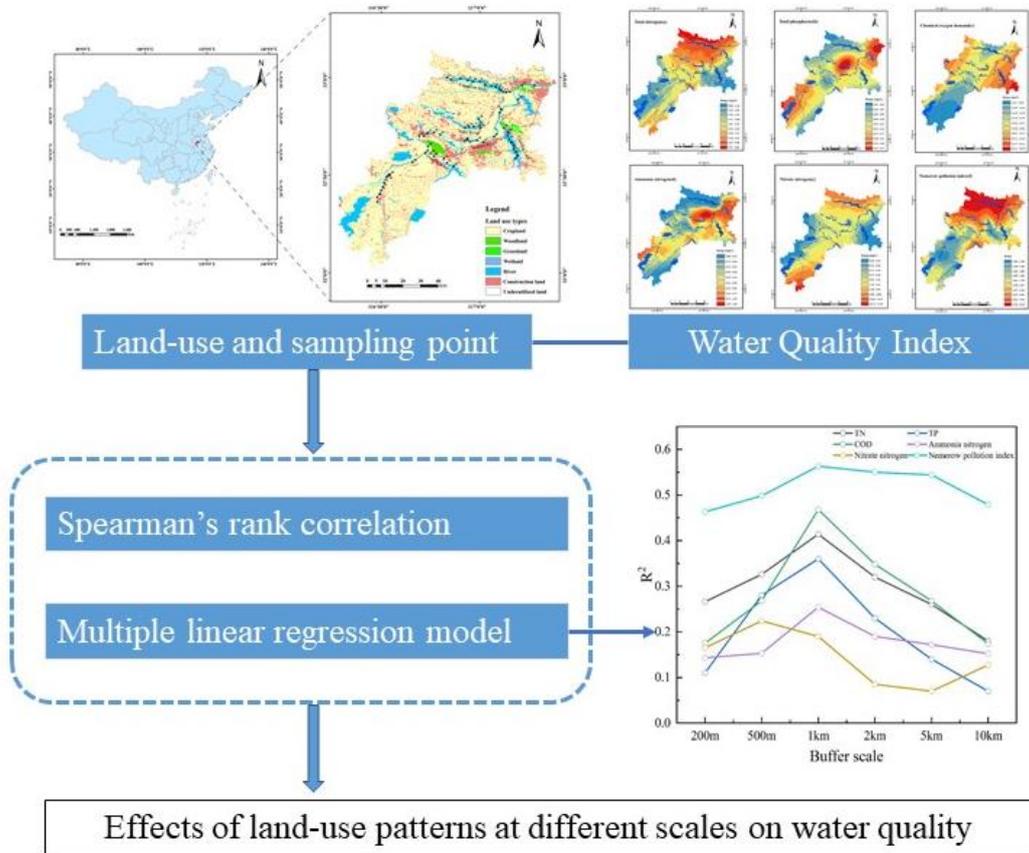


Graphical abstract



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36 **ABSTRACT**

37 The influence of land-use patterns on water quality in the Huaihe River Basin was analyzed using
38 the Nemerow index, correlation and multiple linear regression analysis based on the measured data
39 from 62 sampling sites in the Huaihe River Basin (China). In the point buffer zone, a significant
40 correlation was observed between land-use patterns at different buffer zone scales and water quality
41 parameters. Cultivated land and construction land are significantly positively correlated with all water
42 quality parameters, while forest land and grassland were negatively correlated with all water quality
43 parameters. The 1km buffer zone exhibited the greatest interpretation of TN, TP, COD, $\text{NH}_4^+\text{-N}$ and
44 the Nemerow index, while the 500m buffer zone exhibited the greatest interpretation of $\text{NO}_3\text{-N}$. These
45 findings indicate that forest land and grassland were the main land-use types for the interception and
46 consumption of pollutants. While cultivated land and construction land served as ' source of TN, TP,
47 COD, $\text{NH}_4^+\text{-N}$ and $\text{NO}_3\text{-N}$. These findings show that controlling land-use patterns on a small-scale (\leq
48 1 km) land-use pattern, especially the proportion of cultivated land and construction land, can be an
49 effective tool for the protection of water quality basin areas.

50 **Keywords :** land-use ; spatial scale ; Nemerow index ; correlation ; Huaihe River Basin ;

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1. Introduction

Water is a fundamental resource required for the survival and development of all biological organisms, including humans (Prasad et al., 2024; Venkatraman et al., 2025). In light of the escalating water resource scarcity in the 21st century, it is imperative to take immediate action to safeguard the existing water resources (Maruthai et al., 2025; Selvanarayanan et al., 2024). Rivers are the most directly available freshwater resource on Earth, and river water quality is an important indicator of watershed ecosystem health. River water quality can be affected by various natural and anthropogenic factors, including rainfall, geology, topography, soil type, climate conditions, vegetation, urbanization, and industrial or agricultural activities, particularly within basin areas (Bouchareb et al., 2024; Kamaruddin et al., 2015; Mahdiani et al., 2016). Various studies have shown that changes in land-use type and structure can provide an intuitive reflection of the impact of human activities on the ecological environment, with the corresponding impact on non-point source pollution within basin areas, being a main factor contributing to changes in river water quality (Goodarzi et al., 2023; Hong et al., 2016; Wang et al., 2023a; Zhang et al., 2021). Therefore, it is essential that the relationship between land-use patterns and watershed water quality is examined for effective land-use management planning and the protection of aquatic ecological environments and resources.

In recent years, the rapid development of geographic information system (GIS) and remote sensing (RS) technologies, have resulted in an abundance of watershed studies showing correlations between land-use patterns and river water quality. Various studies have shown that significant correlations exist between land-use patterns and the regional water environment. In terms of research scope, previous studies have mainly covered several zone types, such as the watershed, sub-watershed, river buffer zone and point buffer zone areas. For example, Wang et al (Wang et al., 2024). used

79 redundancy analysis to analyze the relationship between land-use models and river water quality
80 parameters at the river section scale and seasonal scale. Chen et al(Chen et al., 2020). used GIS and
81 regression analysis to explore the relationship between land-use changes and surface water quality
82 indices in the Mitiga Basin. Mei et al(Xiao-Mei et al., 2023). used the investment model to calculate
83 the impact of land-use transformation on water quality in the Ciyao River Basin. Mu et al(Mu et al.,
84 2023). used redundancy, multiple linear regression and Spearman's correlation analysis to study the
85 impact of land-use types at different buffer scales, on water quality in the middle section of the Huaihe
86 River Basin. Tan et al(Tan et al., 2024). used correlation and redundancy analysis to explore the
87 impact of land-use composition on water quality at different spatial and temporal scales.

88 The Huaihe River Basin is located in Eastern China, with the main land-use types being typical
89 agricultural production and industrial development, showing that with the rapid development of a
90 regional social economy, land-use types along the Huaihe River are gradually changing, and the scale
91 effect of land use types on the surface water environment is a critical area of concern for ecological
92 environmental protection.(He et al., 2023; Liang et al., 2021; Wang et al., 2018), which is a critical
93 area of concern in ecological and environmental protection. To ensure the safety of the water
94 environment in the Huaihe River Basin, this study selects the Huaihe River Basin as the research
95 subject. Based on remote sensing imagery, land use types are extracted, and combined with surface
96 water quality monitoring results, the relationship between land use types and the water environment
97 in the Huaihe River Basin is analyzed. The study explores the impact mechanisms of different land
98 use types on the water environment in the Huaihe River Basin, aiming to provide a scientific basis
99 for sustainable watershed management and the construction of an ecological security barrier in the
100 Huaihe River Basin, or similar environments worldwide.2 Materials and methods

101 *2.1 Overview of the study area*

102 The study area was located in the middle section of Huaihe River Basin (Anhui area), which was
103 located between 111° 55 ' -121° 20 ' E and 30° 55 ' -36° 20 ' N, as shown in Figure. 1. This region has
104 an average annual precipitation of 913.6 mm, with rainfall distributed more in the south than in the
105 north, with an uneven spatial and temporal distribution(Xu et al., 2023; Xu et al., 2019b). Within the
106 study area, the north bank of the Huaihe River has a flat region of the Huaibei Plain, while south of
107 the Huaihe River has a hilly terrain. The main land-use type was cultivated land, accounting for 70.84%
108 of the total area of the study area, while forest land, grassland, construction land and water bodies

109 accounted for 0.93 %, 1.14 %, 15.70 % and 9.85 % of the total study area, respectively.

110 2.2 Collection and processing of water quality data

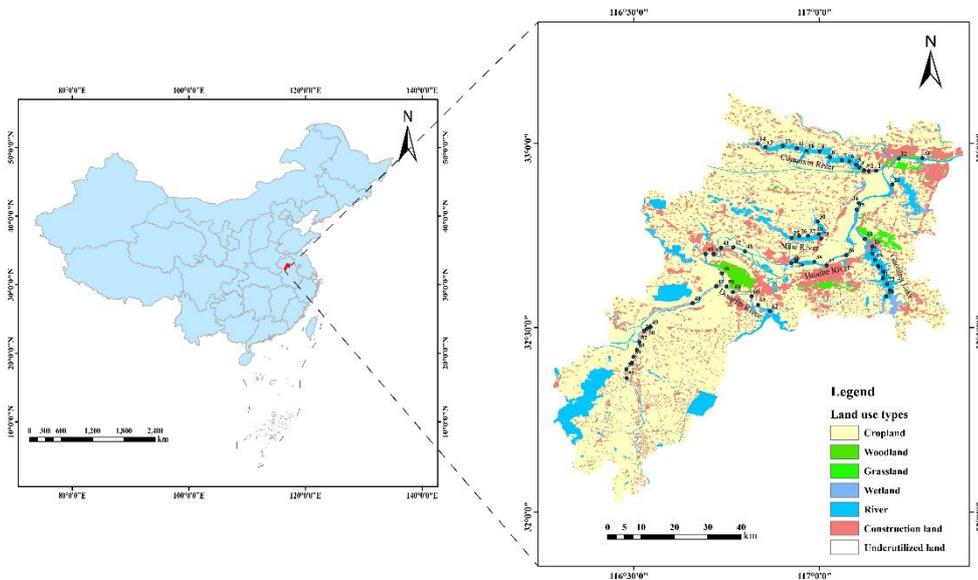
111 In this study, two field surveys were conducted along the main stream and main tributaries of
112 the Huaihe River during the dry season (January 2021), with the distribution of the 62 sampling points
113 shown in Figure. 1. Samples were stored in a high-density polyethylene container and were stored at
114 4 °C. Water quality parameters were measured, such as total nitrogen (TN), total phosphorus (TP),
115 ammonium nitrogen (NH_4^+ -N), nitrate nitrogen (NO_3^- -N) and COD, with the specific determination
116 methods used shown in Table 1. The equipment required for the experiment is listed in Table 2.

117 Table 1 Determination method used for each water quality index parameter

Index	Assay method
TN	Alkaline potassium persulfate digestion UV spectrophotometry (HJ 636-2012)
TP	Ammonium molybdate spectrophotometric method (GB 11893-89)
COD	Potassium dichromate rapid digestion-photometry (HJ 924-2017)
NO_3^- -N	Ultraviolet spectrophotometry (HJ/T346-2007)
NH_4^+ -N	Nessler's reagent spectrophotometry (HJ 535-2009)

118 Table 2 Laboratory Instruments

Serial Number	Instrument	Model	Manufacturer
1	Ultraviolet Spectrophotometer	N5000PLUS	Shanghai Youke Instrument and Meter Co., Ltd.
2	High-pressure Steam Sterilizer	DGL-50B	Jiangsu Dengguan Medical Equipment Co., Ltd.
3	8-Well Microcrystalline COD Digester	SH-108	Jiangsu Haihuan Instrument Equipment Co., Ltd.
4	UPR Ultra Pure Water System	UPI-I-5T	Chengdu Tangning Kangning Technology Development Co., Ltd.



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Figure 1 Map of the study area, showing the land-use map and sampling point distribution

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Previous studies have shown that land-use patterns within the 200 m~ 10 km buffer zone have a significant impact on river water quality(Dai et al., 2024; Xu et al., 2019a; Xu et al., 2021). In order to study the influence of land-use patterns at different buffer zone scales on water quality in the Huaihe River Basin, buffer zones with radii of 200 m, 500 m, 1 km, 2 km, 5 km and 10 km were compared, with each sampling point as the center. The land-use types were divided into six categories, including cultivated land, water area, construction land, wetlands, forest land and grassland. The area of each land-use type within the buffer zone was determined and the relative proportion that each land-use type accounted for was calculated.

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2.3 Analysis methods

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Firstly, the Kriging interpolation was used to construct the spatial distribution map of surface water pollutant concentrations and establish the Nemerow index value for the area. The Nemerow pollution index was used to analyze regional changes in water quality throughout the Huaihe River Basin area. Secondly, Spearman's rank correlation coefficients were determined between the relative

134 proportion of each land-use type area and different water quality parameters at various buffer zone
135 scales. The proportion of land-use type area at different buffer zone scales was used as an explanatory
136 variable and each water quality parameter was used as a response variable for the construction of a
137 multiple linear regression model, the adjusted R^2 value of the model represents the degree to which
138 land use patterns explain the water quality parameters. Using this method, the effects of land-use
139 structure on the water quality of the main stream and main tributaries of the Huaihe River were
140 calculated at different buffer zone scales. Microsoft Excel 2019 and SPSS Statistics 27 were
141 employed for data analysis and processing. Statistical graphics were generated using Origin 2021,
142 ArcMap 10.8, and Adobe Illustrator 2023. Specifically, ArcMap 10.8 was utilized to extract the areas
143 of land use types within different buffer zones and to simultaneously create land use maps and Kriging
144 interpolation maps. Additionally, SPSS Statistics 27 was applied to calculate the Spearman correlation
145 coefficients between each water quality parameter and the proportions of land use types, as well as to
146 construct a multiple linear regression model. The land-use map used in these studies originated from
147 the Resource and Environmental Science Data Center of the Chinese Academy of Sciences
148 (<http://www.resdc.cn>).

149 Constructing a multiple linear regression model of water quality and land use types can help us
150 understand the impact of different land use types on water quality. The model formula is shown in
151 Equations (1):

$$152 \quad Y = a \cdot (\beta_1 X_1 + \beta_2 X_2 + \beta_3 X_3 + \beta_4 X_4 + \beta_5 X_5 + \beta_6 X_6) \quad (1)$$

153 Where, Y represents the measured water quality index in the watershed; a is a constant; $\beta_1 \sim \beta_6$
154 are the correlation coefficients between the area proportions of the five land use types (crop land,
155 woodland, grassland, wetland, river, and construction land) and the water quality index, respectively;

156 $X_1 \sim X_6$ represent the six land use types (crop land, woodland, grassland, wetland, river, and
157 construction land).

158 The Nemerow index method can not only highlight the most serious pollution factors, but also
159 takes into account other evaluation factors that may be more suitable, to a certain extent. In addition,
160 the Nemerow index method can avoid the subjective influence of artificially assigning a weight to
161 each factor during the calculation process (Su et al., 2022a; Su et al., 2022b). In the present study, TN,
162 TP and COD were used to calculate the Nemerow index, according to Equations (2)-(4), as follows:

$$163 \quad I_i = \frac{C_i}{C_{oi}} \quad (2)$$

$$164 \quad \bar{I} = \frac{1}{n} \sum_{i=1}^n I_i \quad (3)$$

$$165 \quad I_p = \sqrt{\frac{I_{i,max}^2 + \bar{I}^2}{2}} \quad (4)$$

166 Where, I_i represents the pollution index of the first evaluation factor; \bar{I} represents the average value
167 of the pollution index of n evaluation factors; I_p represents the Nemerow pollution index value;
168 $I_{i,max}$ is the maximum pollution index value in all pollution evaluation factors; C_i is the measured
169 value of the first evaluation factor; C_{oi} is the water quality standard value for the evaluation factor i
170 (specific grading degrees shown in Table 3) (Wang et al., 2022). This study did not consider the
171 functional classification of surface water quality categories and C_{oi} adopted the 'Surface Water
172 Environmental Quality Standard' (GB 3838-2002) III standard.

173 Table 3 Water quality level based on Nemerow pollution index method

Class of pollution	Nemerow index	Pollution level
1	$I_p \leq 1$	Cleaning
2	$1 < I_p \leq 2$	Light pollution
3	$2 < I_p \leq 3$	Moderate pollution
4	$3 < I_p \leq 4$	Heavy pollution
5	$4 < I_p \leq 5$	Serious pollution

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3 Results analysis and discussion

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3.1 Land-use structure at different buffer zone scales

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For each sampling point, the relative proportion of each land-use type within buffer zone areas

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of 200 m, 500 m, 1 km, 2 km, 5 km and 10 km is shown in Figure 2. The area consisting of river,

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cultivated land and construction land was relatively large, while the area consisting of other land-use

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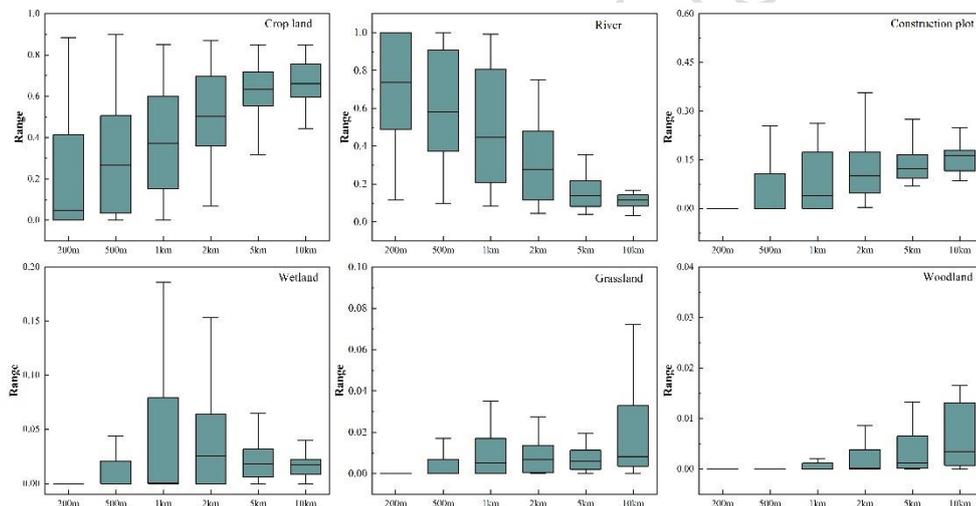
types was relatively small. The sampling points were mainly surrounded by rivers, cultivated land

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and construction land. The proportion of cultivated land and construction land increased at larger

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spatial scales, while the proportion consisting of rivers decreased at larger spatial scales.



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183 Figure 2 The relative proportion of land-use types within the study area at different buffer zone
184 scales

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3.2 Spatial variation in water quality index values

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The descriptive statistics of water quality indicators for the Huai River Basin are presented in

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Table 4. The spatial distribution of TN concentrations across the Huaihe River Basin varied

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significantly. TN concentrations across the study area mainly exhibited a trend of higher

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concentrations in the north of the study area and lower concentrations in the south, while in the central

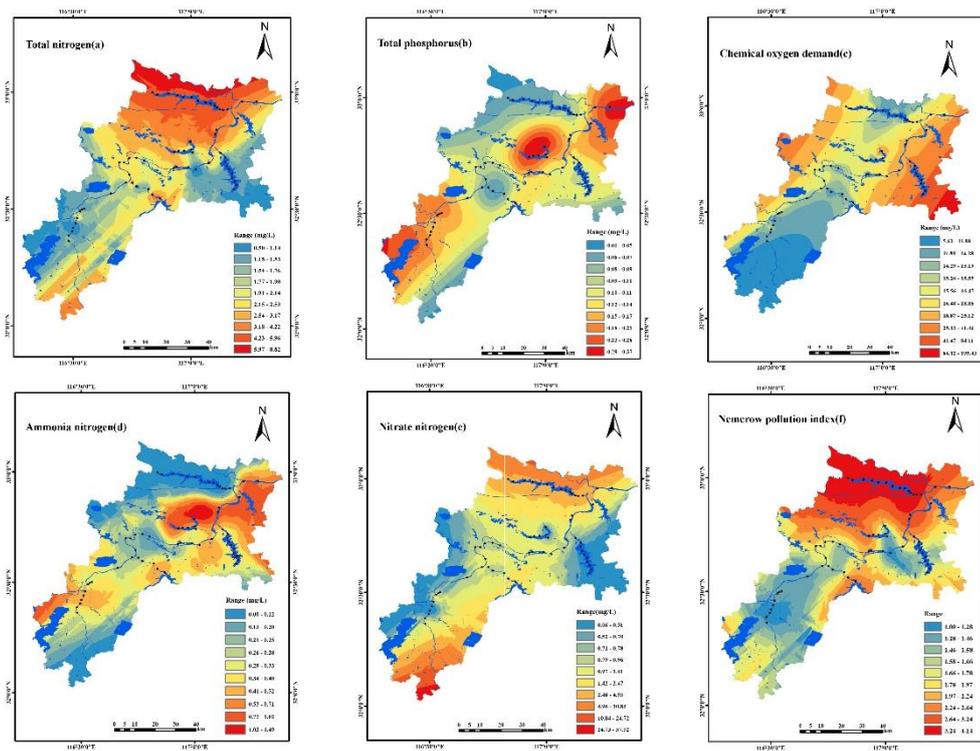
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area along the Huaihe River TN concentrations were in the mid-level (Figure. 3a). As shown in Figure.

191 3b, the TP concentrations within the study area mainly exhibited a trend of higher TP concentrations
 192 in the upper and lower reaches of the study area and lower concentrations in the central region.
 193 Similarly, COD concentrations exhibited an overall trend of higher concentrations in the north of the
 194 study area and lower concentrations in the south (Figure. 3c), while the mass concentration of NH_4^+ -
 195 N mainly exhibited a trend of higher concentrations in the upper and lower reaches of the study area
 196 and lower concentrations in the central region (Figure. 3d). Finally, it can be seen from Figure. 3e,
 197 that the concentration of NO_3^- -N mainly exhibited a trend of higher concentrations in the north and
 198 lower concentrations in the south, with concentrations in the main stream of the central Huaihe River
 199 remaining at a mid-level. Finally, as can be seen from Figure 3f, Nemerow pollution index
 200 concentration also predominantly displays a trend of being higher in the north and lower in the south,
 201 with the central section of the main stem of the Huai River at a moderate level. A comparison of
 202 Figure. 3a and Figure. 3f showed that the Nemerow pollution index distribution was similar to that
 203 of TN, which indicates that TN is an important factor affecting water quality in Huaihe River Basin.

204 Table 4 Descriptive statistics for water quality indicators in the ($\text{mg}\cdot\text{L}^{-1}$)

Parameter	Maximum value	Minimum value	Mean value	Standard deviation
TN ($\text{mg}\cdot\text{L}^{-1}$)	5.46	0.61	2.78	1.52
TP($\text{mg}\cdot\text{L}^{-1}$)	0.72	0.03	0.14	0.12
NO_3^- -N($\text{mg}\cdot\text{L}^{-1}$)	4.72	0.12	1.78	1.54
NH_4^+ -N($\text{mg}\cdot\text{L}^{-1}$)	1.56	0.05	0.33	0.27
COD($\text{mg}\cdot\text{L}^{-1}$)	47	7	19	10



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Figure 3 Distribution of pollutant concentrations and water quality index parameters at sampling points across the Huaihe River Basin study area

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3.3 Watershed pollution index characteristics

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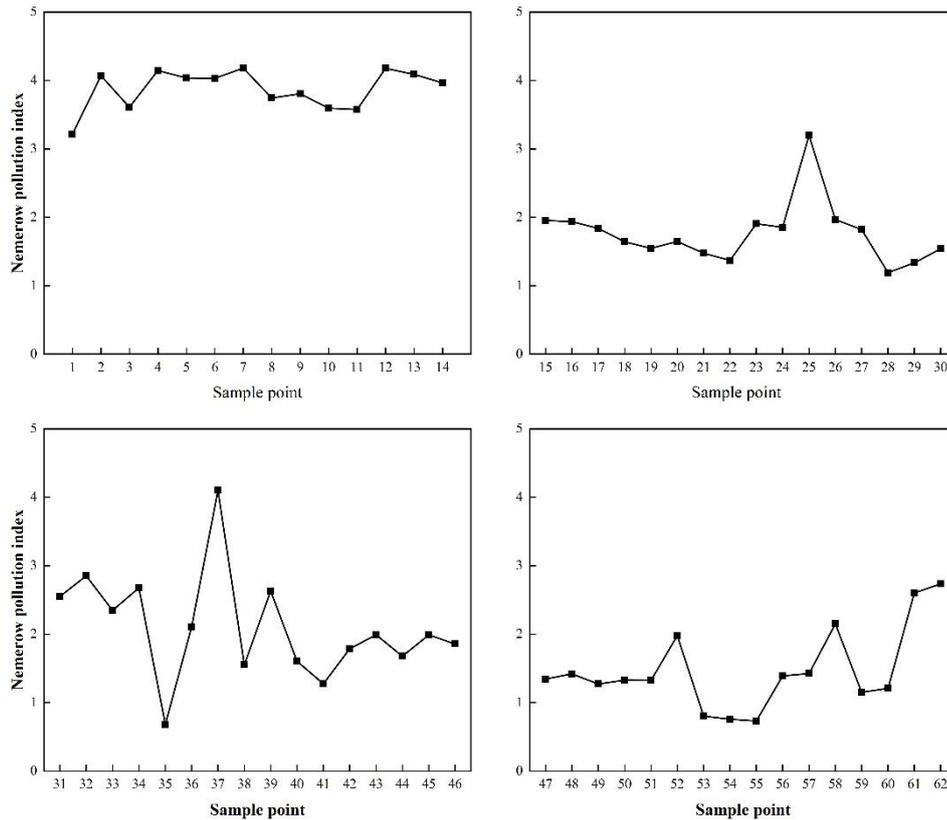
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The distribution of the Nemerow index in the Huaihe River Basin is shown in Figure. 4(The calculation method is shown in Equations (2)-(4)). The mean Nemerow index value across the basin area was 2.3, with a variation range of 0.7~4.2. The Nemerow index values in the Cihuaixin River (the first tributary of the Huaihe River) ranged from 3~4, while in the lower reaches of Huaihe River in the study area ranged from 2~3, the Nemerow index of the upper reaches of Huaihe River in the study area ranged from 1~2 and the Nemerow index of Gaotang Lake and the Nihe River ranged from 1~2. Among these, the Nemerow pollution index values in the lower reaches of the main stream of the Huaihe River fluctuated greatly, while the degree of fluctuation in the upper reaches and tributaries were relatively small.



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Figure 4 Nemerow pollution index value distribution throughout the Huaihe River Basin

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3.4 The correlation between land-use types and water quality in the basin

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At different buffer-zone scales, TN was significantly negatively correlated with cultivated land, while being significantly positively correlated with grassland, wetlands, rivers and construction land (Table 5). Among these, the correlation with grassland within a 1 km buffer zone was the highest ($r = 0.547$, $P < 0.01$). TP was significantly negatively correlated with forest land, wetlands and river, while being significantly positively correlated with cultivated land and construction land, among which the highest correlation was with rivers within a 200 m buffer zone ($r = -0.419$, $P < 0.01$). A significant negative correlation existed between COD and both cultivated land (200 m~1 km) and construction land (2 km~5 km), while significant positive correlations existed between COD and forest land (10 km), wetlands (200 m~500 m) and rivers (200 m~10 km). Among these, the

230 correlation was highest between COD and rivers within the 5 km buffer zone area ($r = -0.416$, $P <$
231 0.01). NH_4^+ -N was significantly negatively correlated with rivers (200 m ~ 2 km) and forest land (5
232 km), while being significantly positively correlated with cultivated land and construction land.
233 Among these, the correlation between NH_4^+ -N and rivers was highest within the 200 m buffer zone
234 area ($r = -0.422$, $P < 0.01$). A significant positive correlation existed between NO_3^- -N and both
235 grassland (500 m ~ 5 km) and construction land (500 m ~ 2 km), with the highest correlation
236 observed with grassland within the 1 km buffer zone ($r = 0.539$, $P < 0.01$). The Nemerow pollution
237 index was significantly negatively correlated with cultivated land and exhibited a high correlation
238 with cultivated land (200 m ~ 500 m) ($r < -0.6$, $p < 0.01$), while being significantly positively
239 correlated with grassland, wetlands, rivers and construction land, with the highest correlation
240 observed with river areas (500 m ~ 5 km) ($r > 0.6$, $p < 0.01$).

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Table 5 Relationships between water quality parameters and land-use types at various buffer

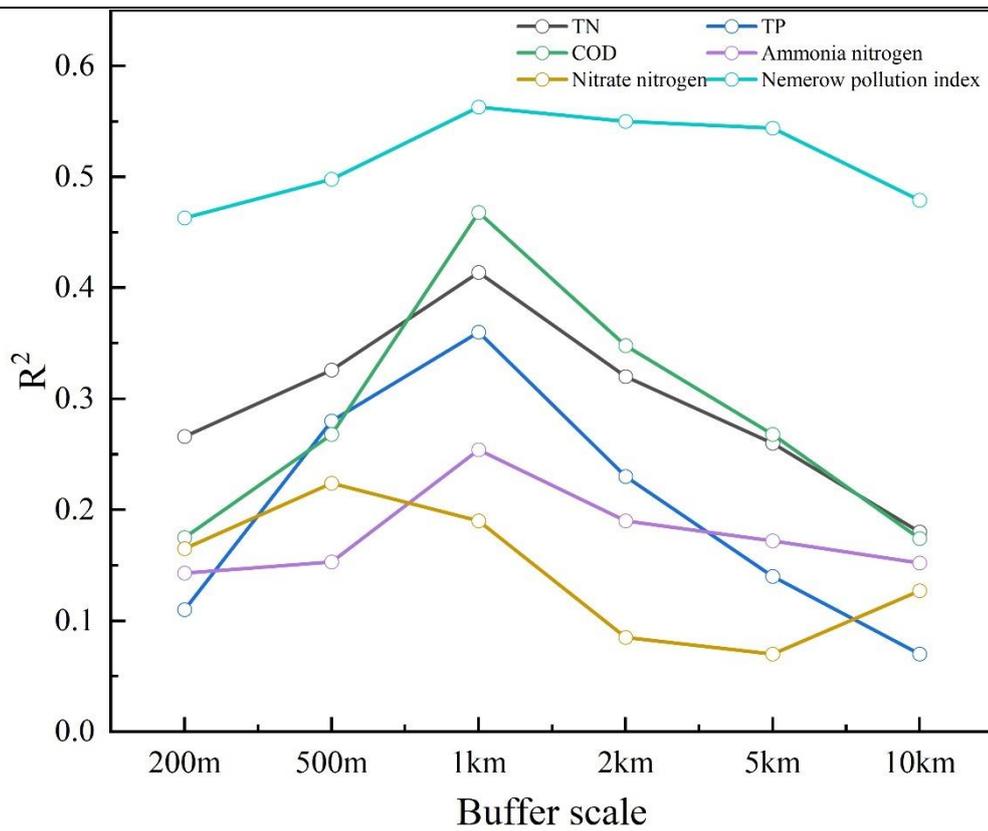
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zone scales, based on Spearman's rank correlation coefficient data

Water quality parameter	Buffer zone scale	Land-use type					
		Cropland	Woodland	Grassland	Wetlands	River	Construction land
TN	200m	0.278*			0.263**	0.266*	
	500m	0.209*		-0.307*	0.270*	0.274*	0.253*
	1km	0.216*		-0.547**		0.277*	0.283*
	2km			-0.510**			0.227*
	5km		-0.269*	-0.539**			
	10km			-0.255*			
TP	200m	0.290*			0.278*	0.419**	0.275*
	500m	0.225*			0.238*	-0.285*	0.266*
	1km	0.203*				-0.220*	
	2km						
	5km		-0.339*				
	10km		-0.301*				
COD	200m	-0.355**			0.362**	0.335*	
	500m	-0.287*			0.302*	0.317*	
	1km	-0.296*				0.355**	
	2km					0.297*	0.297*
	5km					0.331**	0.416**
	10km		-0.317*			0.271*	
NH ₄ ⁺ -N	200m	0.254*				-0.422*	0.314*
	500m	0.240*				-0.316*	0.209*
	1km					-0.208*	
	2km					-0.253*	
	5km		-0.281*			-0.256*	
	10km						
NO ₃ -N	200m						
	500m			-0.337**			0.420**
	1km			-0.539**			0.324*
	2km			-0.500**			0.249*
	5km			-0.475**			
	10km						
Nemerow pollution index	200m	0.635**			0.366**	0.528**	
	500m	0.612**			0.544**	0.639**	0.481*
	1km	0.583**			0.542**	0.675**	0.463*
	2km	0.439**		-0.298*		0.617**	0.431**
	5km	0.331*		-0.446**		0.617**	0.284*
	10km			-0.530**		0.461**	

254 Note: Only explanatory variables with significant correlations were listed. * indicates $P < 0.05$; ** indicates $P <$
255 0.01.

256 Multivariate linear regression analysis showed that the influence of land-use patterns on water
257 quality parameters increased initially and then decreased with increasing buffer zone scale (Figure
258 5). The R^2 values generated from the multiple regression model of TN, TP, COD, NH_4^+ -N and the
259 Nemerow pollution index reached a maximum level within the 1 km buffer zone, reaching 0.414,
260 0.36, 0.468, 0.254 and 0.563, respectively ($P < 0.01$). The R^2 value of the $\text{NO}_3\text{-N}$ multiple linear
261 regression model reached a maximum level within the 500 m buffer zone of 0.224 ($P < 0.01$). The
262 land-use patterns within the large-scale buffer zone area (5 km \sim 10 km) provided a relatively weak
263 explanation for the spatial distribution of water quality parameters. In general, land-use types
264 explained the spatial differentiation of the Nemerow pollution index more than other parameter, while
265 the association between land-use types and $\text{NO}_3\text{-N}$ concentrations was the lowest.



266

267 Figure 5 The explanatory power of different water quality parameters based on the multiple
 268 linear regression model (R^2 value) data for land-use patterns at different buffer zone scales.

269 3.5 Discussion

270 Based on the combined analysis of land-use and pollutant distribution maps (Figure. 1 and Figure.
 271 3), the northern part of the study area (mostly cultivated land and construction land) was shown to be
 272 highly affected by frequent anthropogenic activities, with regional rivers greatly disturbed (Wang et
 273 al., 2023b; Yang et al., 2023). The TN, $\text{NO}_3\text{-N}$ and Nemerow index values were relatively high.
 274 Furthermore, in the dry season, the amount of rainwater replenishment was small in the north of the
 275 study area and the self-purification capacity of water bodies was poor. In contrast, the concentration
 276 of pollutants in forest and grassland areas in the central and southern regions of the study area was
 277 relatively low and the vegetation coverage of forest and grassland areas was higher than other land-
 278 use types. In the dry season, abundant plant root growth reduced pollutant migration through surface

279 runoff due to factors such as plant root absorption and soil retention, effectively intercepting
280 pollutants and reducing their transfer to rivers, reducing the scale of disturbance to the river water
281 body(Wu et al., 2021; Zhang et al., 2023).

282 TN, TP and NH_4^+ -N concentrations were significantly positively correlated with cultivated land
283 within a 1 km buffer zone (Table 5), which was attributed to agricultural activities and the excessive
284 use of pesticides and fertilizers, with the unabsorbed fractions of pollutants flowing into the water
285 body with the slope water, causing a reduction in water quality(Zhang et al., 2022). However, a poor
286 correlation was observed with cultivated land within a larger buffer zone(2km~10km), which may
287 be due to cultivated land accounting for a small proportion of the large-scale buffer zone, while
288 pollutant may also be adsorbed and deposited during transport over a larger scale. COD exhibited a
289 significant negative correlation with cultivated land within a 1 km buffer zone, showing that
290 cultivated land played a role as a COD sink. A significant negative correlation existed between TN,
291 TP, COD, NH_4^+ -N and woodlands within the large-scale buffer zone(5km~10km). There was a
292 significant negative correlation observed between TN, NO_3^- -N and grasslands within a 5 km buffer
293 zone, due to forest land and grasslands being rich in vegetation, reducing pollutant migration through
294 surface runoff and mitigating water pollution through biochemical effects such as vegetation root
295 absorption and soil retention(Li et al., 2019). TN, TP and COD were positively correlated with
296 wetlands in small-scale buffer zones(200m,500m), which may be due to the decomposition of aquatic
297 plants in wetlands during winter, resulting in an increase in pollutant concentrations within the basin
298 area. There was a significant positive correlation between TN, COD and rivers within a 1km buffer
299 zone, which may be due to cultivated land and construction land and other exogenous inputs. TP and
300 NH_4^+ -N were significantly negatively correlated with rivers in small-scale buffer zones(200m~1km),

301 which may be due to the self-purification function of the river itself. TN, TP, COD, NH_4^+ -N and NO_3^-
302 -N were significantly positively correlated with construction land, due to the high-intensity of
303 anthropogenic activities in construction areas, producing domestic sewage and industrial wastewater,
304 thereby affecting water quality(Yao et al., 2023).

305 The Nemerow index was significantly positively correlated with cultivated land within a
306 500m~5km buffer zone, wetlands, water bodies and construction land, indicating that agricultural
307 source pollution, residential sewage discharge, wetland litter degradation and river exogenous
308 pollutant invasion were the main causes of watershed pollution within the study area. There was a
309 significant negative correlation observed with grassland within a 2km~10km buffer zone, indicating
310 that grassland can effectively intercept and absorb pollutants.

311 Multivariate linear regression analysis showed that land-use types had the highest explanatory
312 power for the Nemerow pollution index and the lowest explanatory power for NO_3^- -N pollution.
313 Multivariate linear regression was performed to interpret the effects of land-use patterns on water
314 quality parameters, exhibiting an initially increasing trend, followed by a subsequent decrease. The
315 impact of TN, TP, COD and NH_4^+ -N pollution reached a maximum within a 1 km study area, while
316 the impact of NO_3^- -N reached a maximum at a buffer zone scale of 500 m, which may be due to the
317 relatively singular land-use type within the small-scale buffer zone (200 m, 500 m). Within the large-
318 scale buffer zone (5 km, 10 km), organic pollutants (COD) and inorganic pollutants (nitrogen and
319 phosphorus) were easily absorbed by soil particles, allowing them to be utilized by vegetation or
320 transformed into gas (such as nitrification and denitrification) or other insoluble substances (such as
321 low-solubility or insoluble phosphate) during the transport process. This shows that optimizing small-
322 scale spatial land-use patterns is of high significance to the overall quality of water within the Huaihe

323 River Basin.

324 **4 Conclusions**

325 (1) The surface water environment is closely related to land-use. The concentrations of TN, TP,
326 COD, $\text{NH}_4^+\text{-N}$, $\text{NO}_3\text{-N}$ and the Nemerow index values were significantly positively correlated with
327 cultivated land, wetlands and construction land, while a significant negative correlation was observed
328 with forest land and grassland. In summary, crop land, wetlands, and Construction land are
329 significant contributors to water pollution during winter, whereas woodlands and grasslands exhibit
330 notable mitigating effects on water pollution. Therefore, in the context of land management and water
331 resource protection, rational planning of land use, including increasing the coverage of woodlands
332 and grasslands while curbing the expansion of cultivated land and built-up areas, would be beneficial
333 for enhancing surface water quality.(2) There was a significant correlation observed between land-
334 use patterns and water quality parameters at different buffer zone scales. Among them, land-use
335 patterns within the 1 km buffer zone had the greatest impact of TN, TP, COD, $\text{NH}_4^+\text{-N}$ and the
336 Nemerow index, the R^2 values generated from the multiple regression model reaching 0.414, 0.36,
337 0.468, 0.254 and 0.563, respectively ($P < 0.01$). while land-use patterns within the 500 m buffer zone
338 had the greatest impact on $\text{NO}_3\text{-N}$, the R^2 values generated from the multiple regression model
339 reaching 0.224 ($P < 0.01$). These results highlight the importance of optimizing land-use patterns
340 within a 1 km area, in order to effectively protect water resources within the Huaihe River Basin.

341 Taking the Huai River Basin as the study area, this research preliminarily analyzes the impact of
342 land use types on water quality within circular buffer zones, which may provide certain reference
343 value for land use planning and water quality protection in the basin. Future studies could further
344 integrate hydrological models and ecosystem models to simulate the responses of the water
345 environment under different land use scenarios, offering more precise decision-making support for
346 watershed management. Additionally, interdisciplinary collaboration should be strengthened to
347 comprehensively consider the impacts of climate change and socio-economic factors on land use and
348 the water environment, thereby constructing a more holistic watershed management framework.

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350 Funding

351 This research was supported by Anhui Provincial Key Research and Development Project(The
352 Ecological Environment Special Project) (NO.202004i07020012)

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