

1 **Facile Prepared Fe<sub>3</sub>O<sub>4</sub> Nanoparticles as a Nano-Catalyst on photo-Fenton process**  
 2 **to remediation of methylene blue dye from water: Characterisation and**  
 3 **Optimization**

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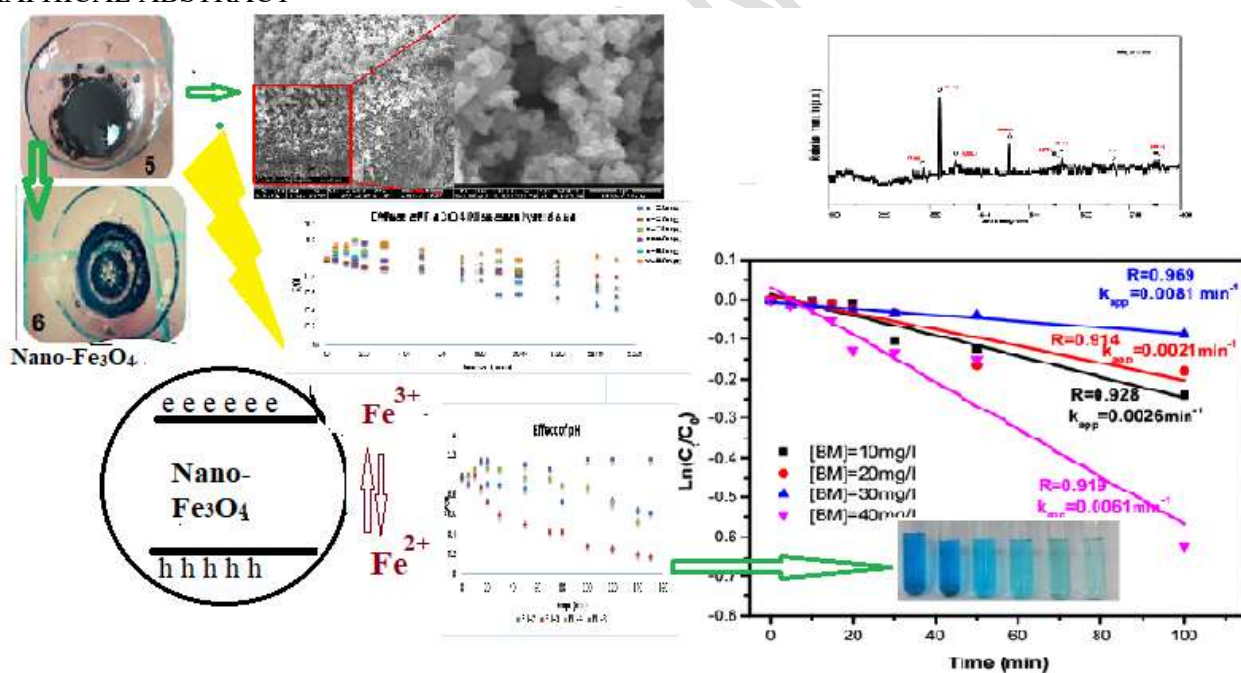
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14 GRAPHICAL ABSTRACT



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## ABSTRACT

In this research, a simple method for the preparation of Fe<sub>3</sub>O<sub>4</sub> nanoparticles (Fe<sub>3</sub>O<sub>4</sub>NPs) with an average size of 38.05 nm via co-precipitation was investigated. X-ray diffraction (DRX), scanning electron microscopy (SEM), and infrared spectroscopy (FT-IR) were used to characterize obtained Fe<sub>3</sub>O<sub>4</sub> nanoparticles. These Fe<sub>3</sub>O<sub>4</sub>NPs were then used as nano-catalysts to degrade Methylene Blue (MB) in an aqueous solution via the Photo Fenton-like process. Also, under room solar light and low temperature, the photocatalytic activity of Fe<sub>3</sub>O<sub>4</sub>NPs for degrading MB was optimized through various experimental factors such as pH (ranging from 2 to 8), H<sub>2</sub>O<sub>2</sub> concentration (from 10<sup>-2</sup> to 5 × 10<sup>-1</sup>M), catalyst amount (20 to 60 mg), and target organic compound concentrations (10 to 40 mg/L). The optimal experimental conditions were found to be a pH of 3, H<sub>2</sub>O<sub>2</sub> (0.5M) , a dye concentration of 40 mg/L, and 40 mg of Fe<sub>3</sub>O<sub>4</sub>NPs as nano-catalyst. These conditions led to a high degree of removal (>86%) of MB dye from water. The pseudo-second-order kinetic model was the suitable model to describe the degradation of MB dye with a coefficient value of 0.969. From this, it was concluded that Fe<sub>3</sub>O<sub>4</sub>NPs could act as an effective nano-catalyst for a sustainable and environmentally friendly way to eliminate organic pollutants in water and wastewater.

**Keywords:** Fe<sub>3</sub>O<sub>4</sub>Nps, structural characterization, MB degradation, photo-Fenton Oxidation process, pseudo-first-order Model, Nanocatalyst.

## 51 **1. Introduction**

52 Contamination of water with organic pollutants is a significant environmental problem that has  
53 negative impacts on both ecosystems and humans. Chemicals such as pesticides, herbicides, petroleum,  
54 and dyes are often released into water and wastewater through industrial and agricultural practices  
55 (Chowdhary et al. 2020).

56 Organic pollutants are harmful to aquatic life, the food chain, and water quality. They can also cause  
57 health problems for humans, such as cancer and neurological damage (Okoye et al. 2022). Therefore,  
58 the treatment of organic contaminated water and wastewater is crucial. Additionally, water scarcity  
59 due to drought has been a global issue in recent years. Thus, reusing treated wastewater is an effective  
60 approach to conserving freshwater resources (Manikandan et al. 2022). However, before releasing  
61 wastewater into the environment or using it for human consumption or agricultural irrigation, it must  
62 undergo treatment using methods such as filtration, bioremediation, and chemical treatments (Kesari  
63 et al. 2021). These methods aim to remove or degrade pollutants to ensure human safety and maintain  
64 the health and sustainability of the ecosystem (Kesari et al. 2021). Industrial activities are a significant  
65 source of organic pollutants. The textile industry, in particular, uses numerous organic compounds and  
66 dyes, which are discharged as effluent during dyeing and finishing processes. These organic dyes have  
67 low biodegradability, thus leading to the long-term pollution of water sources (Kesari et al. 2021). This  
68 contamination poses a threat to ecosystems and human health.

69 The photo-Fenton degradation process is a powerful technique for treating contaminated water, It is  
70 an advanced oxidation process (AOP) category, which involves the use of catalysts (materials) to  
71 generate highly reactive hydroxyl radicals to degrade pollutants in water and wastewater under light  
72 irradiation (Pandis et al. 2022). One of the advantages of this process are highly effective in removing  
73 various types of pollutants from wastewater, including organic and inorganic compounds. this process  
74 is considered a low-cost technology that does not require expensive reagents or complex equipment. it  
75 can be operated at ambient temperature and pressure conditions. also, this process does not generate

76 toxic by-products and can significantly reduce the amount of sludge generated during the treatment  
77 process [(Zhang et al. 2019b). However, this process has disadvantages as the catalyst used can be  
78 deactivated over time due to fouling or poisoning by pollutants, which reduces the efficiency of the  
79 process. also, this process is highly dependent on optimizing various parameters involved such as the  
80 pH of the media, and the nature of catalysts used, This process may not be suitable for the treatment  
81 of certain emergent contaminates with high concentrations (Zhang et al. 2019b).

82 Recently, a heterogenous Fenton-like process also gets considerable interest in the photodegradation  
83 of organic pollutants with high efficiency and environmentally benign operation, by using different  
84 types of catalysts, such as copper ferrate ( $\text{CuFe}_2\text{O}_4$ ) doped with Al for degradation tetracycline (> 95%)  
85 (Zhong et al. 2023),  $\text{FeCeO}_x$  for degradation of sulfamethazine (> 90%) (Liu et al. 2023),  $\text{Fe}_2\text{O}_3/\text{TiO}_2$   
86 supported by montmorillonite and aluminum pillared clays for removal three emerging organic  
87 contaminates (triclosan, dichlorophenol, and bisphenol ) ( 100%),  $\text{ZnO}/\text{Fe}_2\text{O}_3@\text{Gr}$  catalyst explored  
88 on mineralization of 58.5% to 92.8% of cationic dye crystal violet (Cardona et al. 2023),  $\text{Mn}_3\text{O}_4/\lambda\text{-}$   
89  $\text{MnO}_2$  which reached 90.6% of mineralization phenol (Xing et al. 2023), modified  $\text{CuFeO}_2$  for  
90 degradation of ofloxacin ( 94.2%)(Zhang et al. 2023), and  $\text{FeMnO}_x$  nanocube for degradation of  
91 Rhodamine (>98%)(Su et al. 2023). the process based on continually production of  $^*\text{OH}$  following  
92 equations (1 -2):  $\text{Fe}^{2+} + \text{H}_2\text{O}_2 \text{-----} \rightarrow ^*\text{OH} + \text{OH}^-$  (1) and  $\text{Fe}^{3+} + \text{H}_2\text{O}_2 \text{-----} \rightarrow \text{Fe}^{2+} + \text{HO}_2^* + \text{H}^+$ (2)

93 On the other hand, recently promising nanomaterials such as  $\text{Fe}_3\text{O}_4$ ,  $\text{ZnO}$ , and  $\text{TiO}_2$  which that  
94 characterized by their high surface area, can be improved and reach efficiency on the photodegradation  
95 of organic pollutants at the laboratory scale. Also, the using  $\text{Fe}_3\text{O}_4$  nanoparticles and doped  $\text{Fe}_3\text{O}_4$  on  
96 photocatalytic degradation of carbol fuchsin dye, furacilin, and doxorubicin hydrochloride drugs  
97 (Frolova and Sukhyy 2023; Koli et al. 2019; Weng et al. 2018; Zhang et al. 2019a),  $\text{ZnO}$  nanoparticles  
98 as nanocatalyst to remove of 2-Chlorophenol and on the degradation of lignin (Lenka and Badamali  
99 2023; Zyoud et al. 2021),  $\text{TiO}_2$  on photocatalytic degradation of tebuconazole (TEB), Congo red, and  
100 cefotaxime (Abbood et al. 2023; Gianni et al. 2023; Obaiah et al. 2023).

101 Based on these findings, in the present study, **i)** a facile method was employed for the synthesis of  
102 nanoscale  $\text{Fe}_3\text{O}_4$  Nps particles, and their structural and morphological characteristics were also  
103 investigated, **ii)** additionally, under room sunlight and low temperature, the degradation potential of  
104 the obtained  $\text{Fe}_3\text{O}_4$  NPs was investigated for the removal of Methylene Blue dye from water, **iii)**  
105 experimental conditions affecting the degradation process of MB, including pH, the amount of nano-  
106 catalyst ( $\text{Fe}_3\text{O}_4$  NPs), concentration of  $\text{H}_2\text{O}_2$ , and the initial amount of the target compound (MB) were  
107 optimized. **iv)** The degradation kinetics of MB dye in the presence of  $\text{Fe}_3\text{O}_4$  NPs were examined using  
108 the pseudo-first-order model. As a result, under ambient conditions,  $\text{Fe}_3\text{O}_4$ NPs demonstrated high  
109 photocatalytic efficiency. These results are disclosed in detail in this study.

## 110 **2. Experimental**

### 111 **2.1.Reagents**

112 All reagents used were of chemical reagent grade with good purification.  $\text{FeCl}_2 \cdot 4\text{H}_2\text{O}$  (99%), and  
113  $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$  (97%) were purchased from Sigma Aldrich, Spain. Hydrogen peroxide ( $\text{H}_2\text{O}_2$ , 35%),  
114 sodium hydroxide (NaOH, 98%), and hydrochloric acid (HCl, 37%), were obtained from Panreac,  
115 Germany with analytical grade. Methylene blue (MB) ( $\text{C}_{16}\text{H}_{18}\text{ClN}_3\text{S} \cdot \text{H}_2\text{O}$ ) with a molecular weight  
116 of 319.85 g/mol and purity of 98%, which is chosen as the target organic pollutant was supplied by  
117 Biochem, Germany.

### 118 **2.2. Chemical preparation of $\text{Fe}_3\text{O}_4$ NPs**

119  $\text{Fe}_3\text{O}_4$  nanoparticles were prepared by co-precipitation method with some modifications according to  
120 Kushwaha & Chauhan, (2023) (Kushwaha and Chauhan 2023). Briefly, 16,25 g of  $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$  and  
121 6,35 g of  $\text{FeCl}_2 \cdot 4\text{H}_2\text{O}$  were dissolved into 200 ml of distilled water. After stirring for 1h, chemical  
122 precipitation was achieved at a heated temperature of 35 °C under vigorous stirring by adding NaOH  
123 (2M) solution. The reaction system keeps at 70°C for 5 h with pH=12. Then, the suspension solution  
124 was cooled to room temperature; the precipitates were separated by a permanent magnet and washed

125 with distilled water until pH neutral. Finally, Fe<sub>3</sub>O<sub>4</sub>NPs were washed with acetone and dried in the  
126 oven at 60-70°C. the chemical reaction can be expressed by (3):



### 128 **2.3. Structural characterizations of synthesized Fe<sub>3</sub>O<sub>4</sub>NPs**

129 Nanoparticles and Nanocomposites structures were characterized by X-ray diffractometry (XRD)  
130 (Rigaku MiniFlex 600 X-ray diffractometer Model) using Cu-K $\alpha$  radiation in the Bragg angle 2 $\theta$  range  
131 (10-80°) at 40 kV and 35 mA and rate of 0.04 degree in 4 seconds. The broadening was calculated  
132 from the (101) diffraction peak and the particle size was estimated from the following Scherrer equation

133 
$$(4): D = \frac{0.89\lambda}{\Delta W \cos \theta} \quad (4)$$

134 Where:  $D$  is the average size of the crystallites,  $\Delta W$  is the full width at half maxima (DWHM),  $\lambda$  is  
135 the wavelength of radiation (CuK $\alpha$ =1.5428 Å), and  $\theta$  is the peak position (in degrees).

136 The morphology of particles was observed using a scanning electron microscope (SEM, FEI Quanta  
137 650 FEG). The chemical composition of the synthesized materials was assessed by FT-IR spectroscopy  
138 with FTIR-S64 (Alpha Bruker). UV-Vis spectrometry (Shimadzu UV-1650-PC) was used to measure  
139 BM dye concentration.

### 140 **2.4. Photocatalytic process using Fe<sub>3</sub>O<sub>4</sub>NPs as nano-catalyst**

141 The studies of the BM photodegradation were performed in the batch experiments under environmental  
142 conditions (under low temperature and room sunlight irradiation) by implementing a set of 250 mL  
143 conical flasks under a stirring rate of 150 rpm at 298 K. The photocatalytic process was conducted by  
144 varying different operation conditions, such as the initial concentration of BM (10 - 40 mg/L), the  
145 concentration of peroxide hydrogen H<sub>2</sub>O<sub>2</sub> (0.1 to 1 M), pH was adjusted over the range pH 2-4 using  
146 HCl and NaOH (1M), the dosage of nano-catalyst (Fe<sub>3</sub>O<sub>4</sub>NPs) in range of (0.4 to 1 g/L) and reaction  
147 time over 120 min. photocatalysis reactions were carried out with a thermostatic shaker at 120 rpm.

148 The removal percent of BM (R%) was calculated by using the following equation (5):

149 
$$R\% = \left(\frac{C_t}{C_0}\right) \times 100, \quad (5)$$

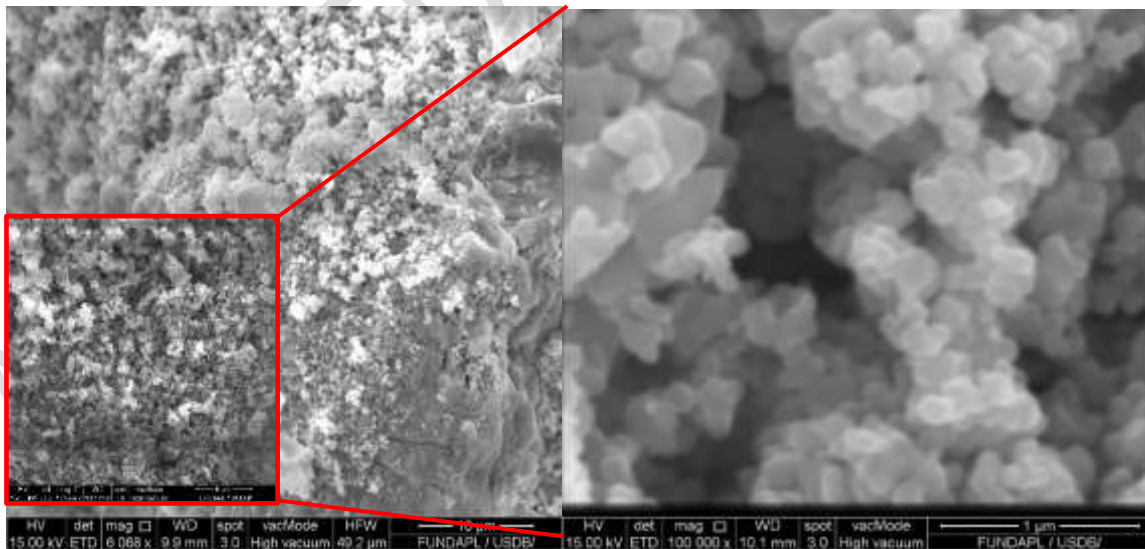
150 where  $C_0$  is the initial BM concentration in the solution ( $\text{mg L}^{-1}$ ), and  $C_t$  is the BM concentration at  
151 time (t) of the removal process.

### 152 3. Results and Discussion

#### 153 3.1. Structural characterization of $\text{Fe}_3\text{O}_4$ NPs

#### 154 3.2. SEM analysis

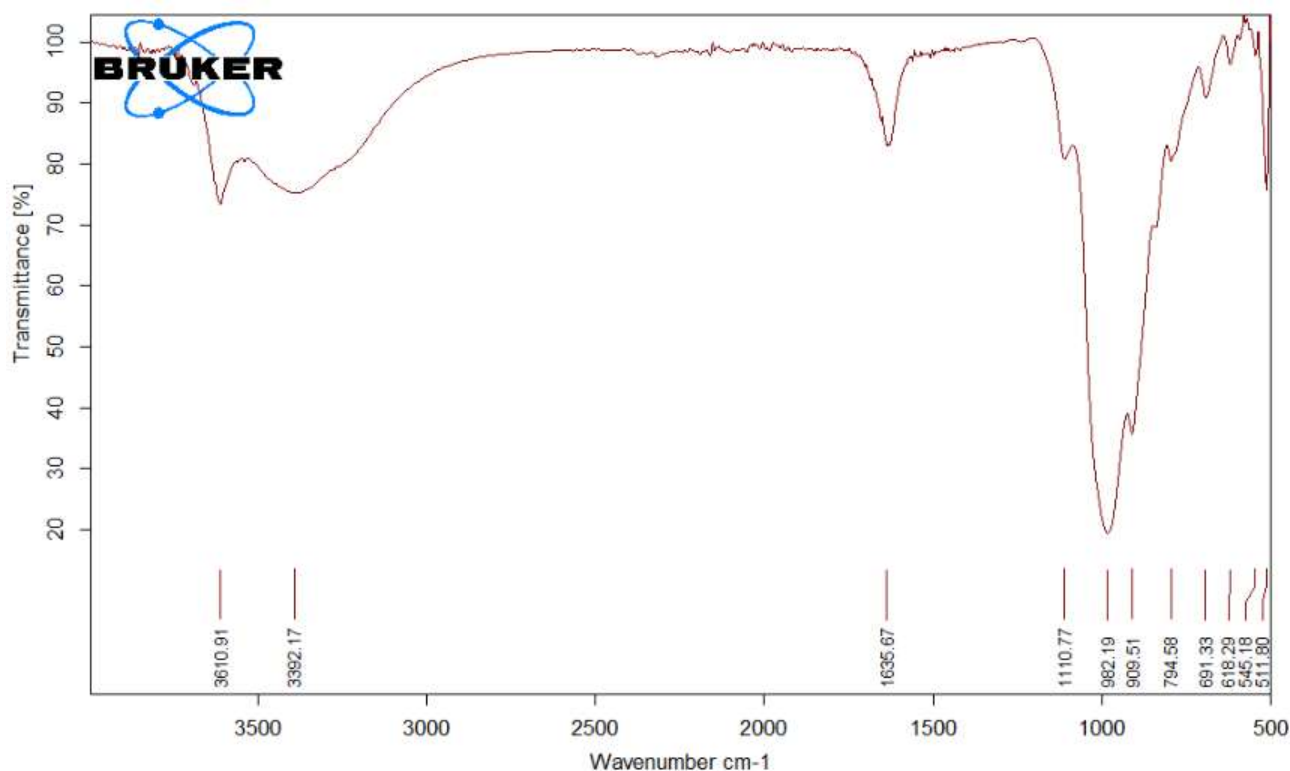
155 To assess the morphology and size of the  $\text{Fe}_3\text{O}_4$  nanoparticles, a scanning electron microscope (SEM)  
156 is used. As shown in Figure 1, SEM images were obtained at various magnifications (fig. 1), It is  
157 clearly found that  $\text{Fe}_3\text{O}_4$  nanoparticles are synthesized and crystallization with successfully. The  
158 morphology of the synthesized nanocomposite was a magnetite-rich powder with uniformly sized,  
159 homogeneously dispersed nanoparticles that had strong agglomeration. The nanoparticles' confirmed  
160 diameter ranges from 0.02 to 0.05  $\mu\text{m}$ . Similar nano-sphere morphologies were reported for  $\text{Fe}_3\text{O}_4$   
161 nanoparticles synthesized via other methods, such as hydrothermal (Daou et al. 2006), sol-gel (Lemine  
162 et al. 2012), and Conventional co-precipitation method (Pham et al. 2019).



163  
164 **Fig. 1.** Scanning electron microscope (SEM) micrographs of the synthesized  $\text{Fe}_3\text{O}_4$ NPs in various  
165 magnifications

#### 166 3.3. FTIR analysis

167 The structural and chemical properties of the synthesized nanoparticles were determined using the FT-  
168 IR spectra. Figure 2 shows FT-IR spectra of Fe<sub>3</sub>O<sub>4</sub>Nps. As seen, Figure 2 indicates the existence of  
169 spectra peaks at 3392 and 1635 cm<sup>-1</sup>, which correspond to a hydroxyl group (-OH) stretching and  
170 bending vibration on Fe<sub>3</sub>O<sub>4</sub>Nps surface, respectively, as reported in the literature (An et al. 2017;  
171 Lezner et al. 2012; Nijpanich et al. 2021; Pucar Milidrag et al. 2022). Additionally, stretching  
172 vibrations associated with (-Fe-O-) that were found in the FTIR spectra of Fe<sub>3</sub>O<sub>4</sub>Nps s at wavelengths  
173 of 982, 691, 618, and 511cm<sup>-1</sup> came from ultrafine magnetite powders (Abbas et al. 2014; Nijpanich  
174 et al. 2021; Pucar Milidrag et al. 2022).



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176  
177 **Fig 2.** FTIR spectra of synthesized Fe<sub>3</sub>O<sub>4</sub>Nps.

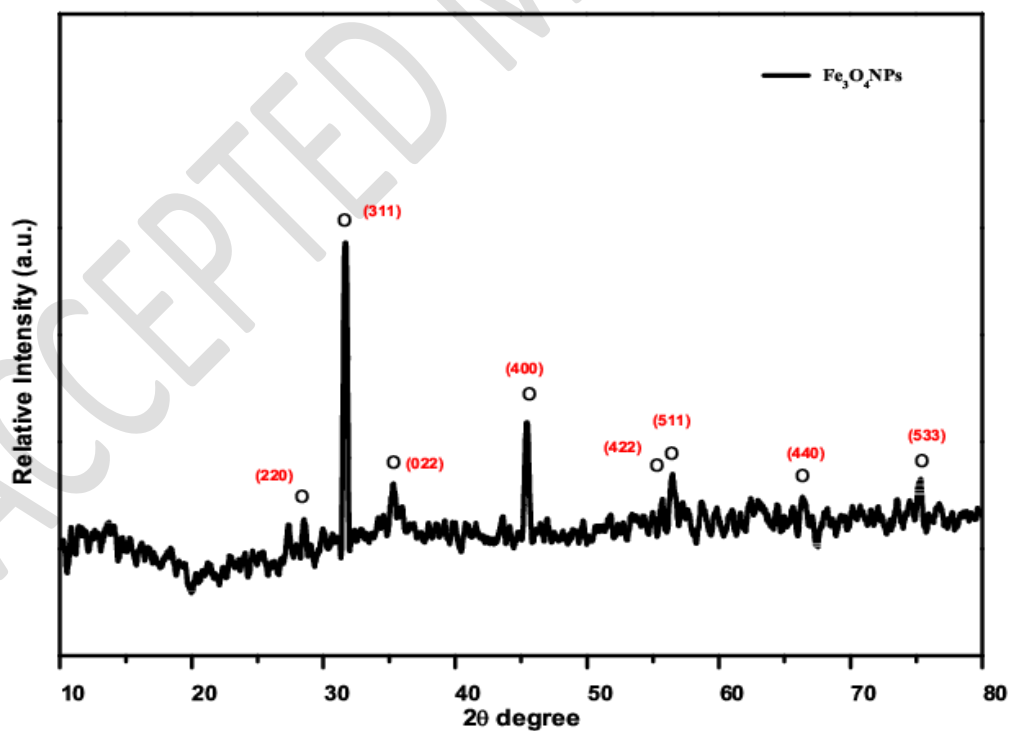
### 178 **3.4. X-ray diffraction analysis**

179 The X-ray diffraction (XRD) method is used to determine the products' crystalline phase, phase  
180 structure, and crystallite size. XRD results of synthesized Fe<sub>3</sub>O<sub>4</sub>NPs samples are revealed in Fig 3.



181 Softwares Match! and X'Pert High Score are used to treat RDX spectra, and the average crystallite  
182 sizes of samples are calculated with the Debye-Scherrer formula.

183 The diffraction pattern of  $\text{Fe}_3\text{O}_4\text{Nps}$  is illustrated in Figure 3, which indicates the major diffraction  
184 peaks of  $\text{Fe}_3\text{O}_4\text{NPs}$  are observed at  $29^\circ$ ,  $32^\circ$ ,  $35.5^\circ$ ,  $45.1^\circ$ ,  $53.4^\circ$ ,  $57.0^\circ$ ,  $62.33^\circ$ , and  $75.0^\circ$ , which  
185 correspond to the crystal planes (220), (311), (022), (400), (422), (511), (440), and (533), respectively.  
186 These peaks are consistent with the data recorded in reference to JCPDS card No. 19-0629.  
187 Additionally, the X-ray peaks of maghemite (210) and (211) are not apparent in the XRD patterns of  
188 nanoparticles. The calculated lattice parameters are  $\Delta a = 0.5245 \pm 0.0004$  (JCPDS file 19-629 used for  
189 calculation). This result confirmed the cubic inverse spinel structure of  $\text{Fe}_3\text{O}_4\text{NPs}$ , which also indicates  
190 the successful synthesis of  $\text{Fe}_3\text{O}_4$  nanoparticles. Also, the XRD of  $\text{Fe}_3\text{O}_4\text{NPs}$  reveals that the oxide  
191 present is principally magnetite. The average size of  $\text{Fe}_3\text{O}_4$  nanoparticles calculated according to the  
192 Scherrer equation is 38.05 nm.



193

194

Fig 3. XRD analysis of  $\text{Fe}_3\text{O}_4$  nanoparticles.

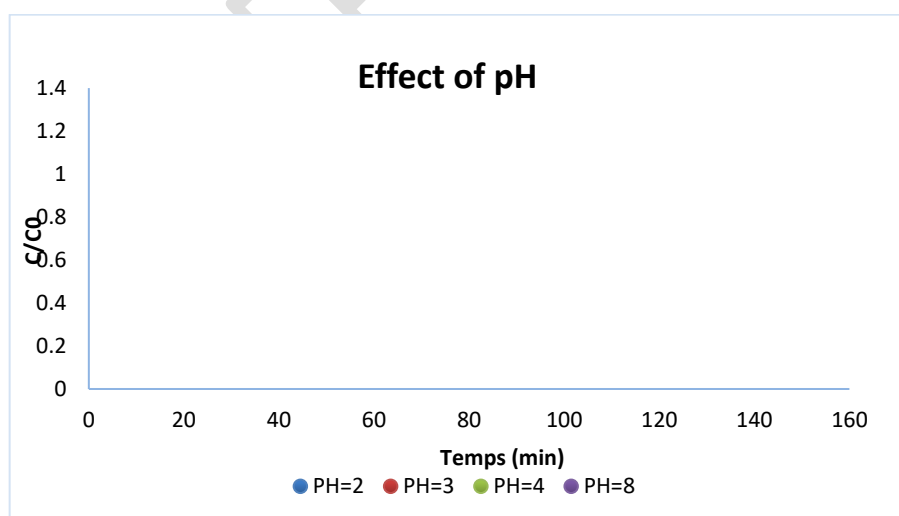
195 **3.5. Photocatalytic activity study of Fe<sub>3</sub>O<sub>4</sub>NPs**

196 To investigate the photocatalytic degradation of organic pollutants with the nanoparticles under visible  
197 light, MB was selected as a model contaminant for photocatalytic decolorization.

198 In this study, the photocatalytic degradation of BM was investigated using nano-catalysts Fe<sub>3</sub>O<sub>4</sub>NPs  
199 in the presence of H<sub>2</sub>O<sub>2</sub> and under visible light, as well as the effect of parameters on the performance  
200 of the photo-catalytic process such as solution pH, photo-catalyst dosage, H<sub>2</sub>O<sub>2</sub> concentration, and  
201 initial dye concentration were evaluated.

202 **3.5.1. Effect of pH**

203 The effects of pH on the photocatalytic activity of BM were investigated. As shown in Figure 4, the  
204 efficiency of BM degradation using Fe<sub>3</sub>O<sub>4</sub>Nps at constant conditions (initial concentration of BM=40  
205 mg/L, duration time of 150 min) is maximum at pH=3 (85 %) and minimum at pH=8 (25 %), indicating  
206 the importance of pH value on BM degradation in photocatalytic process under visible light. At acidic  
207 pH, Fe<sub>3</sub>O<sub>4</sub> nanoparticles can undergo chemical transformations that promote the degradation of MB.  
208 For instance, the reduction of Fe<sup>3+</sup> ions to Fe<sup>2+</sup> ions is favored in an acidic environment. These Fe<sup>2+</sup>  
209 ions can participate in Fenton-like reactions, generating highly reactive hydroxyl radicals (<sup>•</sup>OH) that  
210 are potent oxidants. These hydroxyl radicals can effectively degrade the MB molecules into by  
211 products (less toxic compounds).



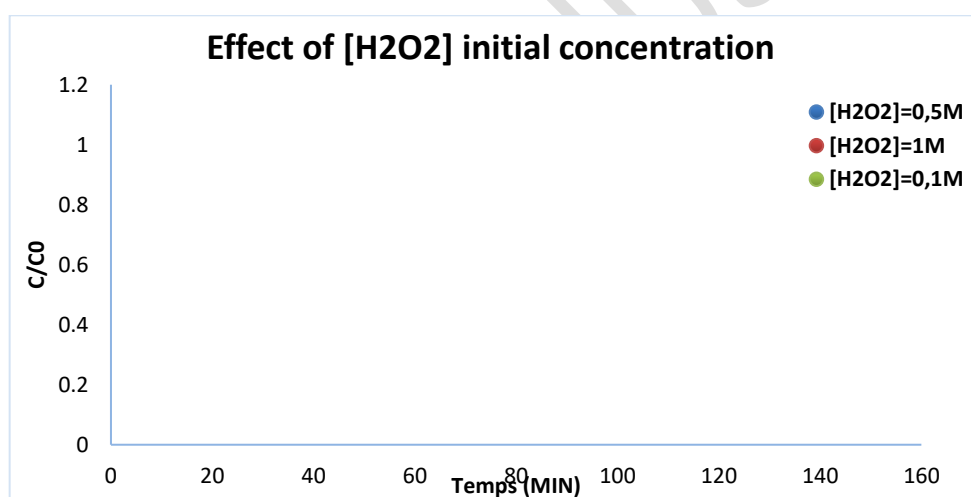
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**Fig. 4.** Effect of pH on photocatalytic removal efficiency of BM using Fe<sub>3</sub>O<sub>4</sub>Nps

### 214 3.5.2. Effect of H<sub>2</sub>O<sub>2</sub> Amount

215 The ·OH radical produced by hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>) is one of the most effective initiators for  
216 increasing fast degradation efficiency. Figure 5 shows the role of H<sub>2</sub>O<sub>2</sub> in the degradation of MB using  
217 Fe<sub>3</sub>O<sub>4</sub>NPs in the presence of natural light. The MB dye degradation experiments were carried out with  
218 different concentrations of H<sub>2</sub>O<sub>2</sub> (30%) in three values: 0.1, 0.5, and 1M. The efficiency of MB  
219 degradation was 83% with 0.5 M of [H<sub>2</sub>O<sub>2</sub>]. We could also see that the degradation is better with  
220 concentrations of [H<sub>2</sub>O<sub>2</sub>]: 1 M, it reaches 75% of the removal rate. For 0.1M we notice a degradation  
221 of 60%. This can be explained by the fact that the increase of the H<sub>2</sub>O<sub>2</sub> concentration allows the  
222 production of a higher quantity of hydroxyl radicals (OH°) and then a massive oxidation of the organic  
223 pollutant (Methylene Blue).

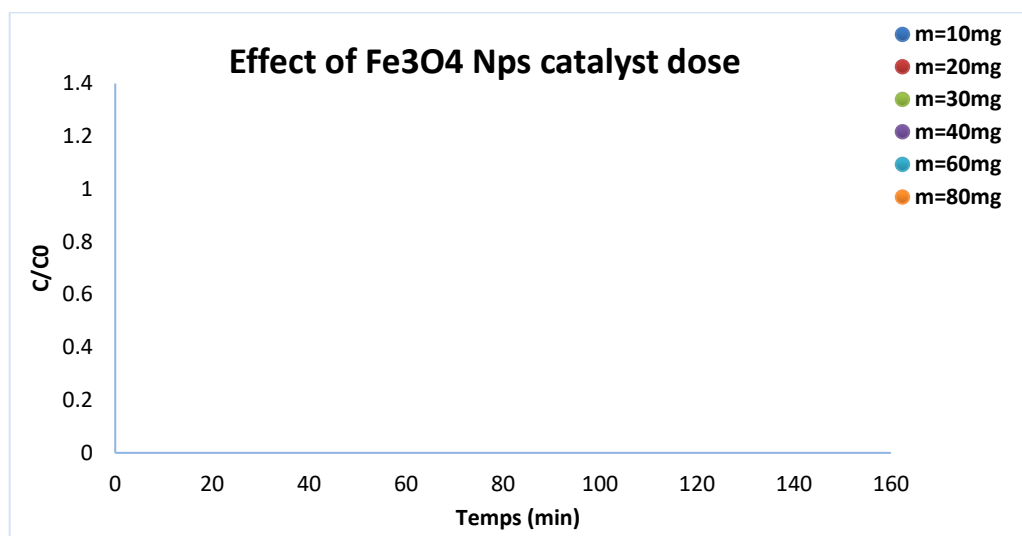


224 Fig.5. Effect of H<sub>2</sub>O<sub>2</sub> concentration for enhanced degradation of MB dye.

### 225 226 227 3.5.3. Effect of Nano-Photocatalyst (Fe<sub>3</sub>O<sub>4</sub>NPs) Amount

228 The effect of Fe<sub>3</sub>O<sub>4</sub>NPs doses on MB removal was investigated, with photocatalyst amounts ranging  
229 from 10 to 80 mg for MB. As shown in Figure 6, increasing the amount of nano-catalyst from 10 to 40  
230 mg increases the MB removal efficiency. However, increasing the nano-photocatalyst amount above  
231 40 mg reduced the removal efficiency, which could be due to decreased light penetration into the  
232 solution and increased light dispersion from the surface of nanoparticles (Ebrahimi et al. 2019). As a

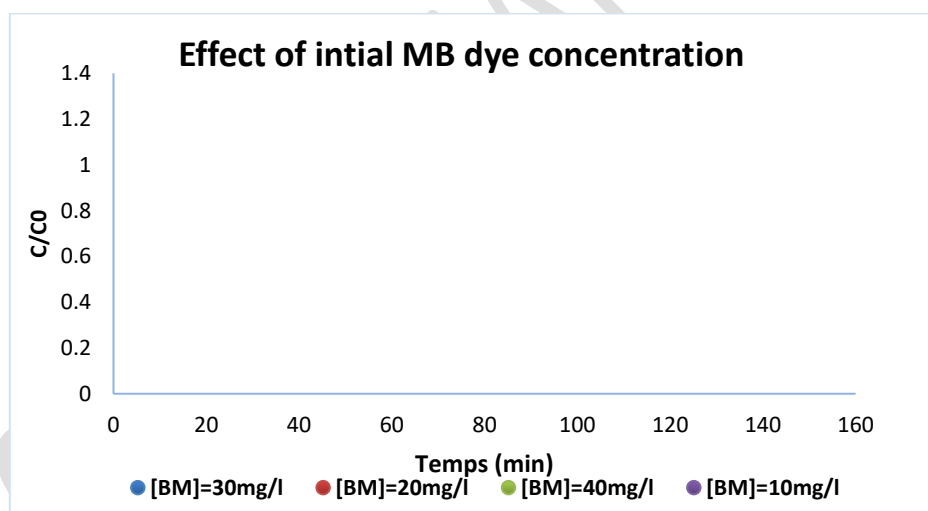
233 result, the optically activated volume decreases, and a small amount of the suspension becomes  
234 activated. As a result, 40 mg of nano-catalyst was chosen as the optimum amount.



235

236 **Fig.6.** Effect of catalyst mass Fe<sub>3</sub>O<sub>4</sub>NPs on MB removal

### 237 3.5.4. Effect of initial MB concentration



238

239 **Fig. 7.** Effect of MB concentration on degradation process.

240

241 The initial concentration of the pollutant is a critical parameter in the photocatalytic degradation of  
242 organic pollutants. The effect of initial MB concentration at four different amounts (10, 20, 30, and 40  
243 mg/l) on photo-catalytic degradation was tested at a specific pH value and optimum amount of nano-  
244 catalyst, which shows that as the initial concentration of MB increases, the photo-catalytic degradation

245 process decreases until the 100 min. As shown in Figure 7, the photocatalytic process is more effective  
 246 at a high MB concentration of 40 mg/l, with 80 % degradation in an exposure duration of 150 min  
 247 under visible light.

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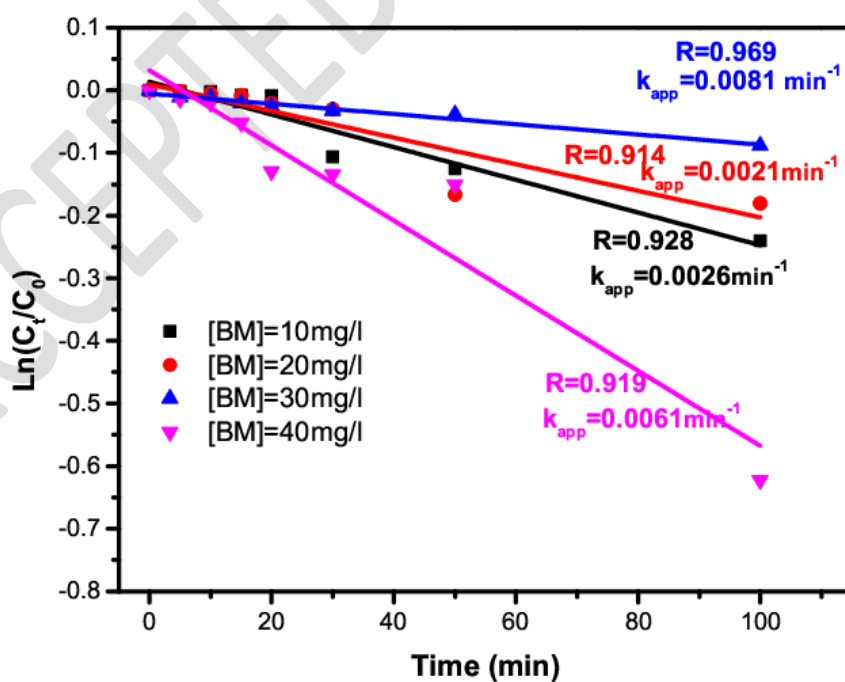
### 249 3.6.Kinetic study of degradation MB dye

250 In order, to evaluate the photocatalytic degradation process of MB dye we applied a pseudo-  
 251 first-order kinetic model using obtained experimental data. The apparent reaction rate constant ( $-k_{app}$ )  
 252 was determined using the first-order reaction model as a following equation (6):

$$253 \ln \frac{C_t}{C_0} = -K_{app} \cdot t \quad (6)$$

254 Where, two concentrations of pollutants,  $C_0$  and  $C_t$ , at initial and at time  $t$  (min) of the reaction,  
 255 respectively, along with a reaction rate constant,  $k_{app}$ , which indicates the efficacy of pollutant  
 256 elimination. The degradation kinetic constants of  $Fe_3O_4$ NPs for MB at different concentrations were  
 257 obtained by first-order kinetic model fitting, indicating that the degradation rate ( $K_{app}$ ) ranging from  
 258 0.021 to 0.0081 $min^{-1}$ , and with coefficient  $R^2 > 0.91$  for all tested concentrations of MB (Fig. 8).

259



260

261 **Fig.8.** Linearized plots of pseudo-first-order MB decolorization at different concentrations.

262

## 263 4. Conclusions

264 In this study, we investigated the degradation of Methylene Blue (MB) in aqueous solution using  
265 nanocatalysts Fe<sub>3</sub>O<sub>4</sub>NPs. DRX, FT-IR, and UV-vis were used to characterize their morphology. The  
266 Fenton photocatalytic activity of Fe<sub>3</sub>O<sub>4</sub>NPs to degrade MB was optimized using various experiment  
267 factors such as pH (2 to 8), H<sub>2</sub>O<sub>2</sub> concentration (0.1 to 1M), amount of nanocomposites catalysts (10  
268 to 80 mg), and MB concentrations. The results show that Fe<sub>3</sub>O<sub>4</sub>NPs have a high efficiency in degrading  
269 MB under the conditions of H<sub>2</sub>O<sub>2</sub> (10<sup>-1</sup>M), catalyst (40 mg), MB (30 mg/L), and pH=3. However,  
270 Fe<sub>3</sub>O<sub>4</sub>NPs have an optimum efficiency under the conditions of H<sub>2</sub>O<sub>2</sub> (10<sup>-1</sup>M), catalyst (40 mg), MB  
271 (40 mg/L), and pH=3. Also, the pseudo-second-order kinetic model was found as the best model to  
272 describe the photocatalytic degradation of MB. The pH value is an important factor in the efficiency  
273 of the oxidation system; in this study, pH=3 demonstrated high efficiencies. The current Photo-Fenton  
274 oxidation process using Fe<sub>3</sub>O<sub>4</sub>NPs as a nano-catalyst can be used as an alternative process to treatment  
275 to degrade organic pollutants in water and wastewater.

## 276 Conflict of interest

277 The authors declare that they have no conflicts of interest.

## 278 References

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