

# Facile prepared Fe<sub>3</sub>O<sub>4</sub> nanoparticles as a nano-catalyst on photo-fenton process to remediation of methylene blue dye from water: characterisation and optimization

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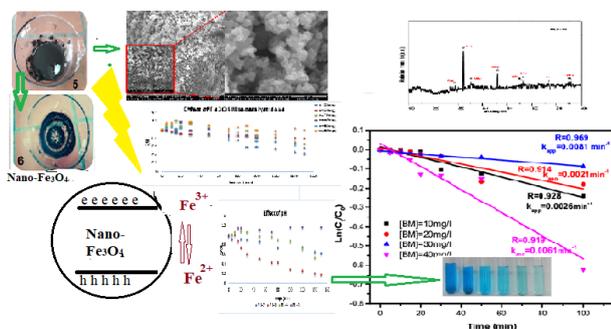
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## Graphical abstract



## Abstract

In this research, a simple method for the preparation of Fe<sub>3</sub>O<sub>4</sub> nanoparticles (Fe<sub>3</sub>O<sub>4</sub>NPs) with an average size of 38.05 nm via co-precipitation was investigated. X-ray diffraction (DRX), scanning electron microscopy (SEM), and infrared spectroscopy (FT-IR) were used to characterize obtained Fe<sub>3</sub>O<sub>4</sub> nanoparticles. These Fe<sub>3</sub>O<sub>4</sub>NPs were then used as nano-catalysts to degrade Methylene Blue (MB) in an aqueous solution via the Photo Fenton-like process. Also, under room solar light and low temperature, the photocatalytic activity of Fe<sub>3</sub>O<sub>4</sub>NPs for degrading MB was optimized through various experimental factors such as pH (ranging from 2 to 8), H<sub>2</sub>O<sub>2</sub> concentration (from 10<sup>-2</sup> to 5 × 10<sup>-1</sup>M), catalyst amount (20 to 60 mg), and target organic compound concentrations (10 to 40 mg/L). The optimal experimental conditions were found to be a pH of 3, H<sub>2</sub>O<sub>2</sub> (0.5M), a dye concentration of 40 mg/L, and 40 mg of Fe<sub>3</sub>O<sub>4</sub>NPs as nano-catalyst. These conditions led to a high degree of removal (>86%) of MB dye from water. The pseudo-second-order kinetic model was the suitable model to describe the degradation of MB dye with a coefficient value of 0.969. From this, it was concluded that Fe<sub>3</sub>O<sub>4</sub>NPs could act as an effective nano-catalyst for a sustainable

and environmentally friendly way to eliminate organic pollutants in water and wastewater.

**Keywords:** Fe<sub>3</sub>O<sub>4</sub>Nps, structural characterization, MB degradation, photo-fenton oxidation process, pseudo-first-order model, nanocatalyst.

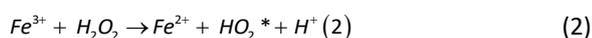
## 1. Introduction

Contamination of water with organic pollutants is a significant environmental problem that has negative impacts on both ecosystems and humans. Chemicals such as pesticides, herbicides, petroleum, and dyes are often released into water and wastewater through industrial and agricultural practices (Chowdhary *et al.* 2020).

Organic pollutants are harmful to aquatic life, the food chain, and water quality. They can also cause health problems for humans, such as cancer and neurological damage (Okoye *et al.* 2022). Therefore, the treatment of organic contaminated water and wastewater is crucial. Additionally, water scarcity due to drought has been a global issue in recent years. Thus, reusing treated wastewater is an effective approach to conserving freshwater resources (Manikandan *et al.* 2022). However, before releasing wastewater into the environment or using it for human consumption or agricultural irrigation, it must undergo treatment using methods such as filtration, bioremediation, and chemical treatments (Kesari *et al.* 2021). These methods aim to remove or degrade pollutants to ensure human safety and maintain the health and sustainability of the ecosystem (Kesari *et al.* 2021). Industrial activities are a significant source of organic pollutants. The textile industry, in particular, uses numerous organic compounds and dyes, which are discharged as effluent during dyeing and finishing processes. These organic dyes have low biodegradability, thus leading to the long-term pollution of water sources (Kesari *et al.* 2021). This contamination poses a threat to ecosystems and human health.

The photo-Fenton degradation process is a powerful technique for treating contaminated water, It is an advanced oxidation process (AOP) category, which involves the use of catalysts (materials) to generate highly reactive hydroxyl radicals to degrade pollutants in water and wastewater under light irradiation (Pandis *et al.* 2022). One of the advantages of this process are highly effective in removing various types of pollutants from wastewater, including organic and inorganic compounds. this process is considered a low-cost technology that does not require expensive reagents or complex equipment. It can be operated at ambient temperature and pressure conditions. Also, this process does not generate toxic by-products and can significantly reduce the amount of sludge generated during the treatment process (Zhang *et al.* 2019b). However, this process has disadvantages as the catalyst used can be deactivated over time due to fouling or poisoning by pollutants, which reduces the efficiency of the process. Also, this process is highly dependent on optimizing various parameters involved such as the pH of the media, and the nature of catalysts used, this process may not be suitable for the treatment of certain emergent contaminates with high concentrations (Zhang *et al.* 2019b).

Recently, a heterogenous Fenton-like process also gets considerable interest in the photodegradation of organic pollutants with high efficiency and environmentally benign operation, by using different types of catalysts, such as copper ferrate (CuFe<sub>2</sub>O<sub>4</sub>) doped with Al for degradation tetracycline (> 95%) (Zhong *et al.* 2023), FeCeO<sub>x</sub> for degradation of sulfamethazine (> 90%) (Liu *et al.* 2023), Fe<sub>2</sub>O<sub>3</sub>/TiO<sub>2</sub> supported by montmorillonite and aluminum pillared clays for removal three emerging organic contaminates (triclosan, dichlorophenol, and bisphenol ) ( 100%), ZnO/Fe<sub>2</sub>O<sub>3</sub>@Gr catalyst explored on mineralization of 58.5% to 92.8% of cationic dye crystal violet (Cardona *et al.* 2023), Mn<sub>3</sub>O<sub>4</sub>/λ-MnO<sub>2</sub> which reached 90.6% of mineralization phenol (Xing *et al.* 2023), modified CuFeO<sub>2</sub> for degradation of ofloxacin (94.2%)(Zhang *et al.* 2023), and FeMnO<sub>x</sub> nanocube for degradation of Rhodamine (>98%)(Su *et al.* 2023). The process based on continually production of \*OH following equations:



On the other hand, recently promising nanomaterials such as Fe<sub>3</sub>O<sub>4</sub>, ZnO, and TiO<sub>2</sub> which that characterized by their high surface area, can be improved and reach efficiency on the photodegradation of organic pollutants at the laboratory scale. Also, the using Fe<sub>3</sub>O<sub>4</sub> nanoparticles and doped Fe<sub>3</sub>O<sub>4</sub> on photocatalytic degradation of carbol fuchsin dye, furacilin, and doxorubicin hydrochloride drugs (Frolova and Sukhyy 2023; Koli *et al.* 2019; Weng *et al.* 2018; Zhang *et al.* 2019a), ZnO nanoparticles as nanocatalyst to remove of 2-Chlorophenol and on the degradation of lignin (Lenka and Badamali 2023; Zyoud *et al.* 2021), TiO<sub>2</sub> on photocatalytic degradation of

tebuconazole (TEB), Congo red, and cefotaxime (Abbood *et al.* 2023; Gianni *et al.* 2023; Obaiah *et al.* 2023).

Based on these findings, in the present study, i) a facile method was employed for the synthesis of nanoscale Fe<sub>3</sub>O<sub>4</sub> Nps particles, and their structural and morphological characteristics were also investigated, ii) additionally, under room sunlight and low temperature, the degradation potential of the obtained Fe<sub>3</sub>O<sub>4</sub> NPs was investigated for the removal of Methylene Blue dye from water, iii) experimental conditions affecting the degradation process of MB, including pH, the amount of nano-catalyst (Fe<sub>3</sub>O<sub>4</sub> NPs), concentration of H<sub>2</sub>O<sub>2</sub>, and the initial amount of the target compound (MB) were optimized. iv) The degradation kinetics of MB dye in the presence of Fe<sub>3</sub>O<sub>4</sub> NPs were examined using the pseudo-first-order model. As a result, under ambient conditions, Fe<sub>3</sub>O<sub>4</sub>NPs demonstrated high photocatalytic efficiency. These results are disclosed in detail in this study.

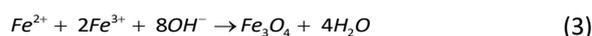
## 2. Experimental

### 2.1. Reagents

All reagents used were of chemical reagent grade with good purification. FeCl<sub>2</sub>·4H<sub>2</sub>O (99%), and FeCl<sub>3</sub>·6H<sub>2</sub>O (97%) were purchased from Sigma Aldrich, Spain. Hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>, 35%), sodium hydroxide (NaOH, 98%), and hydrochloric acid (HCl, 37%), were obtained from Panreac, Germany with analytical grade. Methylene blue (MB) (C<sub>16</sub>H<sub>18</sub>ClN<sub>3</sub>S. H<sub>2</sub>O) with a molecular weight of 319.85 g/mol and purity of 98%, which is chosen as the target organic pollutant was supplied by Biochem, Germany.

### 2.2. Chemical preparation of Fe<sub>3</sub>O<sub>4</sub>NPs

Fe<sub>3</sub>O<sub>4</sub> nanoparticles were prepared by co-precipitation method with some modifications according to Kushwaha & Chauhan, (2023) (Kushwaha and Chauhan 2023). Briefly, 16,25 g of FeCl<sub>3</sub>·6H<sub>2</sub>O and 6,35 g of FeCl<sub>2</sub>·4H<sub>2</sub>O were dissolved into 200 ml of distilled water. After stirring for 1h, chemical precipitation was achieved at a heated temperature of 35 °C under vigorous stirring by adding NaOH (2M) solution. The reaction system keeps at 70°C for 5 h with pH=12. Then, the suspension solution was cooled to room temperature; the precipitates were separated by a permanent magnet and washed with distilled water until pH neutral. Finally, Fe<sub>3</sub>O<sub>4</sub>NPs were washed with acetone and dried in the oven at 60-70°C. The chemical reaction can be expressed by (3):



### 2.3. Structural characterizations of synthesized Fe<sub>3</sub>O<sub>4</sub>NPs

Nanoparticles and Nanocomposites structures were characterized by X-ray diffractometry (XRD) (Rigaku MiniFlex 600 X-ray diffractometer Model) using Cu-K $\alpha$  radiation in the Bragg angle 2 $\theta$  range (10-80°) at 40 kV and 35 mA and rate of 0.04 degree in 4 seconds. The broadening was calculated from the (101) diffraction peak and the particle size was estimated from the following:

$$\text{Scherrer equation } D = \frac{0.89\lambda}{\Delta W \cos \theta} \quad (4)$$

Where:  $D$  is the average size of the crystallites,  $\Delta W$  is the full width at half maxima (DWHM),  $\lambda$  is the wavelength of radiation ( $\text{CuK}\alpha=1.5428 \text{ \AA}$ ), and  $\theta$  is the peak position (in degrees).

The morphology of particles was observed using a scanning electron microscope (SEM, FEI Quanta 650 FEG). The chemical composition of the synthesized materials was assessed by FT-IR spectroscopy with FTIR-S64 (Alpha Bruker). UV-Vis spectrometry (Shimadzu UV-1650-PC) was used to measure BM dye concentration.

#### 2.4. Photocatalytic process using Fe<sub>3</sub>O<sub>4</sub>NPs as nano-catalyst

The studies of the BM photodegradation were performed in the batch experiments under environmental conditions (under low temperature and room sunlight irradiation) by implementing a set of 250 mL conical flasks under a stirring rate of 150 rpm at 298 K. The photocatalytic process was conducted by varying different operation conditions, such as the initial concentration of BM (10 – 40 mg/L), the concentration of peroxide hydrogen H<sub>2</sub>O<sub>2</sub> (0.1 to 1 M), pH was adjusted over the range pH 2-4 using HCl and NaOH (1M), the dosage of nan-catalyst (Fe<sub>3</sub>O<sub>4</sub>NPs) in range of (0.4 to 1 g/L) and reaction time over 120 min. photocatalysis reactions were carried out with a thermostatic shaker at 120 rpm.

The removal percent of BM ( $R\%$ ) was calculated by using the following equation (5):

$$R\% = \left( \frac{C_0 - C_t}{C_0} \right) \times 100 \quad (5)$$

where  $C_0$  is the initial BM concentration in the solution ( $\text{mg L}^{-1}$ ), and  $C_t$  is the BM concentration at time ( $t$ ) of the removal process.

### 3. Results and discussion

#### 3.1. Structural characterization of Fe<sub>3</sub>O<sub>4</sub>NPs

##### 3.2. SEM analysis

To assess the morphology and size of the Fe<sub>3</sub>O<sub>4</sub> nanoparticles, a scanning electron microscope (SEM) is used. As shown in Figure 1, SEM images were obtained at various magnifications (Figure 1). It is clearly found that Fe<sub>3</sub>O<sub>4</sub> nanoparticles are synthesized and crystallization with successfully. The morphology of the synthesized nanocomposite was a magnetite-rich powder with uniformly sized, homogeneously dispersed nanoparticles that had strong agglomeration. The nanoparticles' confirmed diameter ranges from 0.02 to 0.05  $\mu\text{m}$ . Similar nano-sphere morphologies were reported for Fe<sub>3</sub>O<sub>4</sub> nanoparticles synthesized via other methods, such as hydrothermal (Daou *et al.* 2006), sol-gel (Lemine *et al.* 2012), and Conventional co-precipitation method (Pham *et al.* 2019).

##### 3.3. FTIR analysis

The structural and chemical properties of the synthesized nanoparticles were determined using the FT-IR spectra. Figure 2 shows FT-IR spectra of Fe<sub>3</sub>O<sub>4</sub>NPs. As seen, Figure 2 indicates the existence of spectra peaks at 3392 and 1635  $\text{cm}^{-1}$ , which correspond to a hydroxyl group (-OH)

stretching and bending vibration on Fe<sub>3</sub>O<sub>4</sub>Nps surface, respectively, as reported in the literature (An *et al.* 2017; Lezner *et al.* 2012; Nijpanich *et al.* 2021; Pucar Milidrag *et al.* 2022). Additionally, stretching vibrations associated with (-Fe-O-) that were found in the FTIR spectra of Fe<sub>3</sub>O<sub>4</sub>Nps at wavelengths of 982, 691, 618, and 511 $\text{cm}^{-1}$  came from ultrafine magnetite powders (Abbas *et al.* 2014; Nijpanich *et al.* 2021; Pucar Milidrag *et al.* 2022).

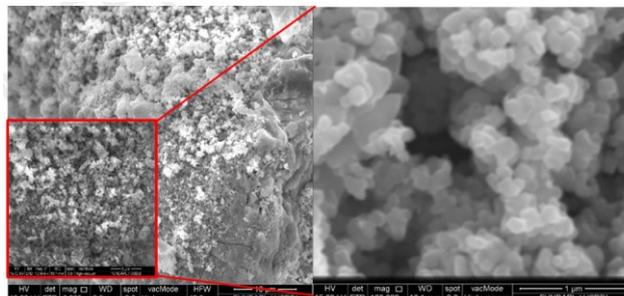


Figure 1. Scanning electron microscope (SEM) micrographs of the synthesized Fe<sub>3</sub>O<sub>4</sub>NPs in various magnifications

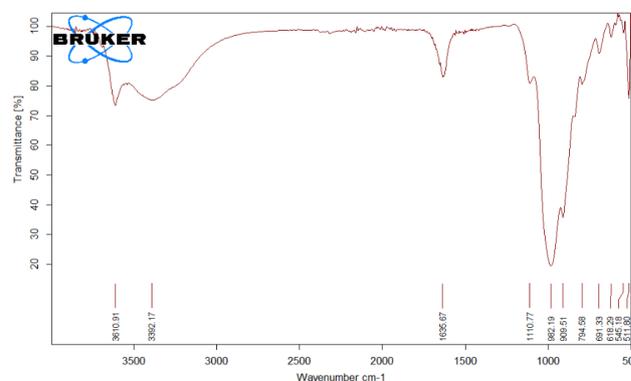
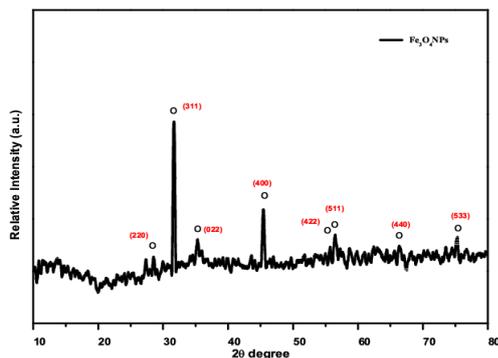


Figure 2. FTIR spectra of synthesized Fe<sub>3</sub>O<sub>4</sub>NPs

#### 3.4. X-ray diffraction analysis

The X-ray diffraction (XRD) method is used to determine the products' crystalline phase, phase structure, and crystallite size. XRD results of synthesized Fe<sub>3</sub>O<sub>4</sub>NPs samples are revealed in Figure 3. Softwares Match! and X'Pert High Score are used to treat RDX spectra, and the average crystallite sizes of samples are calculated with the Debye-Scherrer formula.

The diffraction pattern of Fe<sub>3</sub>O<sub>4</sub>Nps is illustrated in Figure 3, which indicates the major diffraction peaks of Fe<sub>3</sub>O<sub>4</sub>NPs are observed at 29°, 32°, 35.5°, 45.1°, 53.4°, 57.0°, 62.33°, and 75.0°, which correspond to the crystal planes (220), (311), (022), (400), (422), (511), (440), and (533), respectively. These peaks are consistent with the data recorded in reference to JCPDS card No. 19-0629. Additionally, the X-ray peaks of maghemite (210) and (211) are not apparent in the XRD patterns of nanoparticles. The calculated lattice parameters are  $\Delta a = 0.5245 \pm 0.0004$  (JCPDS file 19-629 used for calculation). This result confirmed the cubic inverse spinel structure of Fe<sub>3</sub>O<sub>4</sub>NPs, which also indicates the successful synthesis of Fe<sub>3</sub>O<sub>4</sub> nanoparticles. Also, the XRD of Fe<sub>3</sub>O<sub>4</sub>NPs reveals that the oxide present is principally magnetite. The average size of Fe<sub>3</sub>O<sub>4</sub> nanoparticles calculated according to the Scherrer equation is 38.05 nm.



**Figure 3.** XRD analysis of  $\text{Fe}_3\text{O}_4$  nanoparticles

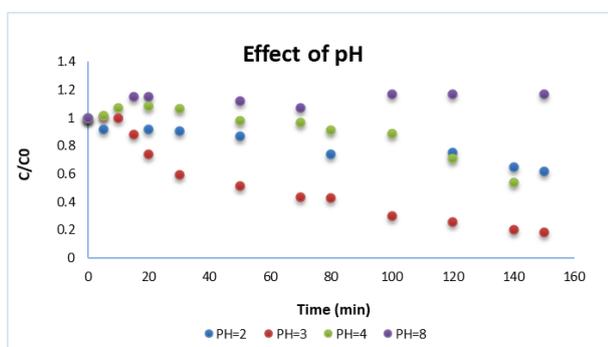
### 3.5. Photocatalytic activity study of $\text{Fe}_3\text{O}_4$ NPs

To investigate the photocatalytic degradation of organic pollutants with the nanoparticles under visible light, MB was selected as a model contaminant for photocatalytic decolorization.

In this study, the photocatalytic degradation of BM was investigated using nano-catalysts  $\text{Fe}_3\text{O}_4$ NPs in the presence of  $\text{H}_2\text{O}_2$  and under visible light, as well as the effect of parameters on the performance of the photocatalytic process such as solution pH, photo-catalyst dosage,  $\text{H}_2\text{O}_2$  concentration, and initial dye concentration were evaluated.

#### 3.5.1. Effect of pH

The effects of pH on the photocatalytic activity of BM were investigated. As shown in Figure 4, the efficiency of BM degradation using  $\text{Fe}_3\text{O}_4$ Nps at constant conditions (initial concentration of  $\text{BM}=40$  mg/L, duration time of 150 min) is maximum at  $\text{pH}=3$  (85 %) and minimum at  $\text{pH}=8$  (25 %), indicating the importance of pH value on BM degradation in photocatalytic process under visible light. At acidic pH,  $\text{Fe}_3\text{O}_4$  nanoparticles can undergo chemical transformations that promote the degradation of MB. For instance, the reduction of  $\text{Fe}^{3+}$  ions to  $\text{Fe}^{2+}$  ions is favored in an acidic environment. These  $\text{Fe}^{2+}$  ions can participate in Fenton-like reactions, generating highly reactive hydroxyl radicals ( $\cdot\text{OH}$ ) that are potent oxidants. These hydroxyl radicals can effectively degrade the MB molecules into by products (less toxic compounds).

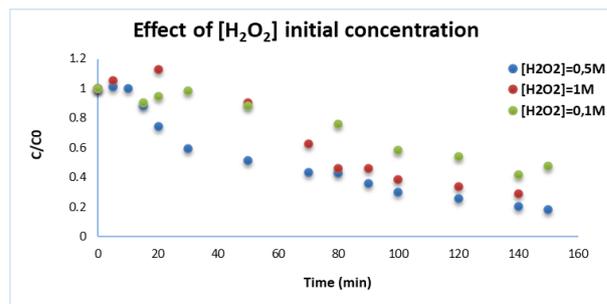


**Figure 4.** Effect of pH on photocatalytic removal efficiency of BM using  $\text{Fe}_3\text{O}_4$ Nps

#### 3.5.2. Effect of $\text{H}_2\text{O}_2$ amount

The  $\cdot\text{OH}$  radical produced by hydrogen peroxide ( $\text{H}_2\text{O}_2$ ) is one of the most effective initiators for increasing fast

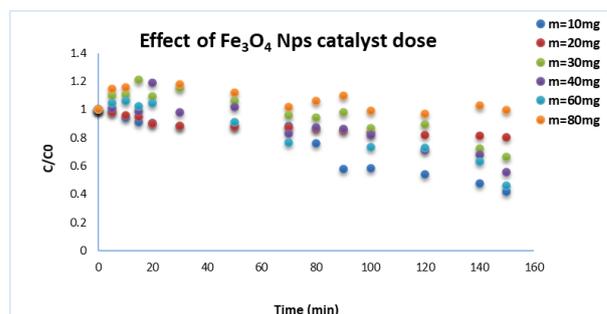
degradation efficiency. Figure 5 shows the role of  $\text{H}_2\text{O}_2$  in the degradation of MB using  $\text{Fe}_3\text{O}_4$ NPs in the presence of natural light. The MB dye degradation experiments were carried out with different concentrations of  $\text{H}_2\text{O}_2$  (30%) in three values: 0.1, 0.5, and 1M. The efficiency of MB degradation was 83% with 0.5 M of  $[\text{H}_2\text{O}_2]$ . We could also see that the degradation is better with concentrations of  $[\text{H}_2\text{O}_2]$ : 1 M, it reaches 75% of the removal rate. For 0.1M we notice a degradation of 60%. This can be explained by the fact that the increase of the  $\text{H}_2\text{O}_2$  concentration allows the production of a higher quantity of hydroxyl radicals ( $\text{OH}^\cdot$ ) and then a massive oxidation of the organic pollutant (Methylene Blue).



**Figure 5.** Effect of  $\text{H}_2\text{O}_2$  concentration for enhanced degradation of MB dye

#### 3.5.3. Effect of nano-photocatalyst ( $\text{Fe}_3\text{O}_4$ NPs) amount

The effect of  $\text{Fe}_3\text{O}_4$ NPs doses on MB removal was investigated, with photocatalyst amounts ranging from 10 to 80 mg for MB. As shown in Figure 6, increasing the amount of nano-catalyst from 10 to 40 mg increases the MB removal efficiency. However, increasing the nano-photocatalyst amount above 40 mg reduced the removal efficiency, which could be due to decreased light penetration into the solution and increased light dispersion from the surface of nanoparticles (Ebrahimi *et al.* 2019). As a result, the optically activated volume decreases, and a small amount of the suspension becomes activated. As a result, 40 mg of nano-catalyst was chosen as the optimum amount.



**Figure 6.** Effect of catalyst mass  $\text{Fe}_3\text{O}_4$ NPs on MB removal

#### 3.5.4. Effect of initial MB concentration

The initial concentration of the pollutant is a critical parameter in the photocatalytic degradation of organic pollutants. The effect of initial MB concentration at four different amounts (10, 20, 30, and 40 mg/l) on photocatalytic degradation was tested at a specific pH value and optimum amount of nano-catalyst, which shows that as the initial concentration of MB increases, the photocatalytic degradation process decreases until the 100 min.

As shown in Figure 7, the photocatalytic process is more effective at a high MB concentration of 40 mg/l, with 80 % degradation in an exposure duration of 150 min under visible light.

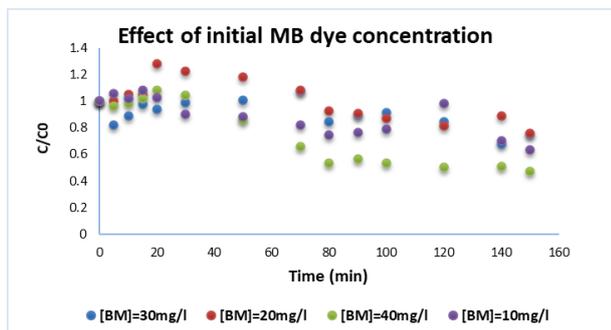


Figure 7. Effect of MB concentration on degradation process

### 3.6. Kinetic study of degradation MB dye

In order, to evaluate the photocatalytic degradation process of MB dye we applied a pseudo-first-order kinetic model using obtained experimental data. The apparent reaction rate constant ( $-k_{app}$ ) was determined using the first-order reaction model as a following equation (6):

$$\ln \frac{C_t}{C_0} = -k_{app} \cdot t \quad (3)$$

Where, two concentrations of pollutants,  $C_0$  and  $C_t$ , at initial and at time  $t$  (min) of the reaction, respectively, along with a reaction rate constant,  $k_{app}$ , which indicates the efficacy of pollutant elimination. The degradation kinetic constants of Fe<sub>3</sub>O<sub>4</sub>NPs for MB at different concentrations were obtained by first-order kinetic model fitting, indicating that the degradation rate ( $k_{app}$ ) ranging from 0.021 to 0.0081 min<sup>-1</sup>, and with coefficient  $R^2 > 0.91$  for all tested concentrations of MB (Figure 8).

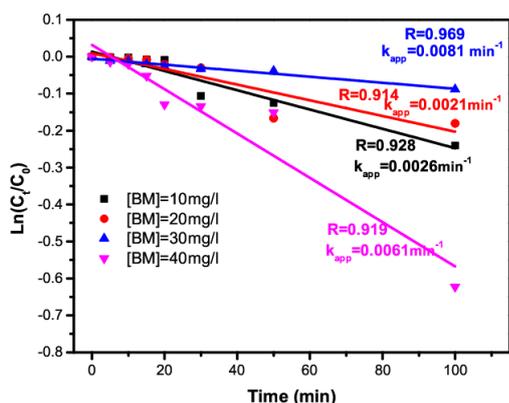


Figure 8. Linearized plots of pseudo-first-order MB decolorization at different concentrations

## 4. Conclusions

In this study, we investigated the degradation of Methylene Blue (MB) in aqueous solution using nanocatalysts Fe<sub>3</sub>O<sub>4</sub>NPs. DRX, FT-IR, and UV-vis were used to characterize their morphology. The Fenton photocatalytic activity of Fe<sub>3</sub>O<sub>4</sub>NPs to degrade MB was optimized using various experiment factors such as pH (2 to 8), H<sub>2</sub>O<sub>2</sub> concentration (0.1 to 1M), amount of

nanocomposites catalysts (10 to 80 mg), and MB concentrations. The results show that Fe<sub>3</sub>O<sub>4</sub>NPs have a high efficiency in degrading MB under the conditions of H<sub>2</sub>O<sub>2</sub> (10<sup>-1</sup>M), catalyst (40 mg), MB (30 mg/L), and pH=3. However, Fe<sub>3</sub>O<sub>4</sub>NPs have an optimum efficiency under the conditions of H<sub>2</sub>O<sub>2</sub> (10<sup>-1</sup>M), catalyst (40 mg), MB (40 mg/L), and pH=3. Also, the pseudo-second-order kinetic model was found as the best model to describe the photocatalytic degradation of MB. The pH value is an important factor in the efficiency of the oxidation system; in this study, pH=3 demonstrated high efficiencies. The current Photo-Fenton oxidation process using Fe<sub>3</sub>O<sub>4</sub>NPs as a nano-catalyst can be used as an alternative process to treatment to degrade organic pollutants in water and wastewater.

### Conflict of interest

The authors declare that they have no conflicts of interest.

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