

Physicochemical Variation of Water, Diversity, and Bacteria Community Structure in Post-coal Mining Ponds of Different Ages in Samarinda, East Kalimantan, Indonesia

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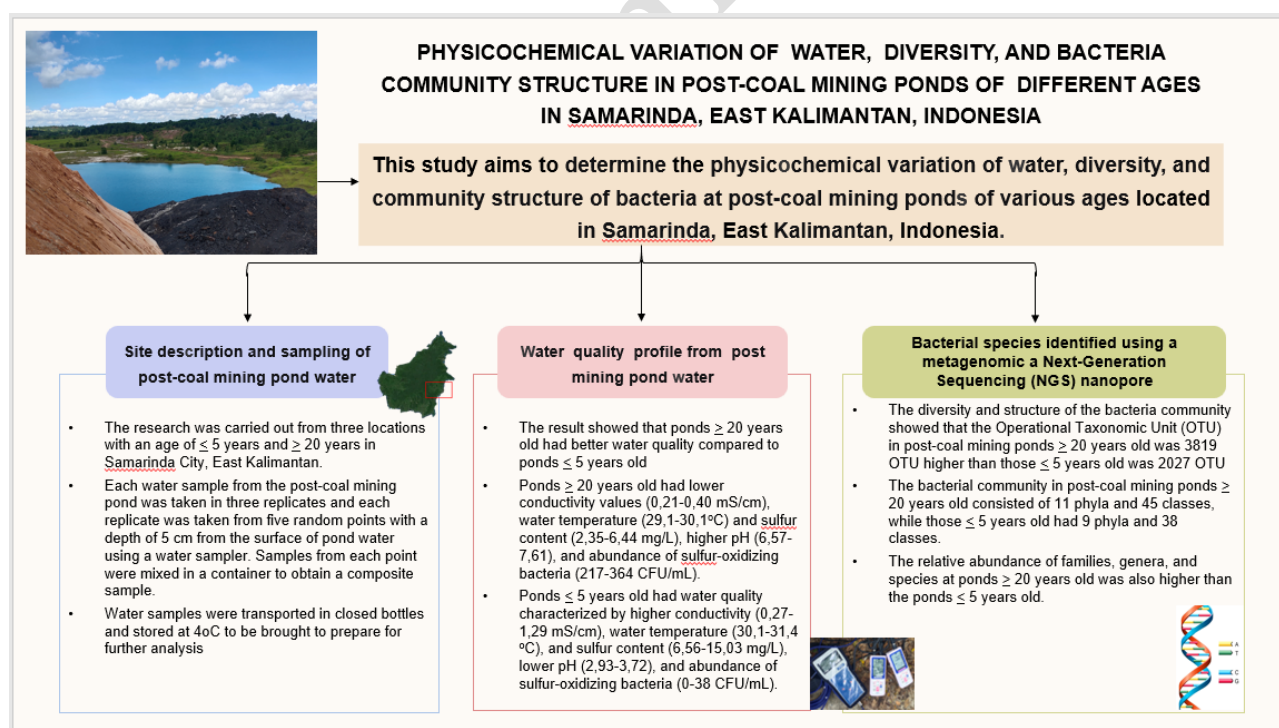
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GRAPHICAL ABSTRACT



ABSTRACT

This study aims to determine physicochemical variation of water, diversity, and community structure of bacteria at post-coal mining ponds of various ages located in Samarinda, East Kalimantan,

18 Indonesia. Bacteria species were identified using a metagenomic approach on the Next-Generation
19 Sequencing (NGS) nanopore platform and analyzed using the QIIME2 pipeline. The result showed
20 that ponds aged ≥ 20 years old had better water quality compared to others ≤ 5 years old. In addition,
21 ponds ≥ 20 years old had lower conductivity values (0,21-0,40 mS/cm), water temperature (29,1-
22 30,1°C), and sulfur content (2,35-6,44 mg/L), as well as higher pH (6,57-7,61) and abundance of
23 sulfur-oxidizing bacteria (SOB) (217-364 CFU/mL). Based on the results, ponds ≤ 5 years old had
24 water quality characterized by higher conductivity (0,27-1,29 mS/cm), water temperature (30,1-31,4
25 °C), and sulfur content (6,56-15,03 mg/L), lower pH (2,93-3,72), and abundance of SOB (0-38
26 CFU/mL). Diversity and structure of bacteria community showed that the Operational Taxonomic
27 Unit (OTU) in post-coal mining ponds ≥ 20 years old was 3819 compared to 2027 in those ≤ 5 years
28 old. Bacteria community in post-coal mining ponds ≥ 20 years old consisted of 11 phyla and 45
29 classes, while those ≤ 5 years old had 9 phyla and 38 classes. Moreover, the relative abundance of
30 families, genera, and species at ponds ≥ 20 years old was also higher. The high water quality,
31 diversity, and community structure of bacteria at ponds ≥ 20 years old showed the successful
32 remediation of ponds water by SOB.

33 **Keywords:** open-cut coal mining, coal mine, bacteria community, pond age, water quality

34 **1. Introduction**

35 East Kalimantan is one of the provinces in Indonesia with abundant coal mining potential due to the
36 widespread availability of resources. In 2019, the Regional Government of East Kalimantan Province
37 issued a total of 386 coal Mining Business Permits (IUP). These permits were distributed across
38 various regions, with 171 in Kutai Kartanegara Regency, 27 in Paser Regency, 23 in Berau Regency,
39 77 in West Kutai Regency, 33 in East Kutai Regency, 18 in North Penajam Paser Regency, and 37 in
40 Samarinda City (Dinas Pertambangan dan Energi Kalimantan Timur, 2021). In addition, Samarinda
41 City is widely known to be the second-largest coal-producing region in East Kalimantan with a flat
42 and hilly topography between 10-200 meters above sea level as well as a land area of 718 km². In the
43 2000s, the development of coal production in Samarinda City increased, leading to its consideration

44 as mining city because approximately 38,814 Ha (54%) of the total 71,823 Ha was used as coal
45 mining areas (Azwari & Rajab, 2021).

46 In line with previous studies, coal mining companies often implement the open-cut mining system
47 (Jiayin et al., 2020; Park et al., 2020). This system typically comprises various stages, namely
48 cleaning the surface of the soil, stripping the overburden, excavating the soil layer or coal seam, and
49 transporting coal to the stockyard for further use (Park et al., 2020). However, the open pit mining
50 system often causes the formation of large openings, which are filled with water containing sulfur
51 (Koščová et al., 2018) and become a habitat for bacteria. The sulfur contained in post-coal mining
52 pool water is obtained from coal containing 2-11% sulfur as well as coal-washing process. During
53 washing, the elements present in coal dissolve or are subjected to several chemical reactions and
54 physical reactions. This causes the washing products to dissolve into the washing solution (Dutta et
55 al., 2018). One of the groups of bacteria species present in post-coal mining ponds is sulfur-oxidizing
56 bacteria (SOB).

57 According to several reports, SOB play an essential role in the oxidation process of H_2S and other
58 reduced inorganic sulfur elements (elemental sulfur (S_0), pyrite (FeS_2), and sulfate (SO_2)). This is
59 primarily because these elements can function as an energy source and sulfur source for bacteria
60 metabolism (Hidayat et al., 2017). Various studies (Pourbabae et al., 2020 and Rana et al., 2020)
61 have shown that SOB is a group of bacteria with bioremediation abilities by carrying out sulfur
62 oxidation. Sulfur is an essential element for life, which is typically found in 2 amino acids and often
63 binds to hydrogen and oxygen ions in water. The element is often found in water in the form of sulfide
64 (S_2^-), hydrogen sulfide (H_2S), ferrous sulfide (FeS), sulfur dioxide (SO_2), sulfite (SO_3), and sulfate
65 (SO_4). The combination of sulfur and hydrogen has been shown to lead to the formation of sulfuric
66 acid. Another study showed that sulfur is often found in combination with alkali metals in lakes and
67 rivers. Despite the potential of SOB, their activity is influenced by environmental factors that drive
68 sulfur cycle (Méndez-García et al., 2015; Zhang et al., 2017) and age of post-coal mining ponds,
69 where the older pond, the higher diversity of life (Tala'ohu & Irawan, 2014). Post-mining ponds

70 generally consist of young and old variants aged < 5 years old and > 20 years old, respectively
71 (Prasetyono, 2015). Therefore, this study aims to 1) determine bacteria community structure isolated
72 from post-coal mining ponds water of various ages in Samarinda, and 2) assess the variations in water
73 quality profiles and the abundance of SOB in post-coal mining ponds aged ≤ 5 years and ≥ 20 years
74 old. The results are expected to provide new insights into indigenous bacteria species in post-coal
75 mining ponds water ecosystems in Samarinda City, East Kalimantan, Indonesia, the correlation model
76 among physicochemical parameters, and the density of SOB of ponds.

77 **2. Materials and methods**

78 *2.1. Site description and water sampling of post-coal mining ponds water*

79 This study was carried out from July-September 2022 in 3 locations of post-coal mining ponds with
80 age of ≤ 5 years and 3 locations of post-coal mining ponds with age of ≥ 20 years in Samarinda City,
81 East Kalimantan. The procedures were also conducted at Water Quality Laboratory of the Faculty of
82 Fisheries and Marine Sciences, Mulawarman University, Samarinda, Indonesia, and the Animal
83 Ecology and Diversity Laboratory, Faculty of Mathematics and Natural Sciences, Universitas
84 Brawijaya, Malang, Indonesia. The details of the sampling locations are presented in Figure 1. Each
85 water sample from post-coal mining ponds was taken in 3 replicates and each replicate was obtained
86 from 5 random points with a depth of 5 cm from the surface of ponds water using water sampler.
87 Samples from each point were mixed in a container to obtain a composite sample. This procedure
88 was repeated with samples from every 3 replicates and 5 different points. For further analysis, water
89 samples were transported in sealed bottles, stored at 4 °C, and transported to the Microbiology
90 Laboratory, Faculty of Mathematics and Natural Sciences, Universitas Brawijaya, Malang, Indonesia.

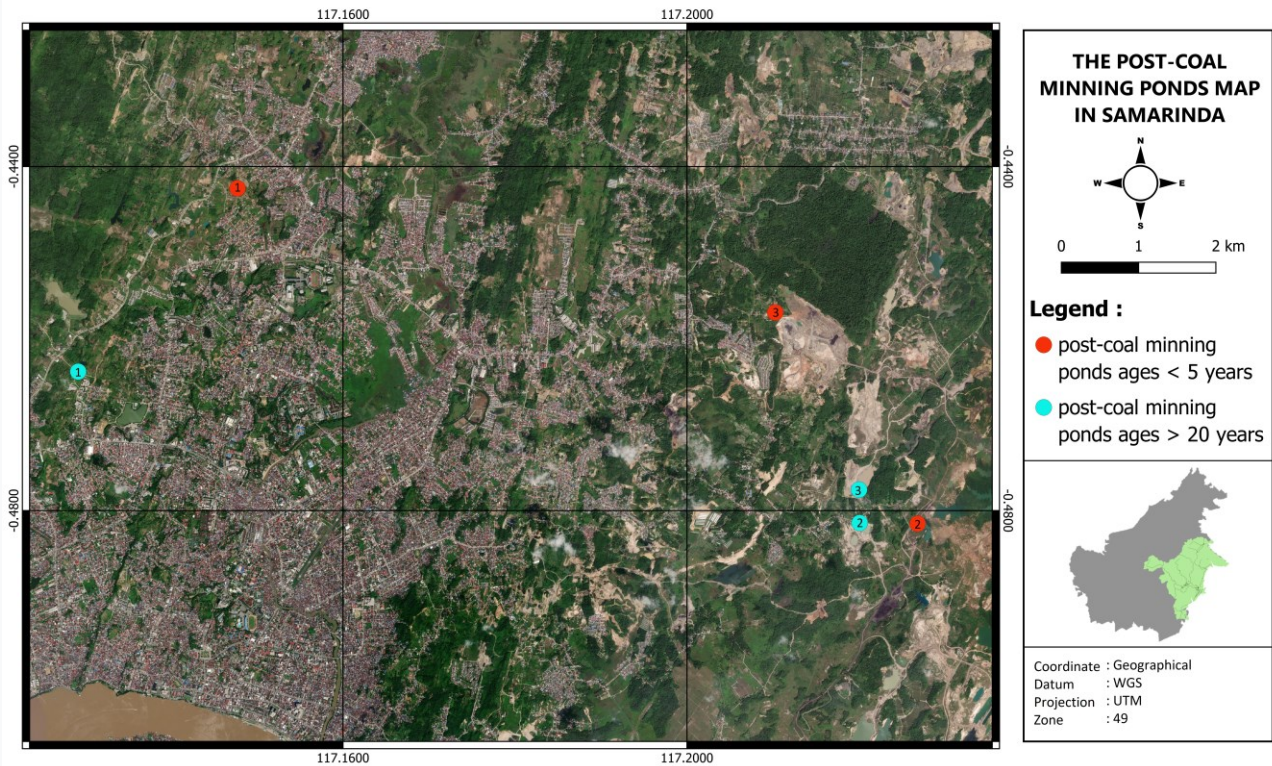


Figure 1. Details of the sampling locations

2.2. Analysis of physicochemical and SOB communities structure of water ponds

Physicochemical parameters of water ponds were analyzed in-situ (direct) and ex-situ (in the laboratory) to determine water quality of age of both post-coal mining. A total of 1 liter of water sample was taken from each pond (Yousef et al., 2019; Ma et al., 2020). Water parameters analyzed in situ were temperature, pH, and dissolved oxygen, while those analyzed in the laboratory were COD (SNI 6989.2:2019), BOD (APHA,5210-B,22ND th.2012), and SO₄ (SNI 6989.20-2009).

In addition, all ponds water samples of the same age were mixed into 1 sample. The mixed ponds water was filtered using a filter membrane (Merck) with a series of pore diameters of 11.0 μm, 0.45 μm, and 0.20 μm. The 0.20 μm filter membrane containing bacteria cells was cut into small pieces and placed into a Lysing Matrix E tube. Bacteria DNA was extracted according to the FastDNA Spin Kit (MPBIO) Germany protocol with modifications, namely 1) increasing the homogenization time of the sodium phosphate buffer sample from 5 to 10 seconds using a vortex, 2) the suspension was incubated for 10 minutes at room temperature (18 ± 1 °C) after which the supernatant protein precipitate solution was inverted. The increase in homogenization time was intended to ensure that

107 the sample and sodium phosphate buffer were well mixed, while incubation of the supernatant-
108 precipitate solution was performed for complete precipitation of protein-supernatant.

109 *2.3. Amplification of bacteria chromosomal 16S rDNA*

110 Extracted DNA was analyzed qualitatively by 1% agarose gel electrophoresis. The DNA
111 concentration and purity were measured using a NanoDrop Spectrophotometer. In addition,
112 amplification of bacteria chromosomal 16S rDNA was carried out through Oxford Nanopore
113 Technology (ONT) 16S ribosomal RNA (16S rRNA) gene sequencing, which provided cover for the
114 full 16S rRNA gene sequence (region V1–V9). Nanopore sequencing was operated by MinKNOW
115 software version 22.05.7. Base-calling was carried out using Guppy version 6.1.5, FASTQ file quality
116 was visualized using NanoPlot, and quality filtering was performed using NanoFilt.

117 *2.4. Taxa richness and diversity analysis*

118 The alpha diversity and beta diversity were subsequently performed using the normalized data. Alpha
119 diversity was applied to analyse species diversity in a sample using observed-species, Chao1,
120 Shannon, Simpson, and ACE expansion with QIIME (Version 1.7.0). In addition, community
121 richness was identified using Chao1, ACE, Shannon, and Simpson indexes. Community diversity was
122 identified using the Shannon and Simpson indexes to assess species richness and diversity for each
123 water sample from different locations (Wang et al., 2018).

124 *2.5. Physicochemical data of water analysis*

125 Physicochemical data of water at post-coal mining ponds was analyzed according to analysis of
126 variance and biplot analysis using Principal Component Analysis (PCA) (Souza et al., 2021; Wang
127 et al., 2018) with the PAST program. Differences in water quality between post-mining ponds were
128 identified by ANOVA followed by the Tukey HSD test (BOD and SOB) (Sheng et al., 2023). The
129 analysis was continued with Brown Forsythe, Games-Howell test (pH, DO, conductivity,
130 temperature, TSS, and Sulfate), Kruskal Wallis, and Mann Whitney test (COD) (Reddy et al., 2020)
131 using SPSS v.22. The correlation among physicochemical parameters and abundance of SOB of post-

132 coal mining ponds water were determined by Pearson correlation analysis (Wang et al., 2018) using
 133 SPSS v.22.

134 3. Results and Discussion

135 3.1. Water quality profile from post-mining ponds water in Samarinda

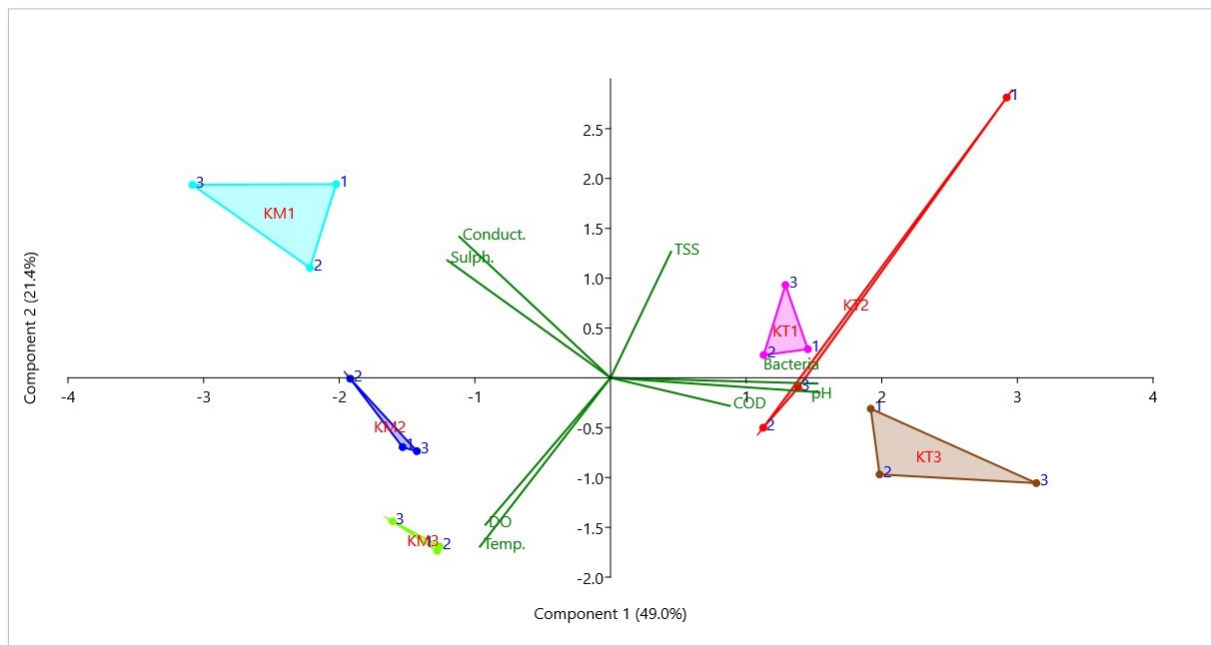
136 Physicochemical and the abundance of SOB in water at post-coal mining ponds ≤ 5 years and ≥ 20
 137 years old are presented in Table 1.

138 Table 1. Physicochemical and SOB parameters of post-coal mining ponds water in Samarinda

No	Parameter	Post-coal Mining Ponds					
		≤ 5 years old			≥ 20 years old		
		1	2	3	1	2	3
1	pH	2,93 ^a	3,23 ^b	3,72 ^c	6,57 ^d	6,84 ^{de}	7,61 ^e
2	DO (ppm)	3,56 ^a	3,71 ^a	3,72 ^a	3,12 ^a	3,27 ^a	3,56 ^a
3	Conductivity (S/m)	1,29 ^e	0,56 ^d	0,40 ^c	0,27 ^b	0,23 ^{ab}	0,21 ^a
4	Temperature (°C)	30,10 ^{ab}	30,43 ^b	31,40 ^c	29,60 ^a	29,13 ^{abc}	30,10 ^{ab}
5	TSS (mg/L)	0,81 ^a	0,05 ^a	0,24 ^a	0,12 ^a	0,98 ^a	0,86 ^a
6	COD (mg/L)	16,20 ^a	15,15 ^a	13,12 ^a	18,14 ^{ab}	23,98 ^b	24,33 ^b
7	BOD (mg/L)	6,40 ^a	6,07 ^a	4,55 ^a	5,54 ^a	5,82 ^a	7,67 ^a
8	Sulfate (mg/L)	15,03 ^{ab}	9,50 ^{ab}	6,56 ^{ab}	6,44 ^{ab}	4,74 ^b	2,35 ^a
9	Abundance of SOB (CFU/mL)	0 ^a	30 ^a	38 ^a	259 ^b	217 ^b	364 ^c

139 Note: numbers followed by the same letter in the same parameter are not significantly different based
 140 on the 5% ANOVA test.

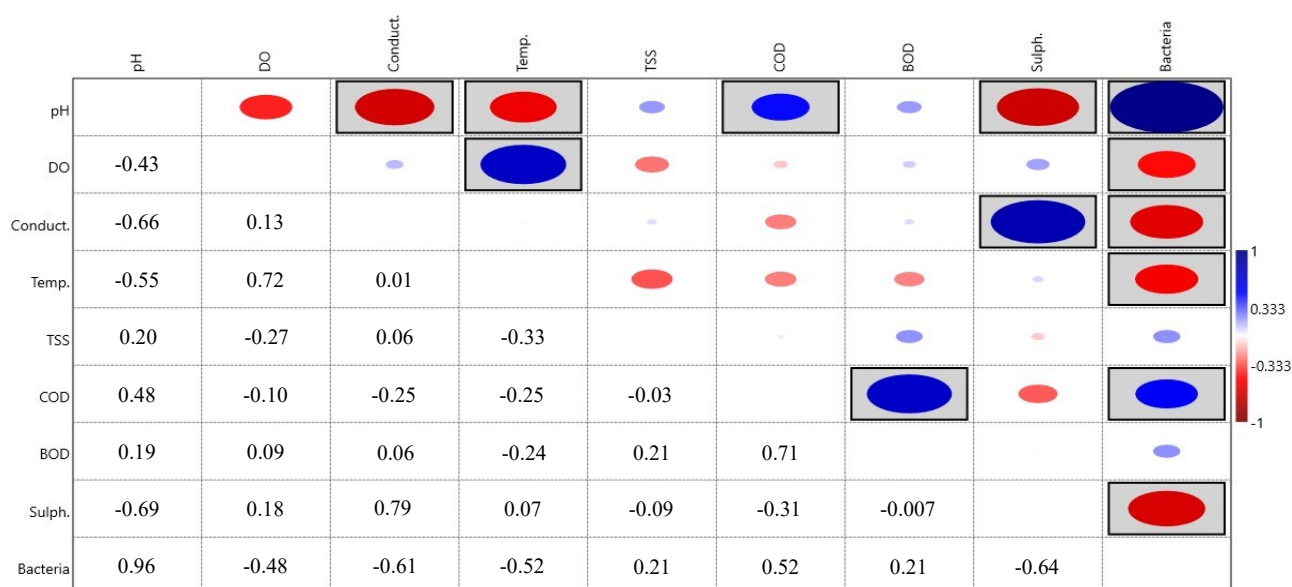
141
 142 Table 1 showed that ponds ≥ 20 years old were characterized by lower conductivity (0,21-0,40
 143 mS/cm), water temperature (29,1-30,1 °C), sulfur content (2,35-6,44 mg/L), higher pH (6,57-7,61),
 144 and abundance of SOB (217-364 CFU/mL). Meanwhile, ponds ≤ 5 years old had lower water quality,
 145 higher conductivity (0,27-1,29 mS/cm), water temperature (30,1-31,4 °C), and sulfur content (6,56-
 146 15,03 mg/L), lower pH (2,93-3,72), and an abundance of SOB (0-38 CFU/mL). Tomal (2020) stated
 147 that the growth and development of microorganisms were stimulated or inhibited by environmental
 148 factors. The main factors affecting bacteria growth included nutrients, pH, water, oxygen
 149 (Chrismanuel et al., 2012), temperature, and humidity in the environment (Tomal, 2020).



150

151 **Figure 2.** Loading plot post-coal mining ponds water in Samarinda based on biplot analysis using
 152 PCA; KM1, KM2, KM3 = post-coal mining ponds ≤ 5 years old; KT1, KT2, KT3 = post-coal
 153 mining ponds ≥ 20 years old
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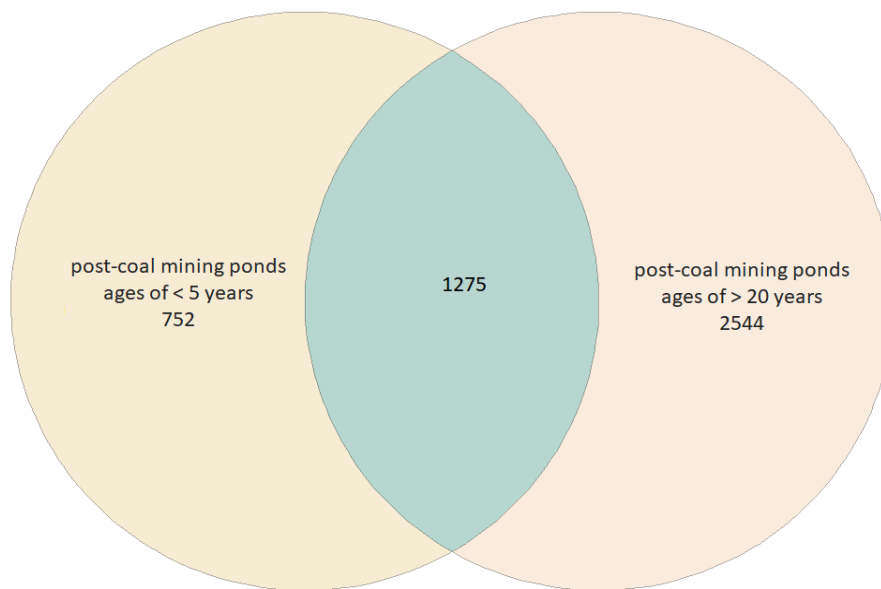
155 Figure 2 showed that the parameters of ponds ≤ 5 and ≥ 20 years old were in one group. The results
 156 showed that there was a significant difference in water quality between the 2 types of ponds. Water
 157 of ponds ≤ 5 years old was characterized by high conductivity, sulfur content, temperature, and low
 158 of density of SOB, COD, and pH. Meanwhile, water of ponds ≥ 20 years old was characterized by
 159 high TSS, density of SOB, pH, and COD, as well as low conductivity, sulfur content, DO, and
 160 temperature. The higher SOB density in ponds ≥ 20 years old led to lower sulfur concentration
 161 compared to those aged ≤ 5 years old. This was because ponds ≥ 20 years old had a more complex
 162 and diverse microorganism community that developed and formed a stable ecosystem. The thriving
 163 microorganism community could have SOB species that had adapted and developed well under ponds
 164 conditions. Stable environmental conditions could support the growth and activity of SOB more
 165 effectively, leading to higher population densities and the accumulation of sulfur and other
 166 compounds necessary for the growth of SOB (Rawlings, 2005).



167 **Figure 3.** Correlation Pearson between physicochemical parameters and SOB density of water
 168 Correlation Pearson analysis was a method for analyzing the relationship between variables of water
 169 ponds. When there was a relationship, the changes that occurred in one of the variables led to changes
 170 in the other variables. The term was said to be a causal term, and it was the hallmark of correlation
 171 analysis. The results of the study in Figure 3 showed there was a positive correlation/relationship with
 172 SOB. This showed that an increase in the density of SOB caused an increment in pH and COD values,
 173 with pH value affecting the microbial metabolism. The gene expression in *Sulfobacillus*
 174 *thermosulfidooxidans* and *Ferroplasma thermophilum* was primarily affected by pH values (Peng et
 175 al., 2019). Maintenance of pH was important due to its significant effect on bacteria growth. In
 176 addition, bacteria growth was determined by optimizing pH and incubation time. The optimization of
 177 environmental conditions was very important for the enhancement of bacteria growth and for
 178 designing an effective biodegradation strategy. The level of the relationship was significant because
 179 it was more than 0.5. Figure 3 also showed the negative correlation between sulfur concentration,
 180 temperature, conductivity, and DO to the density of SOB. This showed that a high density of bacteria
 181 caused a lower concentration of sulfur, temperature, conductivity, and DO. The results showed that
 182 environmental factors affected the activity of enzymes in bacteria, thereby influencing the efficiency
 183 of SOB in treating sulfur (Dong et al., 2023).

184 **3.2. Bacteria community structure based on Oxford Nanopore Technology analysis**

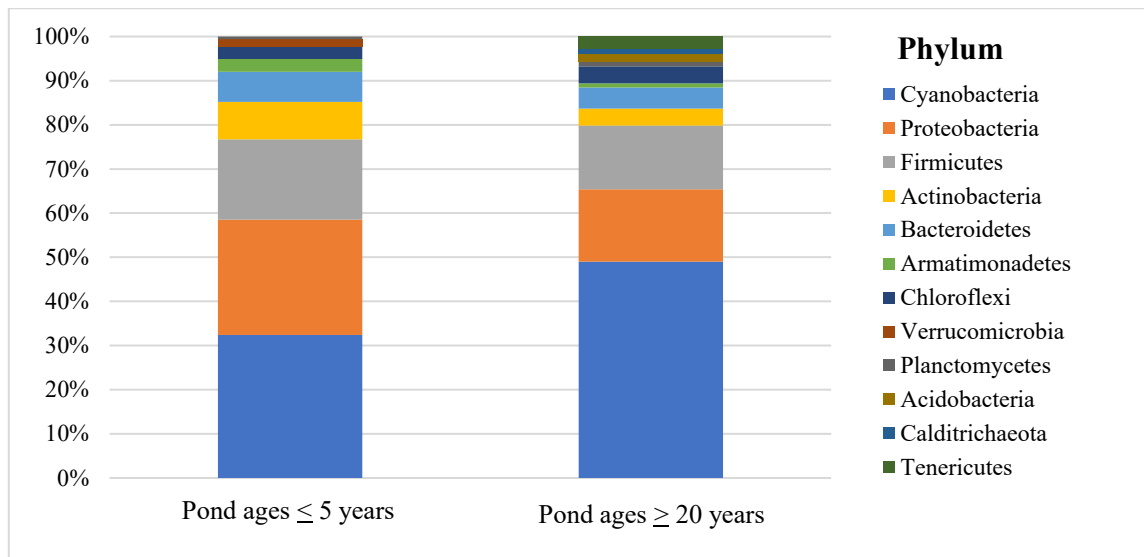
185 Based on the NGS analysis, post-coal mining ponds ≥ 20 years old had 3819 OTUs of bacteria, which
186 was higher than 2027 OTUs of ponds ≤ 5 years old (Figure 4). On the Venn diagram (Figure 4), each
187 circle represented a sample or group of post-coal mining ponds. This showed that age of ponds and
188 physicochemical environmental factors affected the composition and diversity of bacteria. Based on
189 the Venn diagram, there were 1275 OTU in both ages of ponds (≥ 20 years and ≤ 5 years old).



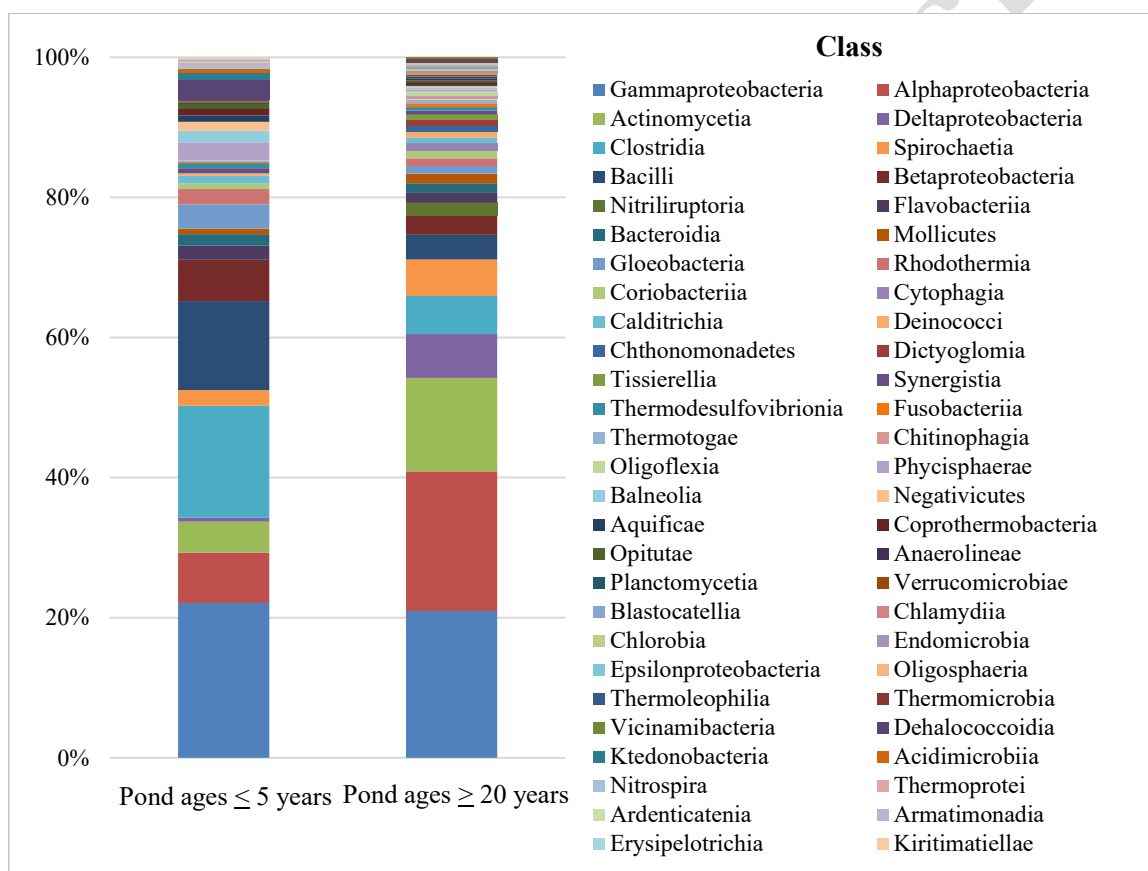
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191 **Figure 4.** Venn diagram depicting the relationship between sample

192 The results of the Venn analysis obtained in Figure 4 showed that coal post-mining ponds samples
193 with age of ≥ 20 years had more OTUs when compared to those aged ≤ 5 years. This showed that
194 ponds age and physicochemical environmental factors influenced the composition and diversity of
195 bacteria. In this study, the dominant bacteria phyla in both post-coal mine ponds were Cyanobacteria,
196 Proteobacteria, and Firmicutes (Figure 5). The results were consistent with Wangka et al., (2020),
197 where Proteobacteria, Firmicutes, and Cyanobacteria were the phylum communities found in Bangka
198 Islands sediments based on the results of NGS analysis.



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Figure 5. Relative abundant of bacteria phylum and class from post-coal mining ponds ≤ 5 years and ≥ 20 years old

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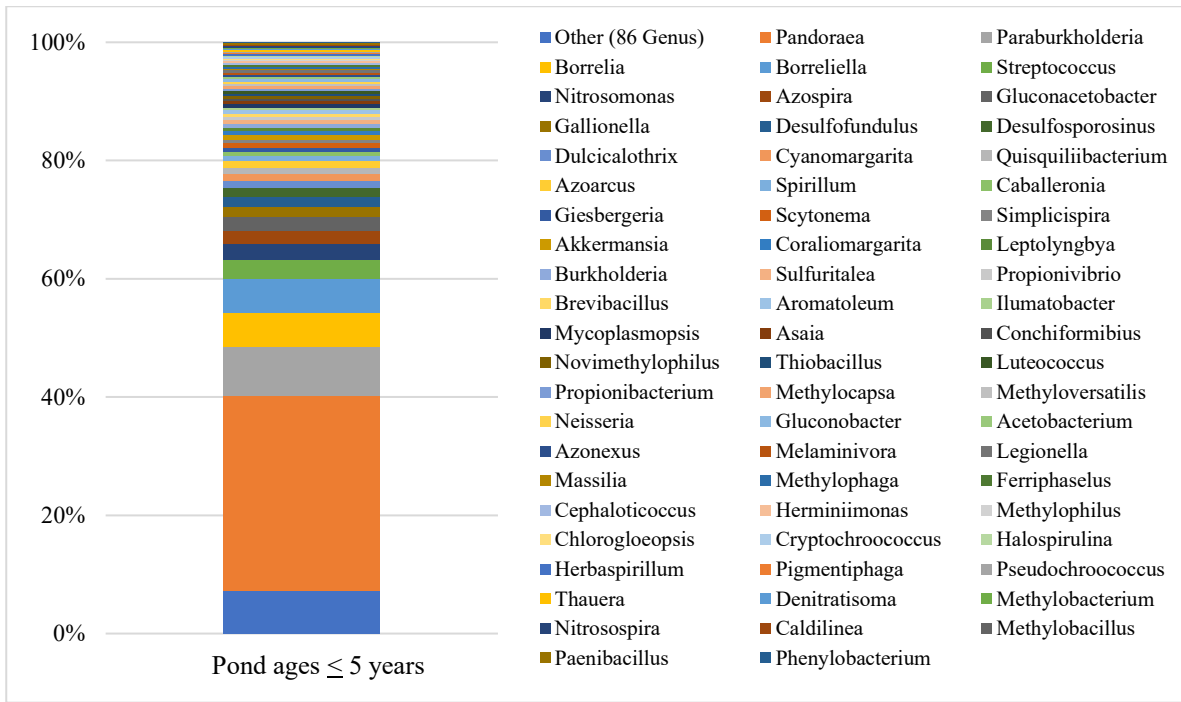
Figure 5 showed that diversity of phyla at ponds ≤ 5 years old consisted of 9 phyla (Cyanobacteria 32%, Proteobacteria 26%, Firmicutes 18%, Actinobacteria 9%, Bacteroidetes 7%, Armatimonadetes 3%, Chloroflexi 3%, Verrucomicrobia 2% and Planctomycetes 1%), while ponds ≥ 20 years consisted of 11 phyla (Cyanobacteria 49%, Proteobacteria 16%, Firmicutes 14%, Actinobacteria 4%,

208 Bacteroidetes 5%, Armatimonadetes 1%, Chloroflexi 4%, Planctomycetes 1%, Acidobacteria 2%,
209 Calditrichaeota 1%, and Tenericutes 3%). The phyla Acidobacteria, Calditrichaeota, and Tenericutes
210 were not found in ponds aged ≤ 5 years, while the phylum Verrucomicrobia was not found in those
211 aged ≥ 20 years. Based on a study conducted by Lee et al. (2009), Verrucomicrobia was a phylum
212 included in the soil and water bacteria community and was typically found in hot springs. Figure 5
213 also showed diversity of classes in ponds aged ≤ 5 years, where there were 27 bacteria classes, namely
214 Gammaproteobacteria (22%), Clostridia (16%), Bacilli (13%), Alphaproteobacteria (7%),
215 Betaproteobacteria 6%, and others (36%). Class diversity in ponds aged ≥ 20 years consisted of 27
216 different bacteria classes, comprising Gammaproteobacteria (21%), Alphaproteobacteria (21%),
217 Actinomycetia (13%), Deltaproteobacteria (6%), Clostridia (5%), Spirocahetia (5%), and others
218 (29%).

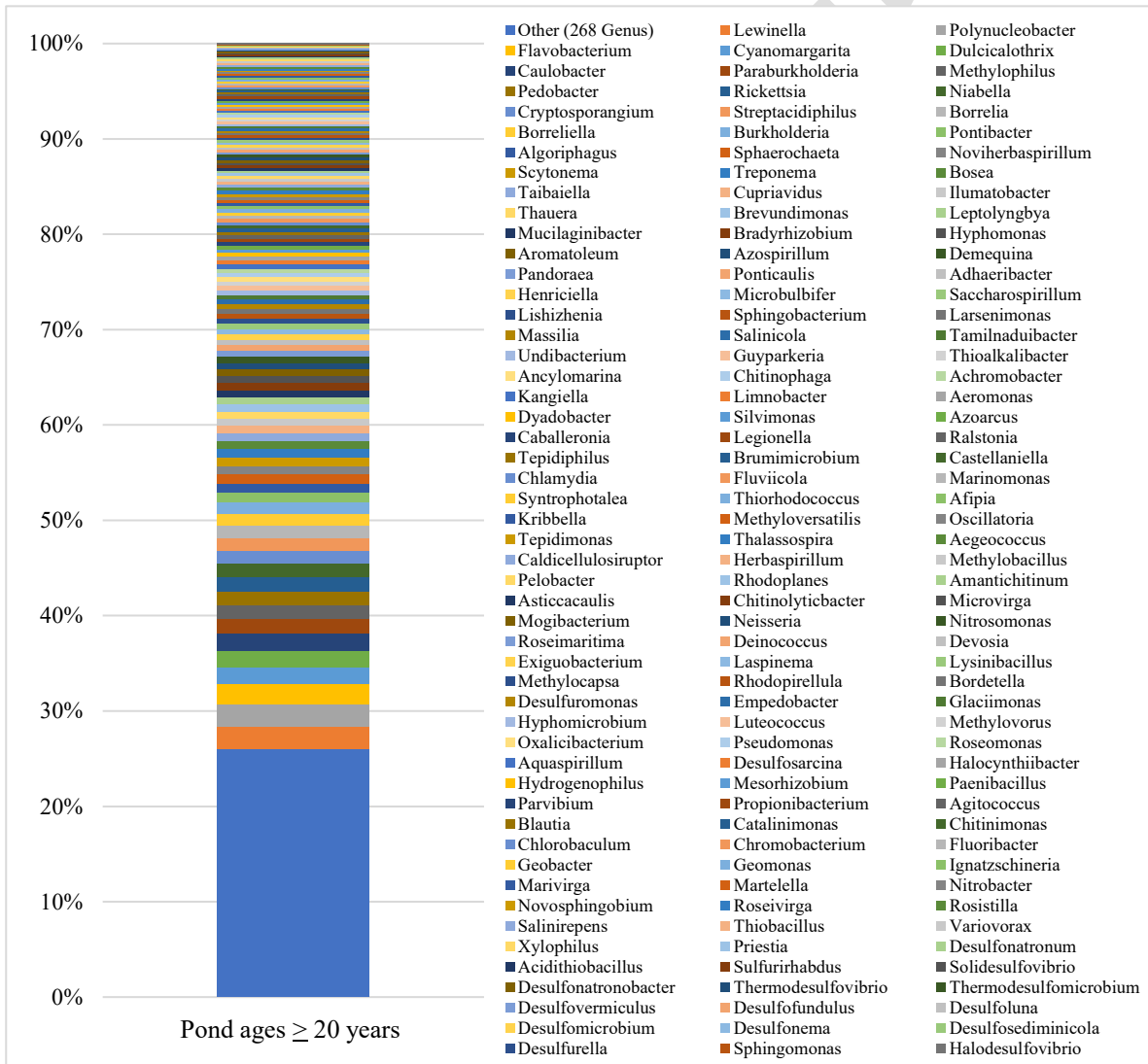
219 The phylum Cyanobacteria was found to be dominant in both age groups of coal post-mining ponds.
220 The main role played by Cyanobacteria in ponds ecosystem was that it performed photosynthesis and
221 was the main producer of oxygen. In this process, Cyanobacteria used sulfur and converted it into
222 compounds that could be used by other organisms. This made an important contribution to the sulfur
223 cycle (Kushkevych et al., 2021). Cyanobacteria could also use various nutrients, such as nitrogen,
224 phosphorus, and carbon dioxide in water. The ability of the microbe to use these nutrients helped
225 control nutrient concentrations in ponds, which in turn affected the growth of other organisms in the
226 food chain (Kamennaya et al., 2012).

227 To determine the relative abundance of genus metagenomically from the 2 ponds ages used as
228 samples in this study, the analysis was carried out and presented in Figure 6.

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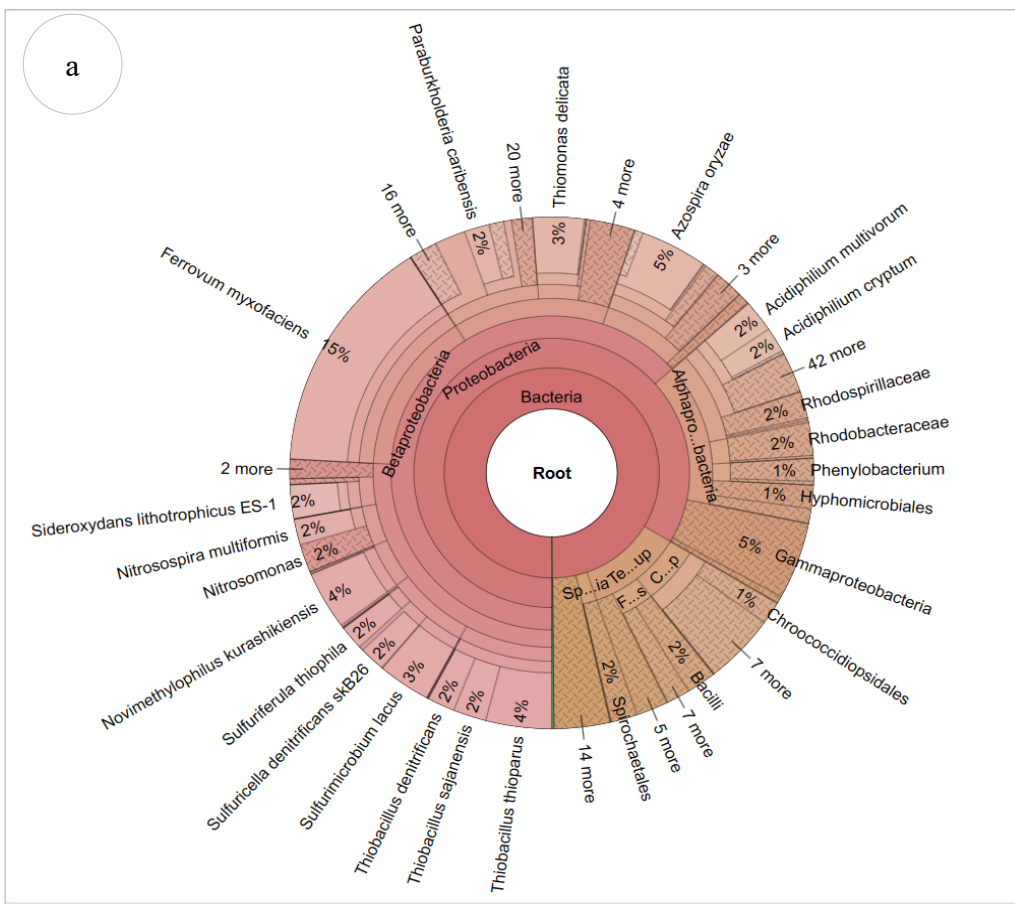
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Figure 6. Relative abundant of bacteria genera of two different ages of post-coal mining ponds

232 Figure 6 showed that bacteria genus richness in ponds ≥ 20 years old was higher compared to ponds
233 ≤ 5 years old. These results showed that ponds age affected bacteria community structure in ponds
234 water. Cole et al., (2019) also showed that bacteria species richness in ponds ≥ 20 years old was
235 significantly higher compared to those aged ≤ 5 years old. This showed that in their activities, SOB
236 were influenced by environmental factors driving the sulfur cycle (Méndez-García et al., 2015; Zhang
237 et al., 2017) and age of post-coal mining ponds because aging caused higher biodiversity (plankton,
238 fish, and other aquatic biota) (Tala'ohu & Irawan, 2014). Diversity and abundance of bacteria in coal-
239 affected environments could be useful bioindicators for post-coal mining ponds restoration. For
240 example, bacteria belonging to the Gammaproteobacteria group could be accurate bioindicators of
241 the biodegradation potential of coal (Akimbekov et al., 2022). Members of the Genus
242 *Acidithiobacillus* were dominantly able to oxidize sulfur in various sulfur-rich environments around
243 the world, and members of this genus were believed to play a key role in the biogeochemical cycles
244 of sulfur. According to the List of Prokaryotic Status in Nomenclature (LPSN) database available at
245 <http://www.bacterio.net>, several species of the Genus *Acidithiobacillus* had been validated, including
246 *Acidithiobacillus thiooxidans* (*Thiobacillus thiooxidans*), *A. albertensis* (*T. albertis*), *A. caldus* (*T.*
247 *caldus*), *A. ferrooxidans* (*T. ferrooxidans*), *A. ferrivorans*, *A. ferridurans*, and *A. Ferriphilus* (Wang
248 et al., 2019). *A. thiooxidans* (*T. thiooxidans*) and *A. ferrooxidans* (*T. ferrooxidans*), were generally
249 considered to be the most important bacteria for sulfur oxidation in soil. In this process, sulfur was
250 oxidized and then entered into the biosynthesis of cysteine and methionine (Kushkevych et al., 2020;
251 Melo et al., 2019).

252 Based on Figure 7, Metagenomic analysis was used to identify the microbial community at post-coal
253 mining ponds ≤ 5 and ≥ 20 years old. The results were helpful in understanding the interaction
254 between microbial communities and the environment.

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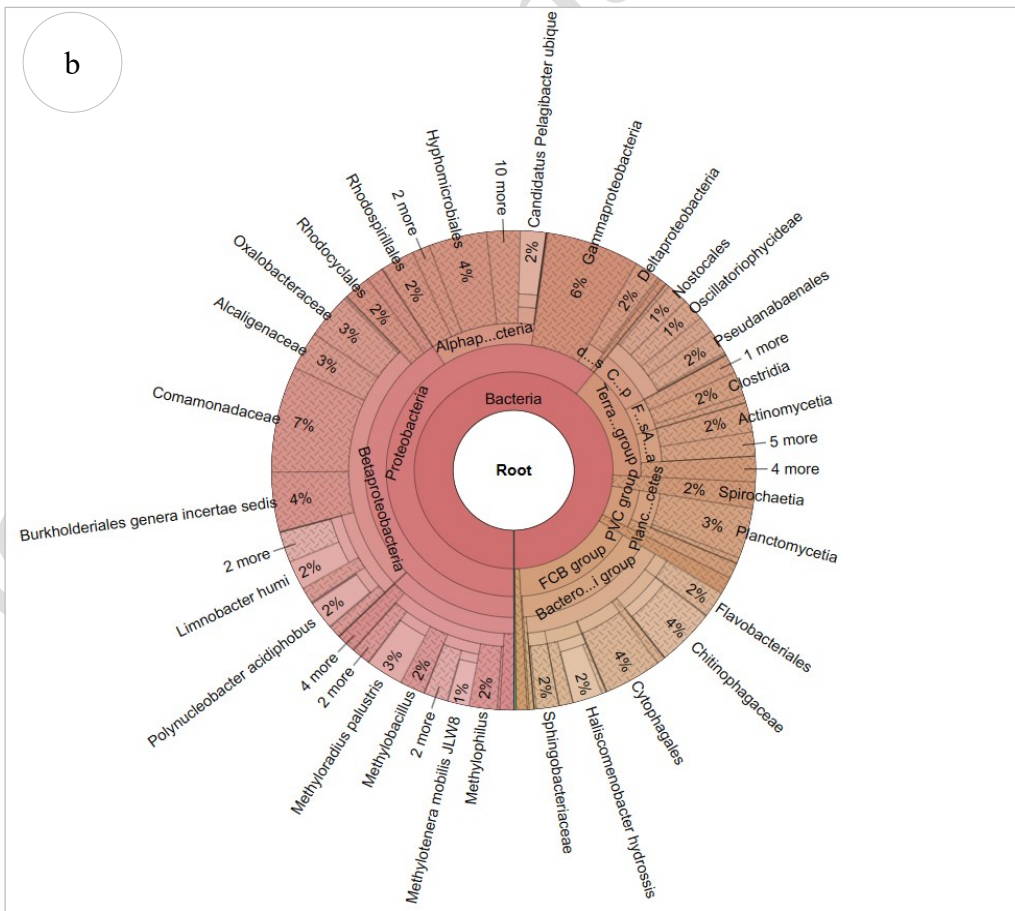
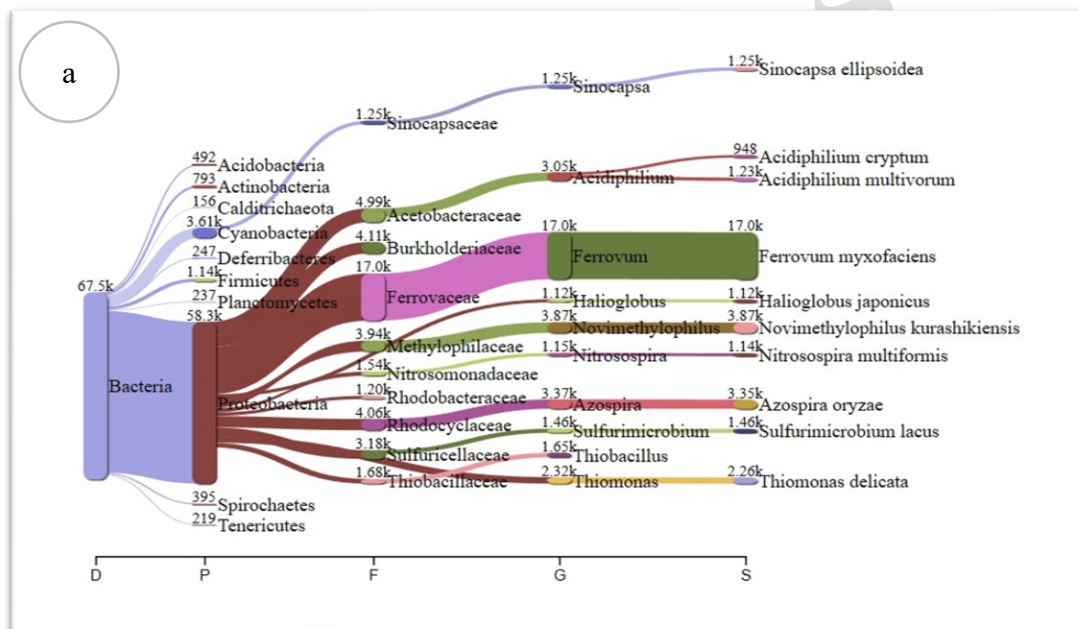


Figure 7. Krona diagram of the relative abundance of bacteria from post-coal mining ponds ≤ 5 years old (a) and ≥ 20 years old (b)

259 Proteobacteria were a very broad and diverse group of bacteria, which were often found in various
260 environments, including coal mine ponds ecosystems. Among proteobacteria, Betaproteobacteria and
261 Alphaproteobacteria were classes commonly found in highly acidic environments, such as coal post-
262 mining ponds. In addition, coal post-mining ponds tended to have highly acidic conditions due to
263 sulfide oxidation that produced sulfuric acid. Proteobacteria, including Betaproteobacteria and
264 Alphaproteobacteria, had several species that could survive and thrive in these extreme environments.
265 Proteobacteria had the metabolic ability to oxidize compounds found in mining environments, such
266 as sulfur and iron. This allowed them to use the resources available in coal post-mining ponds for
267 their growth. Some species of Proteobacteria, specifically in the Betaproteobacteria class, were
268 engaged in the biogeochemical cycling of minerals, such as sulfur and iron. The microbes could play
269 a role in the oxidation of sulfur and iron compounds commonly present in mining environment, as
270 well as in related chemical processes (Akimbekov et al., 2022).

271 Sankey Diagram (Figure 8) were used to visualize microbial species of the samples. At post-coal
272 mining ponds ≤ 5 years old for the species level, it was found that the top 10 dominant species were
273 *Sinocapsa ellipsoidea*, *Acidiphilium eryptum*, *Acidiphilium multivorum*, *Ferrovum myxofaciens*,
274 *Halioglobus japonicus*, *Novimenthylophilus kurashikiensis*, *Nitrosospira multiformis*, *Azospira*
275 *oryzae*, *Sulfurimicrobium lacus*, and *Thiomonas delicata*. The genus *Acidiphilium* were organisms
276 considered to be acidophilic because of their ability to survive and reproduce in an acidic environment
277 due to contamination by sulfuric acid, such as in post-mining ponds ≤ 5 years old with a value of 2-
278 3 (Stan-Lotter & Fendrihan, 2017). Compared to *Acidiphilium*, *Ferrovum myxofaciens* were bacteria
279 that were tolerant to highly acidic environments. These microbes could survive in conditions with
280 low pH caused by sulfuric acid resulting from the oxidation of sulfides in coal. In addition, *Ferrovum*
281 *myxofaciens* was bacteria that oxidized iron and used iron as an energy source. In coal post-mining
282 ponds, there were many oxidized iron minerals and *Ferrovum myxofaciens* could use this as an energy
283 source for their growth (Johnson et al., 2014). At post-coal mining ponds ≥ 20 years old for the species
284 level, it was found that the top 10 dominant species were *Armatimonas rosea*, *Haliscomenobacter*

285 *hydrossis*, *Limnobacter humi*, *Limnobacter thiooxidans*, *Polynucleobacter acidiphobus*,
 286 *Methylophilus methylotrophus*, *Methyloradius palustris*, *Methylotenera mobilis*, *Methylotenera*
 287 *versatilis*, and *Candidatus Pelagibacter ubique*. *Limnobacter* was a genus of bacteria commonly
 288 found in aquatic environments, including coal-mining ponds. In addition, *Limnobacter* were aerobic
 289 bacteria, showing the requirement of dissolved oxygen in water to respire. Post-coal mining ponds
 290 that were open to the atmosphere could have access to sufficient oxygen to support *Limnobacter*
 291 growth. *Limnobacter thiooxidans* had the ability to oxidize sulfur compounds, such as sulfide and
 292 thiosulfate, as an energy source (Spring et al., 2001). In post-coal mining ponds, the sulfur compounds
 293 could be present due to the oxidation process of sulfide minerals in coal.



294

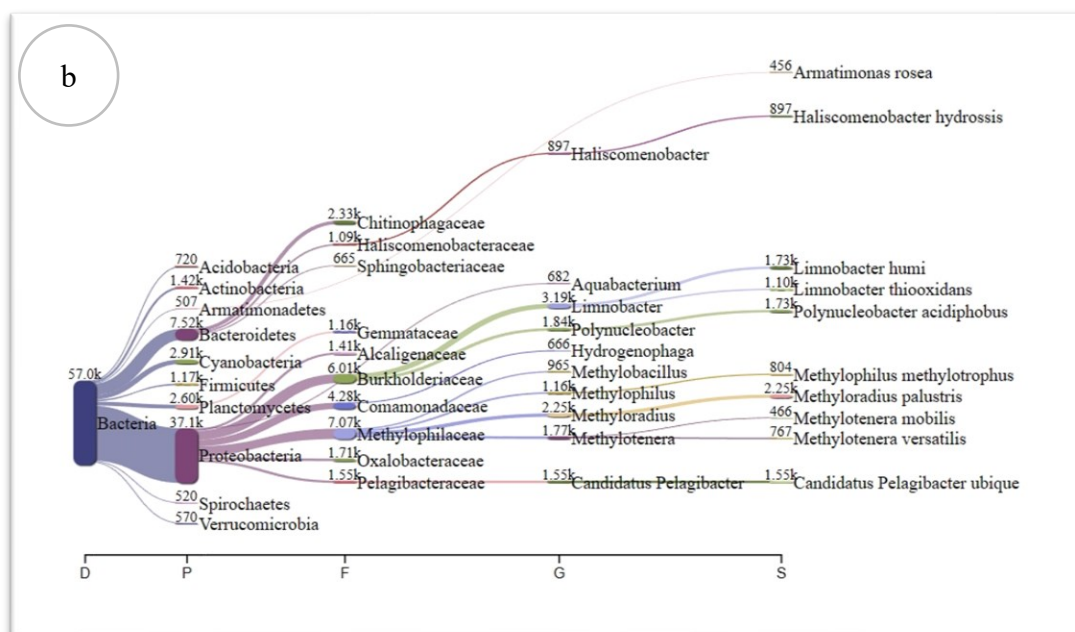


Figure 8. Sankey diagram of the relative abundance of bacteria from post-coal mining ponds ≤ 5 years (a) and ≥ 20 years old (b)

In this study, microbial diversity was determined by the Shannon and Simpson index. The 2 indexes of bacteria diversity provided sample community composition by estimating evenness (Simpson index) and richness (Shannon index). Richness was also measured using 2 types of non-parametric indexes, namely 1) Chao1 to estimate the number of species in community, and 2) ACE to estimate species range. This showed that the number of OTU at post-mining ponds ≥ 20 years old was greater than post-mining ponds ≤ 5 years old. Diversity and richness of bacteria in each post-coal mining ponds showed that diversity at post-mining ponds ≥ 20 years old was higher than at post-mining ponds ≤ 5 years old (Table 2).

Table 2. Bacteria diversity index from post-coal mining ponds water in Samarinda

Sample	Diversity index			
	Chao 1	ACE	Shannon	Simpson
Post-coal mining ponds ≤ 5 years old	3270.47	3301.63	4.18	0.91
Post-coal mining ponds ≥ 20 years old	5449.59	5481.24	6.24	0.99

Based on diversity index in Table 2, the highest diversity was found in post-coal mining ponds ≥ 20 years old. This was influenced by age and physicochemical parameters in post-coal mining ponds. A

309 study conducted by Cole et al., (2019) suggested that ponds age and physicochemical parameters,
310 such as pH and temperature had a significant effect on water quality, abundance of bacteria, and
311 phytoplankton.

312 **4. Conclusion**

313 In conclusion, the results showed that ponds ≥ 20 years old had better water quality and higher
314 bacteria OTU compared to ponds ≤ 5 years old. The richness of phylum, class, family, and genus of
315 bacteria community at post-coal mining ponds ≥ 20 years old was higher compared to ponds ≤ 5 years
316 old. In addition, increasing the abundance of SOB could reduce sulfate concentration of ponds water.
317 The better water quality and higher bacteria community structure at ponds ≥ 20 years old showed the
318 successful remediation of ponds water by SOB.

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