

1 **Current Status and Trends of Municipal Sludge Treatment and Disposal in China**

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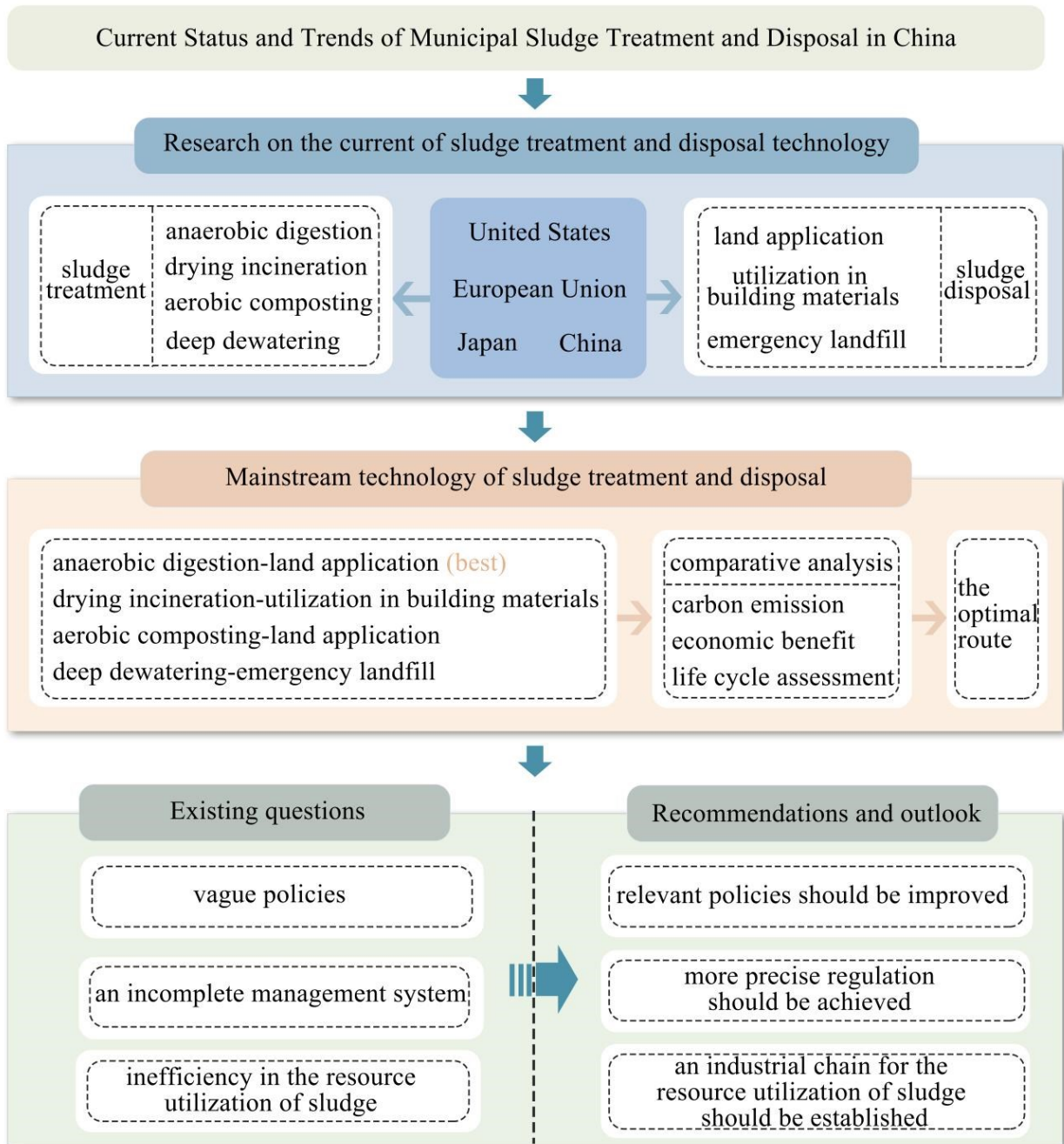
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8 **Graphical abstract**



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## 10 ABSTRACT

11 As urbanization accelerates, an increasingly significant problem of municipal sludge treatment and  
12 disposal has emerged in China. Sludge is both a waste and a resource, therefore, its reasonable and  
13 effective treatment and disposal are crucial for protecting the environment and promoting the  
14 recycling of resources. This article reviewed the current status and technologies of sludge treatment  
15 and disposal, proposed the mainstream routes for sludge treatment and disposal in China, including  
16 anaerobic digestion-land application, drying incineration-construction material utilization, aerobic  
17 composting-land application, deep dewatering-emergency landfill. According to the comparative  
18 analysis of carbon emission, economic benefit and life cycle assessment, anaerobic digestion-land  
19 application was considered the optimal technological route. It has extremely low carbon emissions (-  
20 44.43 kg·t<sup>-1</sup>) (calculated as CO<sub>2</sub>/dry sludge), low net costs (\$31.93/t) and significant environmental  
21 benefits, including SO<sub>2</sub> (-1.9×10<sup>5</sup> kg), electricity (-6.2×10<sup>8</sup> kWh) and fuel (-4.6×10<sup>7</sup> MJ). Based on  
22 this, it highlighted the issues in sludge treatment and disposal, such as vague policies, an incomplete  
23 management system, and inefficiency in the resource utilization of sludge. In response to these issues,  
24 suggestions were made to improve relevant policies, achieve more precise regulation, and establish  
25 an industrial chain for the resource utilization of sludge.

26 In summary, this article offered innovative insights into the optimal technological route for  
27 sludge treatment and disposal in China, while highlighting the practical engineering significance of  
28 addressing policy, management, and resource utilization challenges. Its recommendations have the  
29 potential to drive significant improvements in sludge management practices, contributing to a cleaner,  
30 greener, and more sustainable urban environment.

31 **Keywords:** Municipal Sludge, Treatment and Disposal Technology, Carbon Emission Reduction,  
32 Economic Benefit, Life Cycle Assessment

### 33 **Introduction**

34 Wastewater treatment generally consists of primary, secondary, and sometimes an advanced  
35 treatment process, with different biological, physical, and chemical technologies (Batt et al., 2007).  
36 At present, many sewage treatment processes are used in waste water treatment plants in China,  
37 including conventional activated sludge treatment, anaerobic-anoxic-oxic, anaerobic-oxic,  
38 sequencing batch reactor, oxidation ditch, etc (Jin et al., 2014). During these processes, different types  
39 of sludge are produced. Primary sludge (PS) is generated by the primary settling of municipal  
40 wastewater; secondary sludge, waste activated sludge (WAS) and excess sludge are extracted from  
41 aerobic tanks or secondary settlers or return sludge line; and mixed sludge is a combination of PS and  
42 WAS (Calabrò et al., 2024).

43 The management and treatment of these various types of sludge are crucial, with the expansion of  
44 urban areas and population growth, the production of municipal sludge in China has demonstrated an  
45 increasing trend year by year. According to statistics, China's annual sludge production reached 39.04  
46 million tons (80% water content) in 2019 (Zhou et al., 2022), and this number is expected to continue  
47 growing. The growth rate of sludge production will be even more significant especially in some large  
48 cities and industrially developed regions. Municipal sludge is a major by-product of the sewage  
49 treatment process, containing a large amount of harmful substances such as organic matter,  
50 pathogens, and heavy metals, as well as carbohydrates, proteins, fats, and nutrients such as nitrogen  
51 and phosphorus (Cheng et al., 2022). Sludge possesses the dual characteristics of being both pollutant  
52 and resource. If not treated and disposed in a reasonable and effective manner, it will cause serious  
53 resource waste and environmental pollution, thereby endangering human health (Dai et al., 2022). In  
54 order to mitigate the environmental pollution caused by municipal sludge and enhance the recovery

55 and utilization of resources within sludge, it is necessary to achieve the goals of reduction,  
56 stabilization, harmlessness, and resource utilization (Zhang et al., 2022).

57 In September 2022, the State Development and Reform Commission, the Ministry of Housing and  
58 Urban-Rural Development, and the Ministry of Ecology and Environment jointly issued the  
59 “Implementation Plan for Harmless Treatment and Resource Utilization of Sludge”, providing  
60 important directives on the application of sludge treatment and disposal technologies (Xue et al.,  
61 2023). The Chinese environmental protection departments have explicitly stipulated that complete  
62 harmless treatment and resource utilization of sludge must be achieved by 2035 (Zhou et al., 2022).  
63 However, there remain a series of problems. This article summarized the current status of sludge  
64 treatment and disposal both domestically and internationally, and proposed suggestions and outlooks  
65 based on the encountered problems.

## 66 **1. Current Status of Sludge Treatment and Disposal Technologies**

67 The technologies for the treatment and disposal of municipal sludge vary with different countries  
68 based on environmental policies, economic conditions, and resource feasibility.

69 In the European Union, anaerobic digestion and aerobic fermentation are the most commonly used  
70 technologies for sludge treatment (Kelessidis and Stasinakis 2012). Regarding the final disposal of  
71 sludge, the land application of sludge is the main choice for sludge disposal in EU-15 countries,  
72 accounting for 53%. Ireland and Lithuania commonly utilize land application for sludge treatment  
73 and disposal (Hudcová et al., 2019), whereas Netherlands and Germany tend to prefer incineration  
74 (Kacprzak et al., 2017). Landfilling remains the most common method for sludge disposal in the  
75 newly joined EU countries (Kelessidis and Stasinakis 2012). In the United States, the commonly used  
76 technologies for sludge treatment are anaerobic digestion and aerobic fermentation, and the main  
77 methods of sludge disposal are land application, landfilling, and incineration (Yakameran et al.,

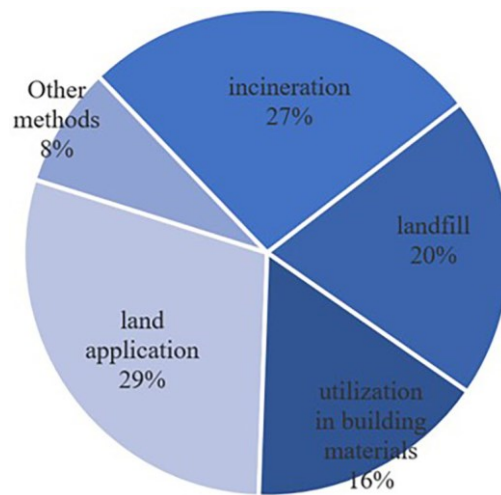
78 2021). According to the investigation of the U.S. Environmental Protection Agency (EPA), its sludge  
79 production in 2019 was about 4.75 million tons of dry sludge. Among them, around 2.44 million tons  
80 of dry sludge was used for land application, around 0.765 million tons of dry sludge was used for  
81 incineration, and around 1 million tons of dry sludge was landfilled and disposed of through other  
82 disposal methods (Qiu et al., 2023). In Japan, sludge treatment typically involves anaerobic digestion  
83 and incineration (Nakatsuka et al., 2020), and the main methods of sludge disposal are landfilling,  
84 and incineration (Lu et al., 2016).

85 China's sludge treatment technologies primarily include thickening, dewatering, anaerobic digestion,  
86 high-temperature aerobic fermentation, and thermal drying (Zhen et al., 2017). The principles and  
87 characteristics of these technologies were detailed in Table 1. At present, incineration and anaerobic  
88 digestion occupy dominant positions (Huang et al., 2023). The main methods of sludge disposal  
89 include land application (Liu et al., 2021), incineration (Fonts et al., 2012), sanitary landfill (Song  
90 and Lee 2010), and utilization in building materials, as illustrated in Figure 1 (Wei et al., 2020).

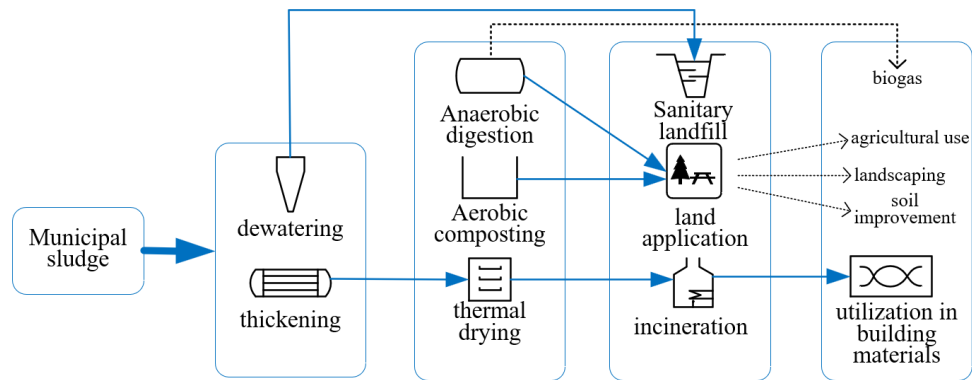
91 Drawing on foreign technologies, China has developed four mainstream sludge treatment and  
92 disposal technologies based on the characteristics of its sludge and regional differences, including  
93 anaerobic digestion-land application, drying incineration-utilization in building materials, aerobic  
94 composting-land application, and deep dewatering-emergency landfill, as shown in Figure 2.

**Table 1.** Sludge Treatment Technologies

<b>Sludge Treatment Methods</b>	<b>Principle</b>	<b>Advantages</b>	<b>Disadvantages</b>	<b>References</b>
Sludge Thickening Technology	The water content in the sludge is reduced by physical or chemical methods to increase the concentration of solid substances in the sludge.	A reduction in sludge volume leads to decreased transportation and processing costs.	The addition of chemical coagulant aids leads to increased treatment costs.	(Radetic 2024)
Sludge Dewatering Technology	This sludge treatment method removes water from liquid raw, thickened, or digested sludge, converting it into a semi-solid or solid clod.	Effective volume reduction, low energy consumption, minimal space requirement, and quick processing time.	The use of flocculants leads to higher treatment costs.	(Cao et al., 2021)
Sludge Anaerobic Digestion Technology	A biological treatment process that converts organic matter into biogas and stabilized residues through microbial metabolism under anaerobic conditions.	Odor and pathogen removal, sludge stabilization, and the production of methane as an energy source.	Inefficiency and low benefits, prolonged processing time, and demanding equipment requirements.	(Appels et al., 2021)
High-Temperature Aerobic Fermentation Technology	A process by which organic matter in sludge is decomposed and converted into humus-like substances through the metabolic activity of microorganisms under aerobic conditions.	Rapid processing speed, reduction in pathogens and parasites.	Increased energy consumption, and generation of odors.	(Mengqi et al., 2021)
Sludge Thermal Drying Technology	This technology further removes moisture from dewatered sludge to reduce its volume by transferring heat between the sludge and a thermal medium.	The volume and weight of sludge are significantly reduced, and the quality of the sludge is enhanced.	Elevated energy consumption, odor emissions, stringent equipment requirements, and increased operating costs.	(Li et al., 2012)



96  
97 **Figure 1.** Proportion of Sludge Utilization and Disposal Methods in China, 2019



98  
99 **Figure 2.** Mainstream Routes for Sludge Treatment and Disposal

100 *1.1 Anaerobic digestion-land application*

101 Anaerobic digestion-land application refers to the application of stabilized sludge from anaerobic  
 102 digestion to agricultural fields, gardens, green belts, and other lands, acting as a soil conditioner or  
 103 fertilizer. It is an effective method of sludge treatment and disposal (Yakameran et al., 2021). This  
 104 process not only stabilizes the biodegradable organic matter in the sludge, reducing the number of  
 105 pathogens and the volume of sludge (Dai et al., 2021), but also recovers organic matter and nutrients  
 106 from the sludge, aiding in soil structure improvement and plant growth promotion (Elmi and  
 107 AlOlayan 2020). Simultaneously, it features low energy consumption and negative carbon emissions  
 108 (Zhao et al., 2024), playing a crucial role in achieving environmental sustainability (Xu et al., 2021).



109 In China, combining anaerobic digestion with land application has become the preferred  
110 technological route for the treatment and disposal of sludge (Feng et al., 2015). According to Calabro's  
111 et al. (2024) statistical analysis, during anaerobic digestion, Eastern Asia, where the data were  
112 concentrated in China, Japan and South Korea, has the lowest mean bio-methane yield, with only  
113  $0.148 \text{ Nm}^3 \text{ kg}_{\text{VS}}^{-1}$ . Only the geographical area, the experiment date and the digested sludge type  
114 significantly influenced the bio-methane yield. It may be that the relatively low biodegradability of  
115 organic matter in municipal sludge in China leads to the lower efficiency of anaerobic digestion  
116 process and the reduction of methane production. Optimizing anaerobic digestion technology can  
117 effectively degrade organic matter and increase gas production. For example, research has  
118 demonstrated that the use of interspecies hydrogen transfer, hydrogen partial pressure, and microbial  
119 electrochemical systems can improve the overall efficiency of the anaerobic digestion process by  
120 enhancing the synthetic interactions among different microorganisms (Anukam et al., 2019). In  
121 addition, anaerobic digestion treatment can co-digest municipal sludge, food waste, and livestock and  
122 poultry manure, as well as other organic wastes, improving sludge treatment efficiency while  
123 increasing the production of biogas. Lan Mu et al. (2020) conducted a series of co-digestion anaerobic  
124 digestion experiments in a semi-continuous mode with different types of municipal sewage sludge,  
125 kitchen waste, and yard waste from different cities. As for co-anaerobic digestion of three feedstocks,  
126 high methane yields of  $314.9 \pm 17.1 \text{ mL/g VS}$  were achieved with a reliable stability. The results  
127 indicated that co-digestion anaerobic digestion not only improved methane yield, content, and  
128 production, but also promoted the sustainability of waste management and energy utilization.

129 After anaerobic digestion, the impact on methanogenic bacteria can significant, potentially lead to  
130 severe acidification of the system (Zeng et al., 2020). Although the organic content of the sludge  
131 decreases, this does not diminish its value for land application (Feng et al., 2015). However, the

132 potential environmental risks must be considered when using sludge for land application. Sludge  
133 contains harmful substances such as heavy metals and pathogens, which may pollute the soil and  
134 groundwater if not properly treated or used excessively. Therefore, a comprehensive assessment of  
135 the toxic and carcinogenic chemicals contained in sludge must be conducted before its land  
136 application (Yakameran et al., 2021). During land application, it is necessary to strictly control the  
137 amount and frequency of sludge application, continuously monitor its impact on the concentration of  
138 heavy metals in crops, and ensure its environmental and agricultural safety (Cocârță et al., 2019).

### 139 *1.2 Drying incineration-utilization in building materials*

140 Sludge drying and incineration technology involves two steps: drying and incineration. Firstly, the  
141 moisture in the sludge is evaporated through thermal drying, transforming wet sludge into dried  
142 sludge (Gao et al., 2023). This not only reduces the volume of the sludge but also prepares the dry  
143 material for the subsequent incineration process, thereby improving the efficiency of incineration  
144 (Xue et al., 2023). Then, the dried sludge undergoes high-temperature aerobic combustion (Dai 2020),  
145 which can completely decompose the organic matter in the sludge, eliminate pathogens and microbes,  
146 and stabilize heavy metals. Drying and incineration can reduce the volume of sludge by more than  
147 90% (Dai et al., 2021). The ash produced after sludge incineration can serve as raw materials or  
148 additives for construction materials (Ni et al., 2022), utilized in the production of lightweight  
149 aggregates, biochemical fiberboards, vitrified aggregates, sludge bricks, pipeline bedding materials,  
150 roadbed aggregates, etc (Zeng et al., 2020). This not only achieves resource utilization of sludge, but  
151 also reduces the dependence on natural resources, which has environmental and economic value  
152 (Ducoli et al., 2021).

153 In densely populated, economically developed cities with concentrated sludge production and scarce  
154 land resources, the drying and incineration technology route is often preferred (Duan et al., 2023).

155 Nena Duan et al. (2023) utilized Aspen Plus software to construct a process model of sludge drying  
156 and incineration and conducted an energy optimization configuration of thermal engineering design  
157 through multi-factor correlation analysis. This established a steady-state operation model of China's  
158 typical sludge drying and incineration process “conductive thermal drying-fluidized bed incineration-  
159 flue gas residual heat preheats air and supplements drying thermal energy” (Yang et al., 2021).  
160 Although the drying incineration-building material utilization route has demonstrated good  
161 environmental protection and resource utilization effects in sludge treatment and disposal, the drying  
162 and incineration processes may produce harmful substances, such as dioxins, posing threats to the  
163 environment and human health. Additionally, the incineration process is characterized by high energy  
164 consumption and requires significant energy input. Currently, sludge drying and incineration  
165 technology have been optimized and improved through deep integration across multiple fields. For  
166 example, Franco Falconi et al. (2020) utilized Linear Quadratic Regulator optimized waste-to-energy  
167 incineration technology, which addressed the shortcomings of traditional Single Input Single Output  
168 strategies. This approach reduced the emission of pollutants by controlling steam flow to manage  
169 energy production and ensuring complete combustion (Falconi et al., 2020). Simultaneously, a  
170 perfected intelligent sensing system has been implemented in current sludge drying and incineration  
171 projects, achieving automation and intelligence (Zhang 2023).

### 172 *1.3 Aerobic composting-land application*

173 Aerobic composting can be applied to various organic wastes, including sludge from sewage  
174 treatment plants and agricultural waste. In this process, sludge and organic matter are thoroughly  
175 mixed and composted under moist, ventilated, and high-temperature conditions, achieving harmless  
176 treatment and resource utilization of sludge (Nowak 2006). The construction and maintenance costs  
177 of the aerobic composting treatment and disposal process are relatively low. Additionally, the simple

178 process of operation and management and the high stability makes it suitable for land application  
179 (Dai 2020). However, this technological process has some disadvantages, such as slow process,  
180 occupying a large area, and having a threaten to the environment and human health. Therefore, Cheng  
181 Qingli et al. (2021) utilized enzymatic pretreatment combined with biological fortification to optimize  
182 the urban sludge aerobic composting technology. The mass fractions of soluble chemical oxygen  
183 demand, soluble protein, and polysaccharides in the sludge increased by 485.22%, 149.15% and  
184 108.76%, thereby improving the efficiency of sludge aerobic composting, reducing the start-up time  
185 of compost fermentation. Moreover, the addition of fortified microbial agents showed significant  
186 nitrogen preservation effects and reduced odor release, achieving rapid and efficient resource  
187 utilization of urban sludge.

188 Currently, the ecological risks associated with the land application of aerobic composting products of  
189 sludge are receiving increasing attention (Chang et al., 2019). Zheng et al. (2021) composted sludge  
190 contaminated with triclocarban (TCC) using wood chips and straw, respectively. The biodegradation  
191 of TCC is influenced by factors such as the type of bulking agents and the duration of composting.  
192 After land application, the soil concentrations of TCC were 2.30 ng g<sup>-1</sup> and 4.45 ng g<sup>-1</sup>, respectively.  
193 Following a risk assessment, the recommended the maximum application amounts for these two types  
194 of compost products are 35.0 t hm<sup>-1</sup> (for wood chip compost) and 18.0 t hm<sup>-1</sup> (for straw compost).

#### 195 *1.4 Deep dewatering-emergency landfill*

196 Through deep dewatering, the water content of sludge can be reduced to a considerably low level,  
197 thus reducing the space and cost for subsequent treatment and disposal (Cao et al., 2021). However,  
198 leachate is generated in the process, which may contain a large number of organic matter, heavy  
199 metals, nutrient salts and other pollutants. If not properly treated, the leachate will pollute the  
200 surrounding surface water and groundwater, causing serious damage to the ecological environment.

201 Emergency landfill refers to the temporary or long-term underground storage of treated sludge in  
202 specific landfill sites. In this process, a large number of greenhouse gases such as CH<sub>4</sub> and N<sub>2</sub>O are  
203 released in a disorderly manner, thus increasing carbon emissions. Deep dewatering-emergency  
204 landfill is a widely adopted sludge treatment and disposal technology in China. It is a low-cost method,  
205 but this technology causes serious secondary pollution, occupies land, wastes resources, and high  
206 carbon emissions. It is considered only as an emergency treatment method (Xu et al., 2021), serving  
207 as a transitional treatment and disposal approach (Dai et al., 2021).

208 Currently, more mature technologies for advanced sludge dewatering include physicochemical  
209 methods such as acid treatment, advanced oxidation technologies, and thermal treatment, as well as  
210 biodegradation methods, aimed at optimizing the dewatering performance and economic aspects of  
211 sludge (Dai 2020). For example, Xie et al. (2022) synthesized poly dimethyl diallyl ammonium  
212 chloride through radiation synthesis and combined it with polyaluminum chloride and calcium oxide  
213 as conditioning agents for advanced sludge dewatering, thereby optimizing the dewatering effect.

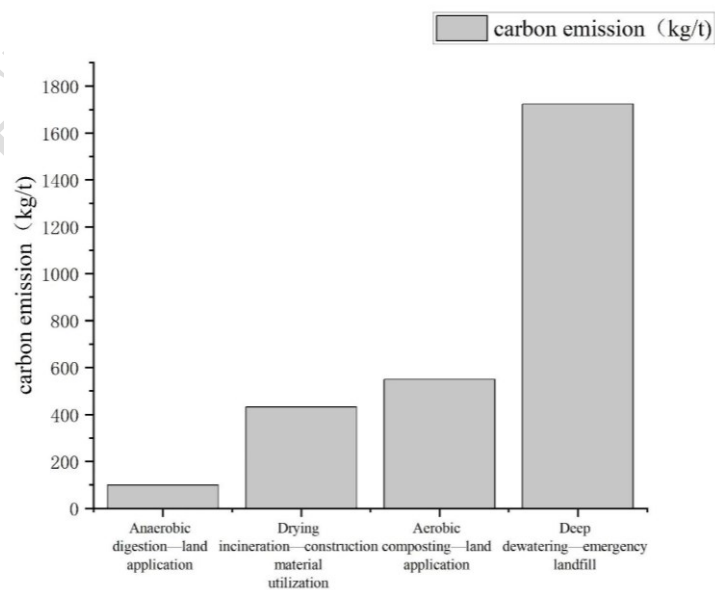
214 In practice, both the background and actual concentrations of toxic metals in the soil should be taken  
215 into consideration, when soils for the disposal of sewage sludge are selected.

### 216 *1.5 Comparative Analysis of Four Mainstream Technologies for Sludge Treatment and Disposal*

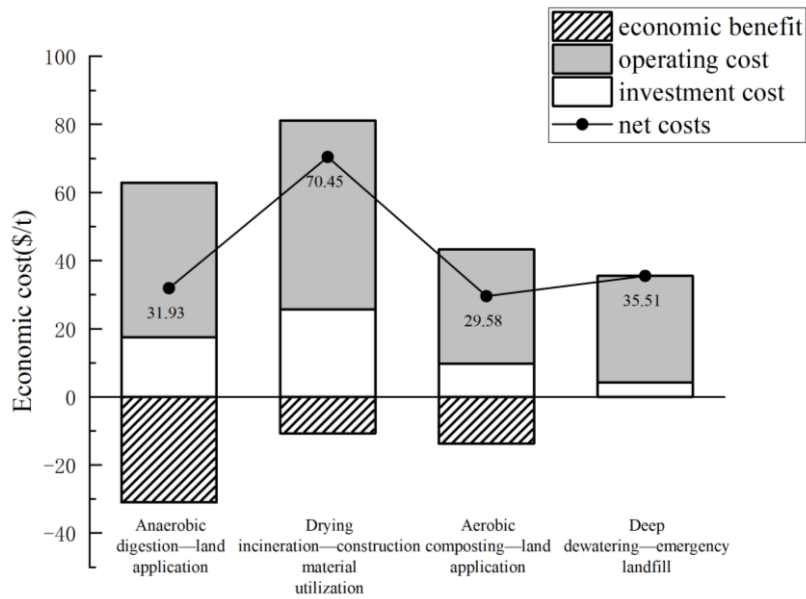
217 With the deep implementation of the dual carbon policy, carbon emissions have become an important  
218 indicator for selecting sludge treatment and disposal technologies. Research by Dai et al. (2021) had  
219 shown that the method with the highest carbon emissions is landfilling after deep dewatering,  
220 followed by sludge drying and incineration. Aerobic fermentation followed by land application has  
221 lower emissions, with the lowest carbon emissions being from land application after anaerobic  
222 digestion. On the other hand, Li et al. (2023) calculated the carbon emissions of each unit of sludge  
223 treatment and disposal, such as thermal drying (1049.24 kg·t<sup>-1</sup>), deep dewatering (960.99 kg·t<sup>-1</sup>),

224 sanitary landfill ( $786.24 \text{ kg}\cdot\text{t}^{-1}$ ), incineration ( $635.52 \text{ kg}\cdot\text{t}^{-1}$ ), anaerobic digestion ( $371.4 \text{ kg}\cdot\text{t}^{-1}$ ) and  
 225 aerobic composting ( $614.17 \text{ kg}\cdot\text{t}^{-1}$ ). The main carbon compensation methods include land application  
 226 ( $-415.83 \text{ kg}\cdot\text{t}^{-1}$ ) and building material utilization ( $-169.75 \text{ kg}\cdot\text{t}^{-1}$ ). Taking a comprehensive view, the  
 227 carbon emission of anaerobic digestion-land application is  $-44.43 \text{ kg}\cdot\text{t}^{-1}$ . Therefore, this technological  
 228 route has greater carbon offset potential and better environmental friendliness.

229 From an economic perspective, the land application route is cost-effective and offers high benefits,  
 230 as illustrated in Figure 4. Additionally, it features higher tolerance for errors, energy recovery  
 231 functions, and is environmentally friendly. In contrast, the use of building materials yields lower  
 232 benefits (Wang et al., 2023). Consideration cost, economic benefits, and the principles of green and  
 233 low carbon, for municipal sludge treatment and disposal, it is recommended to prioritize land use,  
 234 with the use in building materials as a secondary option. From the net costs perspective, aerobic  
 235 composting-land application ( $\$29.58/\text{t}$ ) is the most economical waste treatment route, followed by  
 236 anaerobic digestion-land application ( $\$31.93/\text{t}$ ). The cost difference between the two routes is  $\$2.35$   
 237 per ton of sludge treated, which is relatively small.



238  
 239 **Figure 3.** Carbon Emission Diagram of Mainstream Technologies for Sludge Treatment and  
 240 Disposal



**Figure 4.** Economic Cost Diagram of Sludge Treatment and Disposal Routes

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A comprehensive life cycle assessment identifies anaerobic digestion as the best sludge treatment technology (Xu et al., 2014). Regarding sludge disposal methods, using sludge as fertilizer for land application shows the best results. As indicated in Table 2, anaerobic digestion-land applications can reduce  $\text{SO}_2$  emissions by about  $1.9 \times 10^5$  kg, save about  $6.2 \times 10^8$  kWh of electricity consumption and about  $4.6 \times 10^7$  MJ of fuel consumption. This technical route provides the largest electricity offsets and the lowest fuel consumption (Murray et al., 2008). Considering carbon emissions, economic benefits and life cycle assessment, anaerobic digestion-land application becomes the preferred technology for sludge treatment and disposal due to its low carbon emissions, high economic benefits and significant environmental advantages.

257 **Table 2.** Environmental Assessment of Mainstream Technologies for Sludge Treatment and

258 Disposal

Mainstream Technologies of Sludge Treatment and Disposal	SO <sub>2</sub> (kg)	Electricity(kWh)	Fuel(MJ)
Anaerobic digestion-land application	$-1.9 \times 10^5$	$-6.2 \times 10^8$	$-4.6 \times 10^7$
Drying incineration-construction material utilization	$1.4 \times 10^6$	$-3.2 \times 10^7$	$7.1 \times 10^8$
Composting-land application	$1.2 \times 10^4$	$-5.9 \times 10^8$	$8.3 \times 10^7$
Deep dewatering-emergency landfill	$4.0 \times 10^3$	$5.7 \times 10^5$	$2.8 \times 10^6$

259 Note: positive values represent the emissions or consumption of the indicator, while negative values represent the  
 260 net savings or recovery of the indicator.

## 261 **2. Problems Faced by Sludge Treatment and Disposal**

### 262 *2.1 Policy provisions are not yet clear*

263 Currently, Chinese government departments are increasingly focusing on sludge treatment and  
 264 disposal, continuously clarifying the development direction of sludge treatment and disposal  
 265 technologies from the “13th Five-Year Plan” to the “Water Ten Articles” and then to the “14th Five-  
 266 Year Plan”. However, specific policies related to sludge treatment and disposal are scarce (Lu et al.,  
 267 2024), with a lack of mandatory provisions, standards, and engineering application technical  
 268 guidelines (Xue et al., 2023). For instance, the “Implementation Plan for Harmless Treatment and  
 269 Resource Utilization of Sludge” explicitly highlights the necessity of selecting reasonable and  
 270 diversified technology combinations based on local conditions but fails to provide specific application  
 271 methods (such as methods, seasons, frequency) for sludge treatment and disposal. Moreover, national  
 272 policies concentrate on controlling pollutant indicators before sludge treatment and disposal, but fail  
 273 to establish an environmental and ecological risk assessment system for the sludge treatment and  
 274 disposal process (Cheng et al., 2019).

275 Economically, China's annual investment in sludge treatment and disposal amounts to approximately



276 5.53 billion US dollars, in contrast to about 68.11 billion US dollars annually for wastewater treatment.  
277 However, in developed countries, the investment ratio between sludge treatment and wastewater  
278 treatment is approximately equal (Cocârță et al., 2019). This indicates insufficient investment in  
279 sludge treatment and disposal in China, resulting in the current situation where it remains at the  
280 harmless treatment stage, whereas developed countries have largely achieved the recycling and  
281 utilization of resources (Dai et al., 2022). Simultaneously, policies regarding economic incentives  
282 such as charges, taxes, and subsidies are not sufficiently specific and clear, and the costs of sludge  
283 treatment and disposal have not been effectively integrated into the standards, resulting in a lack of  
284 economic policies to foster industry development.

#### 285 *2.2 Management standard system is incomplete*

286 China's sewage plants, being of a large scale, produce concentrated amounts of sludge. However, due  
287 to an incomplete regulatory framework, there are gaps in supervision and instances of inadequate  
288 oversight. To reduce costs, some sewage treatment plants resort to unorganized emissions, private  
289 landfilling, or direct incineration for disposing of large quantities of sludge. These methods not only  
290 damage the soil and groundwater but also squander the potential for resource utilization of organic  
291 matter in the sludge (Dai et al., 2022).

292 Regarding top-level design, China lacks both a comprehensive management standard system and a  
293 multi-party coordinated management mechanism (Yang et al., 2015). During the management process  
294 of sludge treatment and disposal, sludge management institutions exhibit inconsistent standards (Lv  
295 et al., 2012), involving multiple departments and units, including environmental protection  
296 departments, municipal departments, and sewage treatment enterprises. However, the responsibilities  
297 and authorities of various management entities have not been clearly defined, potentially leading to  
298 management confusion and buck-passing.

## 299 *2.3 Sludge resource utilization is not smooth*

300 China's municipal sludge is characterized by its large volume, high moisture content, and high organic  
301 content, distinguishing it from sludge in other countries. This difference results in bottlenecks in  
302 adopting foreign technologies and equipment, as well as in implementing mature technology routes  
303 (Zhang et al., 2022). Additionally, each method has its downsides, which making the choice of  
304 technology route unclear. Meanwhile, the underdevelopment of sewage treatment systems and sludge  
305 treatment facilities hampers sludge resource utilization (Qu et al., 2019).

306 The current standards for sludge treatment and disposal are fragmented and disjointed. And these  
307 standards mandate that sewage treatment plants adopt a “one-size-fits-all” approach without  
308 considering the entire process of sludge treatment, transportation, and disposal. Moreover, there is an  
309 absence of the mindset that sludge disposal decisions should guide sludge treatment processes (Hu  
310 2019).

## 311 **3. Recommendations and Outlook for Sludge Treatment and Disposal**

### 312 *3.1 Policies related to sludge treatment and disposal technologies should be improved*

313 Technical guides and specialized technical specifications for urban sludge resource utilization should  
314 be developed in detail. First, these documents should elaborate on the main technological routes and  
315 methods currently employed for sludge resource utilization, as well as the corresponding directions  
316 for product development. Second, technical guides should provide comprehensive technical  
317 parameters, operating procedures, and safety guidelines for various technical paths. Finally, these  
318 policies should clarify the quality standards and environmental protection requirements of sludge  
319 treatment and disposal for resource utilization. This provides clear operational guidance and an  
320 evaluation basis for sludge harmless treatment and resource utilization.

321 In financial terms, the government should encourage enterprises and research institutions to engage

322 in technological innovation and product development by providing incentives such as financial  
323 subsidies, tax reductions, and green finance. Besides, a special fund should be established to  
324 specifically support scientific research, development, and demonstration projects for the resource  
325 utilization of sewage sludge, which provides solid financial support for the harmless treatment and  
326 resource utilization of sludge.

### 327 *3.2 Sludge treatment and disposal process to achieve refined regulation*

328 In the process of sludge treatment and disposal, it is necessary to establish a comprehensive regulatory  
329 and tracking system. In detail, this system should encompass every stage, from generation, collection,  
330 transportation, and processing, to the final utilization. This system ensures the transparency and  
331 traceability of information to facilitate the timely discovery and resolution of problems.  
332 Simultaneously, the key parameters of sludge treatment should be monitored in real time, such as  
333 temperature, pH value, and the content of harmful substances. This is crucial for maintaining the  
334 stability of the treatment process and ensuring the safety of the final product.

335 Moreover, departments should enhance coordination and cooperation by establishing a cross-  
336 departmental coordination mechanism composed of environmental protection, urban construction,  
337 agriculture, water affairs, and other departments. This mechanism should foster a synergistic  
338 development pattern for sludge treatment, disposal, and resource utilization. In addition, Wastewater  
339 treatment plants must strictly uphold the direct responsibilities of their governing departments and  
340 improve the management of the sludge treatment and disposal process. Furthermore, they should  
341 tighten the regulation of pollutant emissions in industrial processes and clearly delineate the pathways  
342 for harmless treatment. These pollutants from the sludge treatment process are ensured to be harmless  
343 before being discharged into the system.

344 3.3 Each link interlocks to create an industrial development chain for the resource utilization of  
345 sludge

346 Research institutions should conduct in-depth market demand analysis for products derived from  
347 sludge transformation to guide product development and marketing strategies. Then this will  
348 encourage more enterprises to adopt a product-oriented approach to sludge treatment and disposal,  
349 and to explore the potential to convert sludge into a variety of products, including bio-fertilizers, soil  
350 conditioners, building materials, and bioenergy. Thereby these research applications increasing the  
351 utilization pathways and enhancing the market value of sludge. Moreover, the state must establish a  
352 rigorous quality control system and participate in or promote the development of relevant product  
353 standards, to ensure the quality and safety of products transformed from sludge. In addition, the  
354 productization of sludge after treatment and disposal effectively introduces market capital and  
355 reduces the excessive reliance on government subsidies for sludge treatment and disposal. And then,  
356 the potential of sludge treatment and disposal has been activated to achieve sustainable development  
357 of sludge treatment and disposal in China.

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501 **Tables and Figures**

502 Table 1. Sludge Treatment Technologies

503 Table 2. Environmental Assessment of Mainstream Technologies for Sludge Treatment and Disposal

504 Figure 1. Proportion of Sludge Utilization and Disposal Methods in China, 2019

505 Figure 2. Mainstream Routes for Sludge Treatment and Disposal

506 Figure 3. Carbon Emission Diagram of Mainstream Technologies for Sludge Treatment and Disposal

507 Figure 4. Economic Cost Diagram of Sludge Treatment and Disposal Routes

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