper argues that collaborative efforts across different seasons remain key to mitigating pollution. Regardless of the time of year, the region must strengthen sewage pipeline survey and construction supervision, while strictly regulating the wastewater discharge behavior of livestock farmers. Efforts should be made to promote the use of ecological pesticides within the basin to reduce the usage rate of chemical fertilizers and pesticides. The government needs to guide farmers to reduce fertilizer application before heavy rainfall to prevent nutrient loss. In response to the proliferation of water hyacinth and poor water flow at point R8, local governments must regularly carry out cleaning and dredging to prevent algal outbreaks. The agricultural department should guide villagers to implement crop rotation systems to improve soil structure and reduce fixed nitrogen and phosphorus inputs. In an era of rapid technological development, local governments should also rationally utilize technology to advance environmental protection efforts. Nitrate isotope tracing technology and Bayesian models can be used as routine monitoring tools to accurately locate new pollution hotspots through annual tracing analysis and conduct post-treatment assessments.

## 5. Conclusion

This research witnessed nonpoint source pollution of nitrogen and phosphorus in Mudong River basin. This study adopted a long-term monitoring approach involving water quality sampling, PCA and isotopes. TN, TP, and other parameter data which has been collected were analyzed to reach the following findings:

- (1) Water environmental factors in the surface water repository exhibit distinct seasonal regularity, with TN pollution being the most prominent. The water temperature, DO and TN during the dry season were significantly different from that during the rainy season.
- (2) There are generally three factors that account for water quality. These are environmental factors of water (water temp and DO), nutrient factors (TN and TP), and hydrodynamic conditions (flow velocity and pH). Of the water environmental factors PC1 has the greatest effect.
- (3) Nitrate pollution sources vary fundamentally with the seasons. The isotope analysis of hydrogen and oxygen indicates that replenishment in this basin is dominated by atmospheric precipitation. Using dual-isotope tracking and the simulation of a Bayesian mixed model, researchers found that the main source of pollution was sewage manure during the dry season (contribution rate of 37.9%), while this resource turned into soil organic-N during the rainy season (contribution rate of 63.1%). This shows that during the rainfall season, the leaching of soil nitrogen by rainfall runoff within the basin is the primary cause of nitrogen pollution.

This study integrates the characterization of the spatiotemporal distribution of N and P concentrations along with nitrogen-oxygen isotope analysis to clarify N and P pollution sources, migration routes, the relationship with geological conditions. Based on the nitrate nitrogen and ortho phosphate seasonal trends observed during raining and dry season and its value of nitrogen-oxygen isotopes at the different sampling site in the region, it is shown that agricultural fertilizer, domestic sewage and geological condition determine their pollution migration. This study aims to provide more environmentally friendly approaches for future local cultivation, fertilization, and management practices.

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