

# Eco-Friendly Synthesis of Titanium Dioxide Doped Tin-Zirconium Oxide Nanocomposites for the Degradation of Brilliant Green and 4-Nitrophenol

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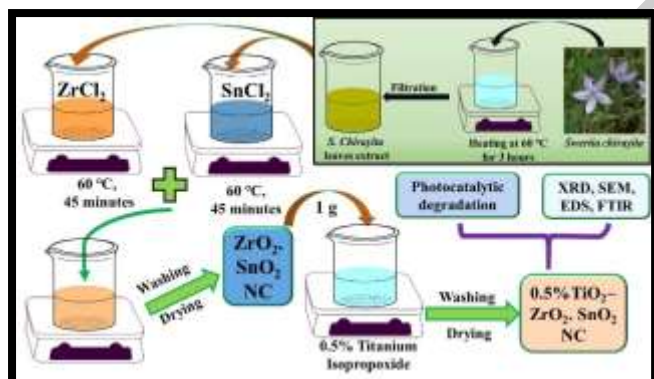
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## Graphical abstract



## Abstract

In this study, a green, cost-effective, and reliable method was employed to synthesize pure and titanium dioxide-doped Tin dioxide-zirconium dioxide nanocomposite ( $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC). The synthesis strategy involves the utilization of leaf extract of the *Swertia chirayita* plant as a stabilizing and capping agent. The surface area of prepared nanocomposites was analyzed through the  $\text{N}_2$  adsorption process, crystalline nature, and other properties of the crystal were investigated by X-ray diffraction (XRD) while scanning electron microscopy (SEM) was used to explore the morphology of pure and  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC. The optical properties of pure and  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC were determined by evaluating the band gap through diffused reflectance (DRS) spectroscopy. The functional groups were studied using Fourier Transform infrared (FTIR)

spectroscopy. The photocatalytic properties of prepared pure and  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC were examined by performing the degradation of brilliant green and 4-nitrophenol. The incorporation of  $\text{TiO}_2$  into  $\text{SnO}_2\text{-ZrO}_2$  led to a decrease in the photocatalytic activity under solar light irradiation.

**Keywords:** Nanocomposite; Zirconia; Titanium; Doping; Characterization.

## 1. Introduction

Water is one of the most important necessities of life, covering 71% of the earth's crust. With a rapid increase in population and industrial activities, global demand for water doubles approximately every 21 years. Water quality is affected by the development and unplanned growth of the industrial sector, therefore, one of the major concerns of modern research is environmental contamination due to the Industrial Revolution (Roushan *et al.* 2016). The root cause of environmental pollution is industrialization which is the key factor for economic development (Malik *et al.* 2014). Human activities such as urbanization, farming, and industries are discharging their wastewater containing organic pollutants in freshwater reservoirs making them unfit for consumption. Improper management of waste materials, produced by different industrial sectors and other anthropogenic activities, is one of the critical issues in developing countries (State 2011).

The textile industry is one of the major contributors to water pollution which discharges 10-25% of textile dyes in

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freshwater reservoirs. The annual production and usage of dyes by textile industry is 1.3 million tons (Chowdhury *et al.* 2019; Wibowo, Safitri, *et al.* 2025). Textile is a globally widespread industry, accounting for 7% of world exports, and generates 1 trillion dollars annually. The industrial sector emerges as one of the largest global polluters due to its consumption of large amounts of fuels and chemicals with 80% of effluents being discharged into water bodies (Lellis *et al.* 2019; Wibowo, Anwar, *et al.* 2025). These residual dyes not only damage the aesthetic beauty of water bodies but also affect the aquatic life. Dyes contain aromatic compounds such as naphthalene and benzamine which are toxic and carcinogenic. In the past, natural dyes were extracted from plants and animals; however, synthetic dyes were discovered in 1856. Complex dye families such as cationic dyes and basic dyes are used in wool, acrylic, and nylon dyeing (Konecoglu *et al.* 2015). Dyes are categorized into direct dyes, vat dyes, basic dyes, reactive dyes, and solvent dyes based on their chromophore structure (Anwer *et al.* 2019).

To treat industrial effluents that contain organic pollutants, different techniques are used. The most commonly employed technologies are adsorption, precipitation, electrochemical technologies, ion exchange membranes, and one most fruitful processes is the photodegradation of organic dyes (Wibowo *et al.* 2023). This process entails exposing the pollutants to UV radiation and converting them to carbon dioxide and water. The development of nanoscience and nanotechnology has captured the interest of modern researchers across various scientific fields. The photocatalytic properties of nanomaterials are enhanced owing to their distinctive characteristics like high adsorption capacity and large surface area, hence, they can be effectively used in the degradation of organic pollutants present in wastewater (Ruan *et al.* 2019). Over the past 2 decades, semiconductors such as  $ZrO_2$ ,  $CeO_2$ ,  $SnO_2$ , as well as  $TiO_2$  have found extensive application in degrading organic pollutants, pesticides, detergents, dyes, and other volatile compounds under the influence of UV radiations (Pouretedal *et al.* 2012). Due to nontoxicity, high stability, and good photocatalytic activity,  $TiO_2$  has been lavishly used for environmental decontamination.  $TiO_2$  absorbs in the near ultraviolet region and has a band gap of 3.2 eV, rendering it unresponsive to the visible region. When either metal or non-metal elements are doped on semiconductors, their photocatalytic activity is enhanced (Sadat *et al.* 2016). The most promising process for wastewater treatment is photocatalysis in which semiconductors are used as photocatalysts and pollutants are degraded by generation of primary oxidant hydroxyl radical. To treat organic wastewater,  $SnO_2$  has become a potential photocatalyst owing to its nontoxic nature, chemical stability, and low cost. Its band gap is 3.6 eV and recombination of electrons and holes occurs rapidly; however, this limitation can be mitigated by utilizing composite semiconductors and doping techniques (Haq, Rehman, *et al.* 2020). The photocatalytic performance of mixed metal composite,  $SnO_2$ - $ZnO$ , is better than pure  $SnO_2$  (Haq *et al.* 2022; Haq, Rehman, *et al.* 2021; Haq, Shoukat, *et al.* 2020).  $ZrO_2$  is an n-type semiconductor whose band

gap is 5.0 eV and holds excellent technological significance. It is extensively utilized in the fields of extraction, catalysis, solid electrolytes, and gas sensors because of its low toxicity and high extraction efficiency as well as adsorption capability (Haq, Afsar, Ali, *et al.* 2021; Haq, Afsar, Din, *et al.* 2021). Some other principal properties of zirconia include high hardness, good frictional behavior, high density, low thermal conductivity, temperature capability up to 2400°C, electrical insulation, and wear resistance (Bona *et al.* 2015).

Depending on the requirement and appropriateness, a variety of physical and chemical methods can be used for the synthesis of nanocomposites. Physical methods like mechanical milling, plasma, aerosol, ablation, electrodeposition, and arc discharge along with chemical methods involving hydrothermal, microemulsion, sol-gel, and sonochemical processes are commonly employed for the fabrication of nanocomposites (Karak 2019). A simple salt-assisted combustion approach was used for the synthesis of  $Nd_2Sn_2O_7$ - $SnO_2$  nanocomposite which is a highly active photocatalyst used in degrading methyl orange dye (Sadat *et al.* 2016). A facile hydrothermal process was employed to prepare Cerium zirconium oxide nanocomposite ( $Ce_xZr_yO_2$ ) that was used for the remediation of sulfonamide pollutants (Li *et al.* 2019). Thus study was planned to develop cost-effective, sustainable, and environmentally benign approaches for the synthesis of nanomaterials used in wastewater treatment. Conventional synthesis methods for photocatalysts often involve hazardous chemicals, high energy consumption, and produce toxic by-products, which limit their practical applicability, especially in developing regions. To address these challenges, our study focuses on the green fabrication of  $TiO_2@SnO_2$ - $ZrO_2$  NC using *S. chirayita* plant extract. The primary motivation is to utilize natural resources for the synthesis of advanced inorganic materials that are not only effective in degrading common organic pollutants like Brilliant Green and 4-Nitrophenol but also align with the principles of green chemistry. Furthermore, doping with  $TiO_2$  is intended to enhance photocatalytic efficiency and environmental stability, broadening the scope of application in real-world wastewater remediation under solar light (Haq *et al.* 2018; Salama *et al.* 2018). *S. chirayita* is an annual, branched medicinal herb belonging to the family Gentianaceae. It grows up to 1.5 meters in height and has broad, opposite, and sessile leaves. It was initially introduced in 1839 in Edinburgh pharmacopeias to be used either as a tincture or an infusion. *S. chirayita* is a 1.5 m tall annual branching plant with broad, opposite, and sessile leaves. It is a bitter tonic that is distributed throughout temperate regions extending from Kashmir to Bhutan. It contains natural products such as terpenoids, iridoids, flavonoids, xanthenes, and secoiridoid glycosides (Mahmood *et al.* 2014). It has laxative, febrifuge, anti-inflammatory, and anti-helminthic properties. *S. chirayita* is also used as a chronic ingredient in anticancer drugs, ayurveda health tonics, antidiabetic preparations, skin creams, hair oils soaps, and even in liver tonics (Shrestha *et al.* 2015). Herein, we focus on the green fabrication of  $SnO_2$ - $ZrO_2$  and  $TiO_2@SnO_2$ - $ZrO_2$  NCs using *S. chirayita* for first time as capping agent. The physicochemical properties

of prepared NCs were investigated via XRD, SEM, DRS, and FTIR techniques. The photocatalytic potential of the prepared nanocomposites was examined against brilliant green and 4-nitrophenol under direct solar light illumination.

## 2. Materials and methods

### 2.1. Preparation of plant extract

*S. chirayita* plant was collected from a local area of the Sudhnoti district. It was subsequently washed with distilled water, boiled at 60°C in 1000 mL water for 3 hours, cooled at room temperature, and then filtered. The filtrate was then stored in an air-tight bottle and subsequently used for the biosynthesis of nanocomposites.

### 2.2. Synthesis of $\text{Sn}(\text{OH})_2$

For the synthesis of  $\text{Sn}(\text{OH})_2$ , 0.98 g of  $\text{SnCl}_2$  salt was taken as a precursor and dissolved in 50 mL water, followed by the addition of 30 mL of plant extract as a stabilizing agent. The solution was heated while being stirred for 45 minutes. Subsequently, NaOH was added to adjust pH which resulted in the preparation of  $\text{Sn}(\text{OH})_2$  in the basic medium. The solid precipitated nanoparticles were filtered and washed multiple times using distilled water and ethanol to remove any unwanted organic substances. The washed nanoparticles were then dried in an oven at 90°C yielding solid  $\text{Sn}(\text{OH})_2$  nanoparticles.

### 2.3. Synthesis of $(\text{ZrOH})_2$

$(\text{ZrOH})_2$  was prepared by dissolving 1.28 g of  $\text{ZrCl}_2$  in 50 mL water. To this solution, 30 mL of extract was added, acting as a capping and stabilizing agent. It was then heated under stirring for 45 minutes. Following this, NaOH solution was added dropwise to adjust the pH and to promote the precipitation of  $(\text{ZrOH})_2$ . The solid  $\text{Zr}(\text{OH})_2$  nanoparticles are collected by **filtering** the solution. The nanoparticles are washed several times with **distilled water** and **ethanol** to remove any unreacted precursors, excess NaOH, and any remaining plant extract or organic by-products. After thorough washing, the nanoparticles were dried in an oven to remove moisture to get dried nanoparticles.

### 2.4. Synthesis of $\text{SnO}_2\text{-ZrO}_2\text{ NC}$

The synthesized  $\text{Sn}(\text{OH})_2$  and  $\text{Zr}(\text{OH})_2$  (0.5 g each) were hydrolyzed in 100 mL distilled water. To this solution 0.1 M NaOH was added dropwise to attain pH 10. Mixture was heated as well as stirred for 4 hours and then left for aging for 12 hours at room temperature. The resulting precipitates were washed several times using distilled water, oven-dried at 100°C, and then stored in an air tight glass bottle.

### 2.5. Synthesis of $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2\text{ NC}$

To synthesize  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2\text{ NC}$ , 1 g of the synthesized  $\text{SnO}_2\text{-ZrO}_2\text{ NC}$  was hydrolyzed in 100 mL distilled water to which, the required volume of titanium isopropoxide (for 0.5 % doping) was added that was heated and stirred for 4 hours. The resulting solution was placed under ambient conditions for 12 hours of aging. The precipitates formed were cooled at room temperature, washed several times

using distilled water, oven-dried overnight, and finally stored in an air-tight glass bottle.

### 2.6. Characterization

To evaluate the physicochemical characteristics of prepared samples, various characterization techniques were used. To assess their crystalline properties, X-ray diffraction (XRD) analysis was conducted using the Philips X'Pert model. The crystallite size was calculated by using Debye-Scherrer equation. To examine their microstructure and surface topology, a field emission scanning electron microscope (FE-SEM), specifically the JEOL JSM-5600LV model from Tokyo, Japan, was utilized. The particles size from the SEM images was estimated by using imageJ software. The NOVA Quantachrome version 11.05 was used to study the  $\text{N}_2$  adsorption capacity of the prepared samples at a temperature of 77.35 K followed by a degased process at 300°C. DRS analysis was utilized to examine the light absorbance properties of the samples. Surface functional moieties were analyzed using the Nicolet 560 FTIR spectrometer, covering the range of 4000 to 400  $\text{cm}^{-1}$ .

### 2.7. Photocatalytic activity

Photocatalytic activity of pure and  $\text{TiO}_2$  doped  $\text{SnO}_2\text{-ZrO}_2$  NC was checked against brilliant green and 4-nitrophenol. For each experiment, 20 mg of prepared NCs was added to 50 ml of brilliant green and 4-nitrophenol solutions individually. Prior to sunlight irradiation, the reaction mixture was stirred for 30 min under dark conditions to achieve adsorption-desorption equilibrium. After specific time intervals, the solutions were placed under sunlight, and using a double-beam spectrophotometer, absorbance spectra were recorded to monitor the photocatalytic degradation of brilliant green and 4-nitrophenol.

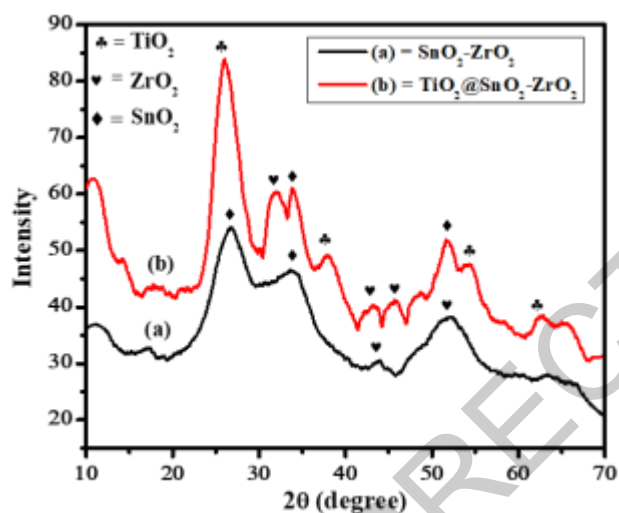
## 3. Results and Discussion

The *S. chirayita* extract acted as a green reducing and stabilizing agent during the synthesis. The extract's rich content of bioactive compound i.e. flavonoids, alkaloids, and xanthenes that reduced the metal precursors to their metallic or oxide nanoparticle forms. The polyphenols and flavonoids in the extract donate electrons to reduce metal ions and subsequently adsorb onto the nanoparticle surface, forming a stabilizing organic moiety. This functionality controls nanoparticle size and morphology and prevents agglomeration, enabling homogeneous dispersion within the nanocomposite. Further, these compounds cap and stabilize the nanoparticles, preventing agglomeration and enhancing uniformity in size and dispersion. Thus, the extract plays a dual role in both the eco-friendly synthesis and the stabilization of the nanocomposites, aligning with principles of green chemistry. It is expected that the residual phytochemicals on NPs surfaces enhance the generation of ROS during photocatalytic process.

### 3.1. XRD analysis

XRD was used to determine the average size of crystals as well as the structural composition of the crystalline phase. **Figure 1** shows the XRD pattern of pure and  $\text{TiO}_2$ -doped  $\text{ZrO}_2\text{-SnO}_2\text{ NCs}$ . The XRD pattern indicates that the particles

have a nanoscale crystallite size, as evidenced by the broad peaks. In XRD spectrum (a), the diffraction bands were observed at  $2\theta$  positions of 26.5, 33.86, and 51.8, corresponding to Miller indices (110), (101), and (211), respectively, indicating the tetragonal geometry of nanosized  $\text{SnO}_2$  particles. These observations aligned with reference card 01-077-0448. Additionally, a set of Bragg reflections appeared at  $2\theta$  positions of 31.9, 34.1, 43.1, 45.1, and 52.1, with corresponding  $hkl$  values of (012), (111), (022), (121), and (113), respectively, suggesting the rhombohedral geometrical shape of  $\text{ZrO}_2$ . These findings were consistent with reference card 00-033-1483. In XRD spectrum b, the distinctive Bragg's reflections at corresponding  $2\theta$  positions for the  $\text{TiO}_2$  doped  $\text{ZrO}_2$ - $\text{SnO}_2$  NCs showed diffraction peaks with  $2\theta$  values of 25.9, 37.86, 54.3, and 62.5, corresponding to Miller indices (101), (004), (211), and (204) respectively. These observations indicated the formation of the anatase phase of titanium dioxide, as per reference card 01-071-1167. The separate sets of diffraction peaks for  $\text{ZrO}_2$ ,  $\text{SnO}_2$ , and  $\text{TiO}_2$  proposed the formation of heterojunction.

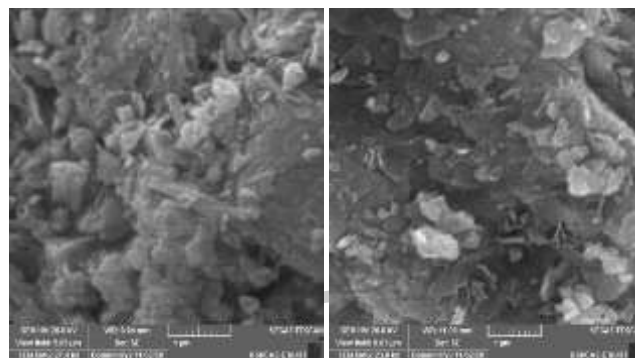


**Figure 1.** XRD spectra of pure and  $\text{TiO}_2$ -doped  $\text{ZrO}_2$ - $\text{SnO}_2$  NCs.

### 3.2. SEM analysis

The SEM micrographs of the prepared pure and  $\text{TiO}_2$  doped  $\text{SnO}_2$ - $\text{ZrO}_2$  NC at various resolutions shown in **Figures 2a and b** demonstrate the morphological features of the sample. In the pure  $\text{SnO}_2$ - $\text{ZrO}_2$  NC (**Figure 3a**), a predominant large compact structure was observed with multiple smaller agglomerates dispersed across its surface. The image shows a mixed morphological shape of the individual particles, where elongated shapes seem dominant. Moreover, the smaller agglomerates were either nearly spherical, elongated, or polyhedral-shaped, which were unevenly distributed and led to the formation of cavities in the sample. The surface of this structure seemed to be non-porous as well as smooth. Similarly, the SEM image of  $\text{TiO}_2$ @ $\text{SnO}_2$ - $\text{ZrO}_2$  NC (**Figure 3b**) displayed a slightly different morphology. Numerous compact structures were obvious where number of cavities had also been decreased. Dispersion of multiple smaller agglomerates over the surface of compact structures was obvious. These agglomerates were multifaceted several of

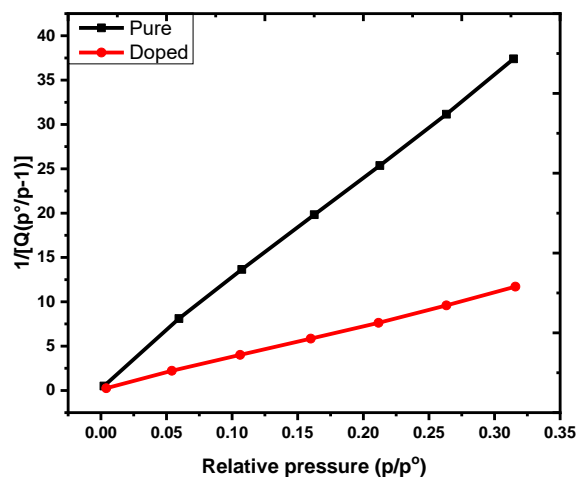
which had distinct boundaries. The size of  $\text{SnO}_2$ - $\text{ZrO}_2$  ranged between 51.68 to 64.38 nm with an average size of 56.09 nm, whereas the size of  $\text{TiO}_2$ @ $\text{SnO}_2$ - $\text{ZrO}_2$  NC lies between 56.85 to 61.68 nm with a mean size of 59.48 nm. The reduction in surface cavities, formation of a more compact structure and increase in average particle size upon  $\text{TiO}_2$  doping shows that the  $\text{TiO}_2$  precursors did not form a uniform dopant layer but instead led to partial aggregation and pore blockage.



**Figure 2.** SEM images of pure (a) and  $\text{TiO}_2$  doped  $\text{ZrO}_2$ - $\text{SnO}_2$  NC (b).

### 3.3. Surface area analysis

The  $\text{N}_2$  adsorption experiment was performed to determine the surface area of the synthesized pure and  $\text{TiO}_2$ -doped  $\text{SnO}_2$ - $\text{ZrO}_2$  NCs, where the adsorption data was analyzed using the BET equation as shown in **Figure 3**. The results show that the adsorption was maximum at zero relative pressure and was found to decrease with increasing relative pressure. The surface area was found to be 96.04 and 29.75  $\text{m}^2/\text{g}$  for  $\text{SnO}_2$ - $\text{ZrO}_2$  and  $\text{TiO}_2$ -doped  $\text{SnO}_2$ - $\text{ZrO}_2$  NC, respectively. Moreover, the surface area of the samples was seen to decrease after doping the  $\text{SnO}_2$ - $\text{ZrO}_2$  NC with  $\text{TiO}_2$ . This decrease might be attributed to the surface coverage of the  $\text{SnO}_2$ - $\text{ZrO}_2$  NC, which preferentially occupies the surface sites (pores), which can reduce the accessibility of the surface area for the adsorbent to adsorb. The insertion of  $\text{TiO}_2$  also induced structural changes in the  $\text{SnO}_2$ - $\text{ZrO}_2$  matrix, which may affect the growth and crystallinity, leading to the formation of larger crystallites that can reduce the surface area-to-volume ratio (Lee *et al.* 2004).

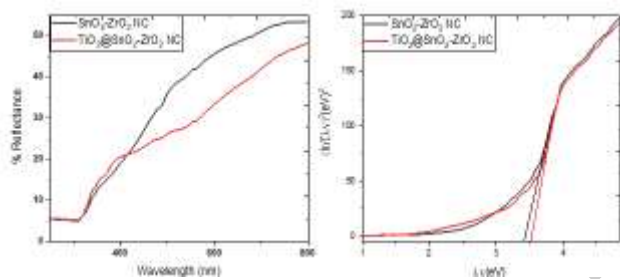


**Figure 3.** BET plots of pure and  $\text{TiO}_2$  doped  $\text{SnO}_2$ - $\text{ZrO}_2$  NC

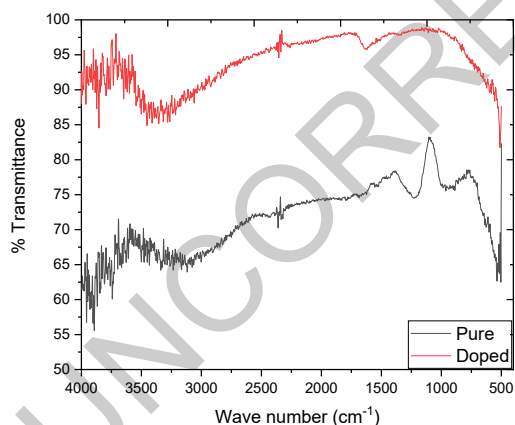


### 3.4. DRS analysis

The DRS spectra of pure and  $\text{TiO}_2$ -doped  $\text{SnO}_2$ - $\text{ZrO}_2$  NCs were recorded within the range of 250-800 nm as depicted in **Figure 4**. DRS spectra indicated the presence of transmittance edge at 300.78 and 294.14 nm. This slight shift of the absorption band to a longer wavelength after doping of  $\text{TiO}_2$  might be due to the quantum size effect and the creation of oxygen vacancies (Kaningini *et al.* 2022). The Tauc plot method was applied to calculate the direct band gap energy of both samples, which are found to be 3.41 eV and 3.47 eV. The  $\text{TiO}_2$  doping leading to an increase in band gap from 3.41 eV to 3.47 eV is a key indicator of a potential phase mismatch. Instead of creating beneficial narrowness electronic states within the band gap, the incorporation of  $\text{TiO}_2$  may have formed a separate, poorly integrated phase or induced lattice strain. This can create energy barriers at the interfaces between Sn-Zr oxide and  $\text{TiO}_2$ , hindering the smooth transfer of photogenerated charge carriers i.e. electrons and holes and promoting their recombination before they can participate in surface reactions (Azam *et al.* 2012).



**Figure 4.** DRS spectra and the Tauc plot for the direct band of pure and  $\text{TiO}_2$ -doped  $\text{ZrO}_2$ - $\text{SnO}_2$  NCs.



**Figure 5.** FTIR spectra of pure and  $\text{TiO}_2$ -doped  $\text{ZrO}_2$ - $\text{SnO}_2$  NC.

### 3.5. FTIR analysis

FTIR spectra of prepared pure and  $\text{TiO}_2$ -doped  $\text{ZrO}_2$ - $\text{SnO}_2$  NCs are depicted in the **Figure 5**. A broad band appeared in the range of 3500-3000  $\text{cm}^{-1}$  corresponding to stretching vibrations of the O-H group present in water molecules and surface metal hydroxide bonds (Aghabeygi *et al.* 2018). The band obtained at 2335  $\text{cm}^{-1}$  was due to surface-absorbed carbon dioxide molecules during the synthesis (Chikkanna *et al.* 2019). The peak observed at 1628  $\text{cm}^{-1}$  in spectra of the pure composite was ascribed to bending

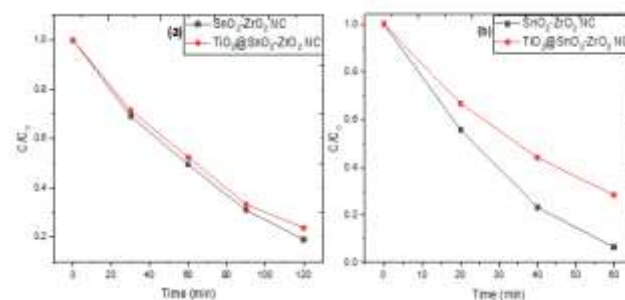
vibrations of OH while the band at 895  $\text{cm}^{-1}$  was accredited to the Sn-O-Zr bond (Qiu *et al.* 2015). Moreover, the absorption band obtained at 626  $\text{cm}^{-1}$  was a characteristic band of tin oxide (Maeda *et al.* 1995). The band observed at 532  $\text{cm}^{-1}$  corresponding to the terminal oxygen vibration of  $\text{SnO}_2$  material while in the titanium dioxide-doped composite, the band obtained at 1622  $\text{cm}^{-1}$  indicated the bending vibration of the Ti-OH bond (Zhao *et al.* 2018). The bands detected at 1346  $\text{cm}^{-1}$  and 586  $\text{cm}^{-1}$  were attributed to the Zr-O bond (Singh *et al.* 2015).

### 3.6. Photocatalytic activity

Pure and  $\text{TiO}_2$ -doped  $\text{SnO}_2$ - $\text{ZrO}_2$  NCs were used to photocatalytically degrade synthetic organic pollutants i.e., brilliant green and 4-nitrophenol. The prepared photocatalysts (20 mg) were added to 50 mL of brilliant green and 4-nitrophenol solutions separately, and then aluminum foil was used to cover the beakers. To attain adsorption-desorption equilibrium, the solution was initially stirred using a magnetic stirrer for 30 minutes in the dark, after which, it was exposed to sunlight between 11 am to 2 pm (18 June 2024). The color of the solution faded with the passage of time suggesting the degradation of the functional moiety responsible for light absorption. After a specific time interval, the progress of the reaction was monitored via a spectrophotometer.

#### 3.6.1. Degradation Profile

The photo-induced degradation of both pollutants was constantly examined by a double beam spectrophotometer and the decrease in the absorbance maxima was noted as function of time interval as shown in **Figure 6 (a and b)**. A gradual decrease in the absorbance maxima was observed, suggesting the degradation of the organic pollutants. The results reveal a sharp decrease in the absorbance maxima of the 4-nitrophenol as compared to brilliant green with respect to irradiation time. This shows the high degradation potential of the prepared nanocatalysts against 4-nitrophenol as compared to brilliant green.



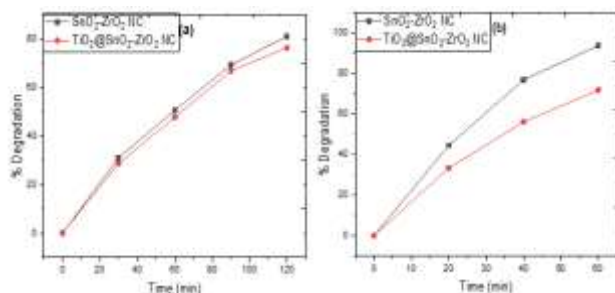
**Figure 6.** Degradation profile of brilliant green (a) and 4-nitrophenol (b).

#### 3.6.2. Percentage degradation

The percentage degradation of both pollutants (**Figure 7**) was determined by using equation 1, where  $C_0$  is the initial concentration while  $C_e$  represents the concentration of pollutants at a specific time. (Bibi *et al.* 2020; Shoukat *et al.* 2019) The results show that 80.9% and 77% of the brilliant green were degraded by  $\text{SnO}_2$ - $\text{ZrO}_2$  NC and  $\text{TiO}_2$ @  $\text{SnO}_2$ - $\text{ZrO}_2$  NC in 120 min, respectively. Likewise, 93.57 and 71.70 % of the 4-nitrophenol were degraded in 60 min in the

presence of  $\text{SnO}_2\text{-ZrO}_2$  NC and  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC, respectively. It is evident from the results that the synthesized nanocatalysts show high efficacy against 4-nitrophenol as compared to the brilliant green. The study demonstrates the potential of  $\text{SnO}_2\text{-ZrO}_2$  NC and  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC as effective catalysts for the degradation of organic pollutants.

$$\% \text{ degradation} = \frac{c_0 - c_e}{c_0} \times 100 \quad (1)$$

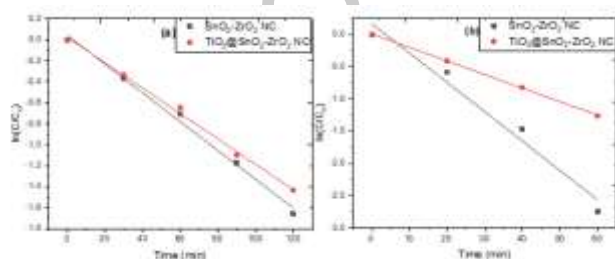


**Figure 7.** Percentage degradation profile of brilliant green (a) and 4-nitrophenol (b).

### 3.6.3. Kinetic study

The kinetic study of the photodegradation reaction of both pollutants was performed by applying equation 2 and the obtained plots are shown in **Figure 8**. The degradation rate constant for  $\text{SnO}_2\text{-ZrO}_2$  NC was found to be  $1.373 \times 10^{-2} \text{ min}^{-1}$  while for  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC, it was  $1.211 \times 10^{-2} \text{ min}^{-1}$  against brilliant green. In the case of 4-nitrophenol, the  $4.56 \times 10^{-1} \text{ min}^{-1}$  and  $2.78 \times 10^{-4} \text{ min}^{-1}$  rate constants were determined for  $\text{SnO}_2\text{-ZrO}_2$  NC and  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC. In the case of both pollutants, the slow degradation rate found for the doped sample suggests the low photocatalytic potential of the catalyst. As it was also assumed from the surface area and band gap analysis, the  $\text{TiO}_2@\text{SnO}_2\text{-ZrO}_2$  NC would have low photocatalytic potential.

$$\ln \frac{c}{c_0} = -kt \quad (2)$$



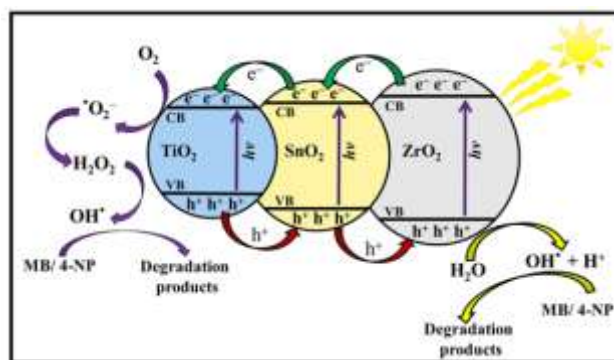
**Figure 8.** Degradation rate constant; (a) brilliant green, b) 4-nitrophenol.

### 3.6.4. Photocatalytic reaction mechanism

When the nanocomposite is exposed to solar or UV light, photons with energy greater than the band gap of the catalyst are absorbed, which leads to the generation of electron-hole pairs in the valance band (VB) of all the three counterparts i.e.,  $\text{SnO}_2$ ,  $\text{TiO}_2$ , and  $\text{ZrO}_2$  of the nanocomposite as shown in **Figure 9**. Once generated, the photoexcited electrons ( $e^-$ ) migrate from the VB to the

conduction band (CB), leaving behind positively charged holes ( $h^+$ ) in the VB. The excited electrons gather in the CB of  $\text{TiO}_2$  whereas the generated  $h^+$  are shifted to VB of  $\text{ZrO}_2$ . The electrons in the CB of  $\text{TiO}_2$  and the holes in the VB are highly reactive and participate in redox reactions with adsorbed species on the catalyst surface. The holes can oxidize organic dye molecules (Dye) adsorbed on the surface, leading to the formation of radical cations ( $\text{Dye}^{+*}$ ). Meanwhile, the electrons in the CB can reduce oxygen molecules ( $\text{O}_2$ ) to form superoxide radicals ( $\text{O}_2^{\cdot-}$ ). The highly reactive radical cations and superoxide radicals generated on the catalyst surface then attack the organic dye molecules adsorbed onto the surface, breaking down their chemical structure into smaller, less harmful molecules, such as carbon dioxide ( $\text{CO}_2$ ) and water ( $\text{H}_2\text{O}$ ). To maintain the photocatalytic activity, minimizing the recombination of electron-hole pairs is essential. Dopants like zirconium dioxide and tin dioxide in the nanocomposite can help to reduce the recombination rate by trapping the photoexcited electrons and holes, thereby prolonging their lifetime.

The variation in the photocatalytic activity of the catalysts might be due to the molecular structure and reactivity of both pollutants. The 4-nitrophenol has a simpler structure with a nitro group attached to the phenol, whereas the brilliant green has a more complex molecular structure with multiple substituted benzene rings. The structural complexity of the pollutants makes them more resistant to degradation. Moreover, the electron-removing nature of the nitro group makes 4-nitrophenol more susceptible to degradation by the reaction of hydroxyl radicals ( $\text{OH}^\cdot$ ).



**Figure 9.** Illustration of photodegradation mechanism of organic pollutants in the presence of pure and  $\text{TiO}_2$ -doped  $\text{ZrO}_2\text{-SnO}_2$  NCs.

### 3.6.5. Comparative analysis

**Table 1** shows the comparative analysis of the photocatalytic performance of our catalyst with values reported in the literature. The metrics presented are the time of reaction and dose of the catalysts required to obtain maximum degradation of the pollutant. Under these standardized criteria, our prepared catalyst demonstrates significantly higher activity, achieving greater degradation in shorter times than most reported catalysts. This enhanced performance is reflected in both the shorter times and low dose to reach to higher degradation percentages.

**Table 1.** Comparative analysis of the-photocatalytic efficacy of the current catalyst with reported data

Catalysts	Dye	Dose (mg/L)	Time (min)	% degradation	References
Sr <sub>0.4</sub> Ba <sub>0.2</sub> Fe <sub>2</sub> O <sub>4</sub>	Atrazine	10	75	86.48	(Yasar <i>et al.</i> 2025)
CuO-NiO-NC	Rhodamine 6G	20	60	80.09	(Lone <i>et al.</i> 2025)
GO-CeO <sub>2</sub>	Methylene blue		120	100	(Fauzia <i>et al.</i> 2024)
ZnO-NPs	Methyl orange	75	250	87	(Mousa <i>et al.</i> 2024)
SnO <sub>2</sub> -ZrO <sub>2</sub> NC	Brilliant green	20	120	80.9	Current study
SnO <sub>2</sub> -ZrO <sub>2</sub> NC	4-nitrophenol	20	60	93.57	Current study

#### 4. Conclusions

The ZrO<sub>2</sub>-SnO<sub>2</sub> and TiO<sub>2</sub>-doped ZrO<sub>2</sub>-SnO<sub>2</sub> NCs were successfully synthesized using an eco-friendly method employing *S. chirayita* extract as a capping and stabilizing agent. XRD analysis demonstrates that ZrO<sub>2</sub>-SnO<sub>2</sub> and TiO<sub>2</sub>-doped ZrO<sub>2</sub>-SnO<sub>2</sub> NCs nanocomposites possess crystalline structures with nanoscale crystallite sizes. DRS analysis showed an increase in band gap energy from 3.42 eV to 3.37 eV which can be attributed to quantum confinement and structural variation. The SEM analysis shows that agglomeration of nanoparticles in both composites with an average size of 56.09 nm for SnO<sub>2</sub>-ZrO<sub>2</sub> and 59.48 nm TiO<sub>2</sub>-doped ZrO<sub>2</sub>-SnO<sub>2</sub> NCs respectively. Moreover, BET analysis was carried out for surface area determination which was found to be 96.04 m<sup>2</sup>/g for SnO<sub>2</sub>-ZrO<sub>2</sub> and 29.75 m<sup>2</sup>/g for TiO<sub>2</sub>-doped SnO<sub>2</sub>-ZrO<sub>2</sub> nanocomposites, respectively. The green synthesized SnO<sub>2</sub>-ZrO<sub>2</sub> demonstrated higher photocatalytic potential for degradation of brilliant green and 4-nitrophenol as compared to doped sample. This boosted activity of SnO<sub>2</sub>-ZrO<sub>2</sub> NC was mainly due to lower band gap energy and higher surface area which cause efficient generation of electron-hole pairs as well as faster adsorption owing to its enhanced surface area. This work offers a green and non-toxic approach utilizing plant extract for the synthesis of nanocomposites. This work can be extended by selectively doping the SnO<sub>2</sub>-ZrO<sub>2</sub> NC with metals to lower its band gap to further improve its photocatalytic performance. Similarly, other plants' extract can also be used to study their influence on morphological variation leading to a variation in photocatalytic potential.

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