1	Diatom records in sediments for eutrophication process of Lake Xian'nv, China since
2	the mid-20th century
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- 9 GRAPHICAL ABSTRACT

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# 11 ABSTRACT

To well understand the eutrophication process for inland lakes where more lands had been used for agriculture and industry without adequate environmental protections, 9 surface sediments and 1 core sediment in Lake Xian'nv which was located in the middle of Jiangxi Province were chosen. Combined with <sup>210</sup>Pb and <sup>137</sup>Cs geochronology, nutrient index

including total nitrogen (TN), total phosphorus (TP), total organic carbon (TOC), Carbon 16 nitrogen ratio (C/N), diatoms assembles and TDI index of eutrophic diatoms were compared 17 18 quantitatively and conducted to elucidate the eutrophication process using statistical methods. The results showed a total of 155 years from 1863 to 2018 were recorded from the entire 19 sediment core. The nutrient level in 23 centimeters below the surface of the sediment core 20 where was synchronized with reservoir construction indicated TN, TP, TOC, and Carbon 21 22 nitrogen ratio had been gradually increased since the construction of Dam. In surface sediments, the three diatoms with higher average relative abundance were Achnanthidium 23 24 minutissimum, Navicula Bory, and Aulacoseira granulata with a value of 54.2 %, 29.4 %, and 15.3 % respectively. Significant changes of the diatom communities from 1955 to2009 had 25 been recorded, where diatom communities could be classified into four distinct periods. In the 26 first few years after the reservoir was built, the dominant diatoms were epiphytic and benthic 27 diatom communities represented by Navicula cincta, Fragilaria ulna, and Cocconeis 28 placentula in zone I on the bottom floor covered from 1955 to 1961. However, planktonic 29 diatom communities were dominant after 1983 which were mainly constituted by Aulacoseira 30 alpigena, Cyclostephanos dubius, Cyclotella meneghiniana, and Cyclotella radiosa. Both TDI 31 index of eutrophic diatoms and Principal Component Analysis (PCA) revealed the succession 32 from benthic diatoms to planktonic diatoms and changes in nutrient levels. 33

Keywords: Diatom community; sediments; nutrient levels; eutrophication process; Lake
Xian'nv

### 36 1. Introduction

37 As one of the most important ecosystems, lake could provide various valuable ecological 38 services for local economic and social development, including drinking water, agriculture,

fisheries, climate regulation, and flooding control, and also serve for biodiversity providing 39 aesthetic and economic benefits (Dearing et al., 2012; Liu et al., 2012; Laj et al., 2022). In 40 recent decades, a large number of artificial lakes have been formed along rivers due to the 41 construction of dams which were aimed at using water energy resources for power generation, 42 navigation, and irrigation (Zhang et al., 2021; Serafim-Júnior et al., 2016; Etourneau et al., 43 2013). However, urban artificial lakes which were used as urban water resources and 44 characterized by stagnant water bodies, fragile ecosystems, and irregular shapes were more 45 easily affected by interference and rapid urbanization from human activities (Yang et al., 2021; 46 Wang et al., 2013). These days, many artificial lakes were deteriorating as ecological system 47 degradation, area shrinking, pollution, and eutrophication due to excessive input of sediment, 48 nutrients, heavy metals, and organic pollutants (Huang et al., 2020; Wang et al., 2012). 49

Input of nitrogen (N) and phosphorus (P) could also cause a significant increase in algal 50 blooms resulting in eutrophication of the water body (Ly et al., 2021; Noges et al., 2020; Rast 51 and Thornton, 2015). Water eutrophication caused by superfluous nutrient contamination has 52 become one of the most prominent environmental problems on a global scale and poses a 53 health risk for aquatic ecosystems and humans (Zolitschka et al., 2015; Jeppesen et al., 2010; 54 Cardinale et al., 2012). Algal biomass and communities in sediments have been widely used 55 to determine the potential level and source nutrients from agricultural, industrial, and urban 56 discharges. The impact of hydrological conditions and climate changes on the level of algae 57 biomass has also been receiving more attention (Mihaljević et al., 2010; Zepernick et al., 58 59 2023). Long-term environmental monitoring for total nitrogen (TN), total phosphorus (TP), total organic carbon (TOC), and algal assembles at the water and sediment was critical for 60

exploring the underlying mechanism of the eutrophication process during different
decomposition (Fontana et al., 2014).

Fortunately, lake sediments could provide valuable information to elucidate catchment 63 variation and processes, which might reflect lake development and help to reconstruct 64 ecosystems (Costa-Böddeker et al., 2012; Yu et al., 2022). In paleolimnology, lacustrine 65 sediments preserved information in previous lacustrine periods has been used as an efficient 66 way to reconstruct the environment changes of lake ecosystems lacking long-term historical 67 records by using the stable isotopes and sedimentary bio information (Smol and Caraballo, 68 2008; Gell et al., 2005; McGowan et al., 2005; Hess et al., 2023). Data before area 69 development identified by stable isotopes technology could trace background threshold values 70 and provide a statistical basis for estimating current chemical and physical activities derived 71 from anthropogenic sources (Zhang et al., 2018; Liu et al., 2012). The sediment deposition 72 rate could also effectively record the erosion intensity and hydrodynamic conditions of the 73 basins (Zhang et al., 2018). Diatom communities were particularly sensitive to a range of 74 environmental factors including nutrient status, hydrology, and water quality, which has been 75 used to trace and evaluate historical ecological conditions for lake (Grundell et al., 2012; 76 77 Bhattacharya et al., 2016). Therefore, multi-proxy analyses based on lacustrine sedimentary stratigraphy had been increasingly accepted to research and explore the environmental and 78 ecological conditions of lakes (Pillsbury et al., 2021; Lan et al., 2018). 79

Since the 1950s, a great number of reservoirs have been constructed in Jiangxi Province to provide valuable clear fresh water, clean energy, and ecological services for downstream (Huang et al., 2017). Located in the northwest of Jiangxi Province, Lake Xian'nv

also known as Jiang'kou Reservoir served as a water source for City Xin'yu and a famous 83 scenic spot was appeared through dam construction on the River Yuan in 1958 (Ouyang et al., 84 85 2018; Zhang et al., 2015). However, rapid population growth and economic development in recent decades had exposed the lake ecosystem to a higher risk, including the decline of water 86 quality, eutrophication risk, as well as biodiversity degradation, and water ecological security 87 (Ji et al., 2020; Li et al., 2022). In order to identify underlying processes and mechanisms, it 88 was necessary to trace the ecological process for both nature and humans in hydrological 89 regimes. However, the lack of long-term monitoring data made it difficult to understand the 90 current state of lake ecology (Zhang et al., 2018). 91

92 Combined with <sup>210</sup>Pb and <sup>137</sup>Cs geochronology, stratigraphic records of sedimentary nutrients 93 loading, diatom assemblages, and TDI index of eutrophic diatoms, this paper aimed to 94 identify the eutrophication process synchronized with anthropogenic nutrient loading. The 95 work purposed to (1) establish a high-resolution record of nutrient level; (2) record the diatom 96 assembles shift since reservoir construction; (3) distinguish the stress of four distinct periods 97 on the diatom succession.

### 98 2. Materials and methods

## 99 2.1 Study area

As illustrated in Figure 1, Lake Xian'nv composed of Qian'yang (upstream) and Wu'long (downstream) and connected by Zhong'shan Gorge was built in 1958 with a surface area of 50 km<sup>2</sup>. It was a large subtropical mountain lake reservoir in south China and the second largest reservoir in Jiangxi Province. Originated from Mountain Wugongshan at the western

boundary of Jiangxi, River Yuan was a tributary of River Ganjiang, passed through Pingxiang 104 and Yichun cities, and finally entered River Ganjiang from lotus pavilion in Zhangshu City 105 106 after leaving Longweizhou, Xinxi Township, and Yushui District. The average annual runoff and sand from April to September were 110.94 m<sup>3</sup>/s and 450,000 tons respectively, accounting 107 for 68.5% and 84% of annual runoff and sand (Zhang et al., 2015). As an important gene pool 108 of subtropical plants, this basin had higher water quality, high species diversity, rich plant 109 species, high forest coverage rate, and good vegetation protection. However, local water 110 quality deterioration and eutrophication have occurred in the reservoir area, as well as 111 abnormal ecosystem security problems in recent years. 112

#### 113 2.2 Sediment sampling

Based on the collected data and investigation, 9 benthic diatoms sampling sites (S1-S9) and 1 114 deposition sampling site (C) were sampled in mid-March 2018. Among them, S1-S5 was 115 collected from the upstream of Lake Qian'yang, and S6-S9 was sampled from Lake Wulong 116 in the downstream. Meanwhile, a 37-cm long sediment core was collected from the northern 117 part of Lake Qian'yang (Figure 1) using a handheld cylindrical gravity corer with an outer 118 diameter of 90 mm (Nanjing Institute of Geography and Limnology, ZZC-90, China). The 119 sampling site of sedimentary columnar samples should meet the following requirements: (1) 120 near the lake center with little interference; (2) avoid disturbance to sediments. The core was 121 immediately sliced into 1cm sections on the shore and all samples were packaged in a 122 self-sealing plastic bag with a serial number predefined by the depth and sites of the samples. 123 124 After all samples were taken back to the laboratory, they were immediately stored in a refrigerator at -20 °C for further analysis. 125



Figure 1. Sampling sites for surface and core sediments from Lake Xian'nv, Jiangxi Province
 2.3 Sediment chronology

Pre-processing of samples: (1) 7mL sample tubes were marked according to the samples to be 129 tested, then these tubes were weighed; (2) samples after screening with 100 mesh sieve were 130 then loaded into the tube, and the total weight were weighed; (3) sample mass was equal to 131 the subtraction between total weight and tube weight. Determination of specific activities: the 132 specific activities of <sup>210</sup>Pb and <sup>137</sup>Cs in the sediments were measured using a Gamma 133 spectrometer (GWL-120-15, ORTEC Corporation, USA). The analytical test was entrusted to 134 the State Key Laboratory of Environment at the Nanjing Institute of Geography and 135 Limnology, Jiangsu province, China. The <sup>210</sup>Pb chronologies were calculated using the 136 constant rate of supply (CRS) model, and verified by the highest <sup>137</sup>Cs activity in 1963 (Lan et 137

139 The corresponding calculation formula was as follows:

140 
$$I_x = I \times e^{-\lambda t}$$

141 Where,  $I_x$  was the accumulation specific activities of <sup>210</sup>Pb above depth x (Bq·cm<sup>-2</sup>); I was the

total storage volume of <sup>210</sup>Pb (Bq·cm<sup>-2</sup>);  $\lambda$  was the decay constant (0.03114·a<sup>-1</sup>).

143 The deposition rate of each layer of sediment could be concluded based on the above 144 equation:

145 
$$t = -\frac{1}{\lambda} \ln \left( 1 - \frac{I_{x1}}{L} \right)$$

146 Where, t was the deposition time of each layer (a);  $I_{x1}$  was the accumulation specific activities 147 of <sup>210</sup>Pb above depth X1 (Bq·cm<sup>-2</sup>); I was the total storage volume of <sup>210</sup>Pb (Bq·cm<sup>-2</sup>);  $\lambda$  was 148 the decay constant (0.03114·a<sup>-1</sup>).

### 149 2.4 Chemistry analysis

TOC was determined by potassium dichromate oxidation-spectrophotometry (HJ 615-2011) 150 (Liao et al., 2017). Potassium dichromate standard solution and concentrated sulfuric acid 151 were added with 0.3g sieving sample, shaken evenly, put in an oil, and bathed at 185-190°C. 152 After that, an o-phenline indicator was added and then titrated with 0.2mol /L ferrous sulfate 153 solution. TN was determined by the Kay method (HJ 717-2014), in which 0.5g sieving soil 154 sample was dissolved with concentrated sulfuric acid, distillated by an automatic Kjeldahl 155 nitrogen analyzer (K9840, Hanon, China), and titrated with 0.02 mol/L hydrochloric acid 156 solution. determined by standard TP was alkali fusion Molybdenum-antimony 157

158	anti-spectrophotometry (HJ 632-2011), in which 0.5g sieving soil sample was dissolved with
159	concentrated sulfuric acid and perchloric acid, transferred to 100mL volumetric bottle, and
160	measured by UV-vis spectrophotometer (TU-1950, Persee, China) at 700nm wavelength.

161 *2.6 Diatom identification* 

Methods mentioned in Battarbee (ECRC) standard diatom treatment methods were used for 162 the pretreatment and preparation of diatom samples: (1) sampling: 0.5g of sediment sample 163 screened with 100 mesh sieve were weighed and put into a 50ml centrifuge test tube; (2): the 164 samples were added hydrochloric acid (HCL) and placed in a 90°C water bath until no 165 bubbles were reacted; (3) 30% hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>) was added until the organic matter 166 was removed by heating in the water bath; (4) cleaned with distilled water and centrifuged 167 (2500 RPM /min, 5min each) 3 times; (5) 100µL sample was soaked and dropped on clean 168 slide, covered with gum drop and finally dried into a permanent piece for microscope 169 observation; (6) diatoms were identified and counted using Olympus microscope (CX31) 170 under 100x10x oil lens. 171

- 172 *2.7 Data analysis*
- 173 2.7.1 Diatom index
- 174 (1) Relative abundance

At least 300 diatoms were counted in each sample, and the relative abundance P of diatom
species was calculated as follows:

177 
$$P = \frac{N_{sp}}{N} \times 100\%$$
 (2-1)

178 Where, N<sub>sp</sub> was the total number of diatom species, and N was the total number of diatom

179 species in each sample.

180 (2) Dominant species

181 Generally, diatom species account for more than 2% of all diatoms were defined as the 182 dominant species, and the calculation formula was as follows:

183 
$$P_{sp} = \frac{n_{sp}}{n} \times 100\%$$
 (2-2)

Where, N<sub>SP</sub> was the total number of diatom species in all samples; n was the total number of
diatoms in all samples.

186 (3) Diatom abundance

187 Abundance index D was used to describe the distribution of diatom diversity in the sediment,

188 and the calculation formula was as follows:

 $D = \frac{S-1}{\log_2 N} \tag{2-3}$ 

Where: N was the number of diatom species in the sample, and S was the number of diatomspecies in the sample.

192 2.7.2 Statistical methods

Multivariate statistical analysis methods including Hierarchical clustering, principal components analysis (PCA), and Redundancy analysis (RDA) were used to determine the diatom communities, and identify the main drivers of diatom community change. In this research, Excel 2013, SPSS 23.0, Canoco 5.0, Origion 8.5, and Photoshop CS6 were used to process the experimental data and draw graphics. The diatom index TDI was calculated by Omnidia7.5.

## **3. Results**

## *3.1 Physicochemical properties of sediments*





Based on the determination results of sedimentary column C, the vertical profile distribution 205 of specific activity of <sup>210</sup>Pbex and <sup>137</sup>Cs was shown in Figure 2(A). It could be confirmed from 206 the figure that the <sup>210</sup>Pb<sub>ex</sub> specific activity with a value ranging from 52.53 to 229.93 Bq·kg<sup>-1</sup> 207 presented an irregular zigzag distribution pattern with the increase of depth, indicating 208 significant human activities and the <sup>137</sup>Cs profile of activities ranged from 0 to 5.27 Bq·kg<sup>-1</sup>. 209 Labeled with <sup>137</sup>Cs, the chronological sequence of the corresponding sedimentary layer 210 identified by the CRS model is shown in Figure 2(B). It could be found from the figure that 211 <sup>137</sup>Cs began to appear at a depth of 5cm, and an obvious peak was recorded at a depth of 19cm 212 which was corresponding to 1963. Using the location of the <sup>137</sup>Cs accumulation peak to 213 calculate the sediment deposition rate, the measured age of the column sample was 155 years 214 in total, corresponding to 1863-2018, and the sediment deposition rate ranged from 0.03 to 215  $0.58 \text{ g}^{\circ}\text{cm}^{\circ}\text{yr}^{-1}$  with an average value of  $0.12 \text{ g}^{\circ}\text{cm}^{\circ}\text{yr}^{-1}$ . 216



Figure 3. Stratigraphic plots of geochemical and physical parameters of TP, TOC, TN, and C/N in the Xian'nv Lake

The average values of nutrient indexes including TP, TOC, and TN in surface sediments were 219 0.52, 25.43, and 2.76 mg  $g^{-1}$ . These indexes in sediment column C are illustrated in Figure 3. 220 Along the vertical section of the sedimentary column, the indexes showed similar trends and 221 gradually increased from bottom to top. Of these, the content of TP was varied from 0.31 to 222 0.88 mg·g<sup>-1</sup> with an average of 0.56 mg·g<sup>-1</sup>. The TOC content ranged from 13.37 to 33.65 223 mg<sup>·</sup>g<sup>-1</sup>, and the TN content ranged from 1.1 to 2.96 mg<sup>·</sup>g<sup>-1</sup>. The average values of TOC and 224 TN were 24.41 mg·g<sup>-1</sup> and 2.04 mg·g<sup>-1</sup> respectively. The C/N ratios fluctuated between 9.53 225 and 13.46 with an average value of 11.86. Combined with the clustering of diatoms and the 226 distribution characteristics of nutrient indexes, the nutrient indexes of the sediment column 227 from bottom to top were divided into four assemblage zones, including zone I (23-20 cm, 228 1955-1961), zone II (20-15 cm, 1961-1983), zone III (15-5 cm, 1983-2009) and surface zone 229 230 IV (5-0 cm, 2009-2018).

Generally, TP, TOC, and TN contents in Zone I were the lowest, and the average C/N ratio 231 was 10.32. After that, TP, TOC, and TN in Zone II with the depth among 20 to 15cm from 232 bottom to top showed a fluctuating upward trend, and maintained at a low level on a whole. 233 The average value of C/N in this period was 11.74, which was higher than that in combination 234 zone I. However, the contents of TOC, TN, and TP in Zone III with a depth from 15 to 5 cm 235 reached the maximum value in the sediment core, and the C/N ratio was also recorded at the 236 highest level with an average ratio of 12.37. Compared with Zone III, the level of nutrient 237 indexes in Zone IV decreased slightly and the C/N ratio of this belt was lower than that of the 238 down belt with an average value of 12.18. 239

240 *3.2 Diatom community characteristics in surface sediments* 

Class	Order	family	genus	Total
Class				diatom
	Coscinodiscaels	Coscinodiscaceae	Cyclotella Kutzing	4
Contribution			Melosira Agardh	2
Centrica			Stephanodiscus	1
e			Ehrenberg	
			Aulacoseira Thwaites	4
	Araphidiales	Fragilariaceae	Fragilaria Lyngbye	4
			Synedra Ehrenberg	3
	Raphidionales	Eunotiaceae	Eunotia Ehrenbery	4
	Biraphidinales	Naviculaceae	Gyrosigma Hassall	20
			Navicula Bory	5
			Pinnularia Ehrenberg	5
Pennatae		Cymbellaceae	Cymbella Agardh	4
		Gomphonemacea e	Gomphonema Ehrenberg	3
	Monoraphidales	Achnanthaceae	Cocconeis Ehrenberg	3
			Achnanthes Bory	3
	Aulonoraphidinale	Nitzschiaceae	Nitzschia Hassall	4
	S	Surirellaceae	Surirella Turpin	2

Table 1 Statistics of species and genera for benthic diatom in surface sediments

Based on the analysis of benthic diatom communities from surface sediments S1-S9, 68 242 species, 16 genera, 9 families, 5 orders, and 2 classes were identified, as listed in Table 1. 243 Diatoms were mainly composed of Centricae and Pennatae, of which Centricae has 11 244 species, 4 genera, 1 family, and 1 order, belonging to Coscinodiscaceae. There were 4 orders, 245 8 families, 12 genera, and 57 species of Pennatae, including Araphidiales, Raphidionales, 246 Monoraphidales and Aulonoraphidinales. According to the classification and statistics of 247 diatoms, the species of Raphidionales were the most, reaching 38 species, followed by 11 248 249 species of Coscinodiscaceae, 7 species of Araphidiales, 6 species of Aulonoraphidinales and 6 species of Monoraphidales. Among all families, Naviculaceae had the largest number of 250 species, reaching 29 species, followed by 11 species of Coscinodicaceae and 7 species of 251 Fragilariaceae, accounting for 42.6%, 16.2%, and 10.3% of the total respectively. Among all 252

genera, *Navicula Bory* had the largest number of species, accounting for 20 species, followed
by 5 species of *Pinnularia Ehrenberg* and 4 species of *Cyclotella Kutzing*, accounting for
29.4%, 7.4% and 5.9% of the total respectively.

In benthic diatoms, Achnathidium minutissimum was the absolute dominant species with the 256 highest relative abundance of 67.2% and the average relative abundance of 54.2%. 257 Aulacoseira granulata was another dominant species with the average relative abundance of 258 15.3%. Besides, there were dominant genera and species, including Navicula Bory with an 259 average relative abundance of 29.4%. The average relative abundances of the other common 260 261 generas, including Pinnularia Ehrenberg, Gomphonema Ehrenberg, Nitzschia Hassall, and Fragilaria Lyngbye, were less than 10%. The dominant species of benthic diatoms were 262 Achnathidium minutissimum, Aulacoseira granulata, Cyclostephanos dubius, Cyclotella 263 meneghiniana and Cyclotella radiosa. 264

265 *3.3 Diatom community characteristics in sediment core* 

Consistent with the species of diatoms in surface sediments, 129 species of diatoms belonging 266 to 22 genera, 10 families, 5 orders, and 2 classes were identified in the effective depth within 267 23 cm of columnar samples. Among them, there are 67 species of Raphidionales, followed by 268 25 species of Aulonoraphidales, 18 species of Coscinodiaceae, 10 species of Araphidiales, 269 and 9 species of Monoraphidales. Among them, Naviculaceae has the largest number of 270 species with 46 species, followed by 18 species of Coccinodiscaceae, 13 species of 271 Nitzschiaceae, and 11 species of Cymbellaceae, accounting for 35.7%, 14%, 10.1%, and 8.5% 272 of the total respectively. Among all genera, Navicula Bory has the largest number of species 273

with 30 species, followed by *Cyclotella Kutzing* with 9 species, and *Gomphonema Ehrenberg*with 8 species, accounting for 23.3%, 7%, and 6.2% respectively. The 9 dominant species and
genera were *Aulacoseira apigena*, *Aulacoseira granulata*, *Cyclotephanos dubius*, *Cyclotella meneghiniana*, *Cyclotella radiosa*, *Coconeis placendula*, *Fragilaria ulna*, *Achnathidium minutissimum*, accounting for 3.59%, 9.46%, 6.52%, 3.85%, 4.53%, 2.42% and 2.08% of
diatoms respectively.

As shown in Figure 4, there were 25 species with relative abundance  $\geq 1\%$  in the sample, 280 accounting for 96.32% of the total number of diatoms. Among them, there were 5 kinds of 281 planktonic diatoms. Below 20 cm of the core sediment, the proportion of Aulacoseira 282 granulata was the highest with its relative abundance ranging between 3.58% and 26.29% 283 and the average value of 14.94%, and it was recorded in 23 samples at the same time. Other 284 planktonic diatoms such as Cyclostephanos dubius, Cyclotella radiosa, Cyclotella 285 meneghiniana, Cyclotella radiosa, and Aulacoseira apigena were also recorded. The species 286 of epiphytic and benthic diatoms accounted for up to 70%, mainly including Achnanthes 287 minutissima, Cocconeis placentula, Stephanodiscus minutulus, Navicula minima, Navicula 288 cincta, Gomphonema gracile, Synedra ulna, etc. 289

The diatom assemblage data could be divided into four assemblage zones by constrained cluster analysis. The combination I belt was located between 23 - 20 cm with corresponding age ranging from 1955 to 1961, in which benthic diatoms were dominated with the highest abundance of 62% and the main dominant species were *Navicula cincta*, *Fragilaria ulna*, and *Cocconeis placentula*. Meanwhile, *Aulacoseira granulata* occupied an absolute advantage with the relative abundance fluctuating between 30% and 37%, showing a downward trend

from bottom to top. In combination II with a depth of 20-15cm, the planktonic diatom species 296 Aulacoseira granulata was still the dominant species with the relative abundance fluctuating 297 298 from 22% to 26%. Nutrient-tolerant species such as Cyclostephanos dubius, Cyclotella meneghiniana, Achnathidium minutissimum began to appear and showed an upward trend in 299 this belt. In combination III zone, the content of Cyclostephanos dubius, Cyclotella 300 meneghiniana, and Achnathidium minutissimum reached the maximum. In combination IV 301 zone, the content of planktonic species Cyclotella radiosa increased significantly. 302 Nutrient-tolerant species began to decrease, and the contents of benthic diatom species 303 Navicula oligotraphenta and Nitzschia solgensis increased. 304

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Figure 4. Stratigraphic plots of dominant diatom species, percentage of planktonic, epiphytic, and benthic diatoms in sediment core from Lake
 Xian'nv

#### 330 4. Discussion

<sup>137</sup>Cs was an artificial radionuclide generated by atmospheric nuclear tests in the 1950s and 331 1970s that flowed gradually into the water body with atmospheric subsidence and surface 332 runoff (Saniewski et al., 2020). As shown in Figure 2, a well-defined <sup>137</sup>Cs peak was observed 333 at 19 cm depth which could mark the 1963 when the testing of nuclear weapons was 334 concentrated in this period (Lan et al., 2020). Lake Xian'nv was formed for water storage 335 after a dam was built in 1958 and completed in 1960, by which the corresponding depth for 336 this period verified by <sup>210</sup>Pb dates was recorded about 22-23cm. Therefore, the sedimentary 337 samples within the effective depth of 23 cm from the surface were used for analysis. 338

As indicated in Figure 3, nutrients in sediments had been gradually accumulated since the 339 construction of the dam which usually led to worse of hydrodynamic conditions (Wu et al., 340 2019). After 1961, the values of TP, TOC, TN, and C/N ratios were slightly increased until 341 1983, and then a significant increase was recorded from 1983 to 2009. After that, the nutrient 342 contents were maintained at a stable level. The continuous increase of nutrients might indicate 343 that the nutrient sources in Lake Xian'nv Lake continued to increase. Meanwhile, the average 344 deposition rate based CRS model revealed that the intensity of all kinds of interference 345 continues to rise. It had been reported that TN and TP contents in Lake Xian'nv increased 346 continuously before 2009, and subsequently were gradually effectively controlled (Luo, 347 2017). 348

The input of nutrients played a key role in the evolution of the ecological environment of Lake Xian'nv (Zou, 2015). Since the 1950s, the development of industry, large-scale

urbanization, and the large-scale development of agriculture had led to a rapid increase in the 351 urban population and the use of chemical fertilizers around Lake Xian'nv (Liao et al., 2008). 352 353 The data shows that the nutrients and water quality of Lake Xian'nv was mainly determined by the following two factors (Zhang et al., 2015). The first source was the discharge of 354 355 industrial and domestic sewage which increased rapidly with the rapid economic development, and the water quality of the lake continued to deteriorate. On the other hand, cage culture had 356 been developed rapidly since the early 1980s. In 2006, the number of cages used for fish 357 farming was estimated at 15,000, which finally caused nutrients continuously accumulate in 358 359 sediments (Luo, 2017).

Besides, nutrient enrichment in the Lake Xian'ny recently was also consistent with the 360 continuous intensification of human activities, including urbanization and fertilizer usage, 361 which were also important sources of phosphorus and nitrogen for lakes (Chen et al., 2013). A 362 significant effect of numerous human activities on diatom succession, species composition, 363 abundance, and seasonal dynamics had been revealed (Kim et al., 2018). Due to the limitation 364 of data collection around the basin, this analysis was based on the population and fertilizer use 365 of Jiangxi Province, as shown in Figure 5. In recent decades, the population and fertilizer used 366 367 in Jiangxi Province had been continuously soared, especially after the 1980s. From 1950 to 1959, the population of Jiangxi Province was less than 20 million, and the population growth 368 rate was also maintained at a low level. However, the population was in a rapid growth stage 369 from 1959 to 1999, and maintained a steady slight growth trend after 2000. In 2017, the total 370 population of Jiangxi Province reached up to 46.22 million. At the same time, the use of 371 chemical fertilizer in Jiangxi Province has increased since the 1980s. 372



Figure 5 Trend of total population and fertilizer usage in Jiangxi Province since 1950

373

Lakes ecosystems were connected with the surrounding environment wherein materials and 375 energy could be exchanged frequently (Vlaičevića et al., 2021). These nutrients could be 376 absorbed by primary producers and consumers, and released after death (Sen and Mallick, 377 2021). The average C/N ratio in Lake Xian'nv at the early stage was 10.32, indicating that the 378 organic matter in the sediment in this period mainly came from both endogenous plankton and 379 terrestrial organic matter.Before 1983, the C/N ratios in Lake Xian'nv had been maintained at 380 12 with a relatively low level. However, a higher value of the C/N ratio at the third stage 381 indicated that the lake might be significantly affected by human activities. Generally, the 382 value of C/N ratio with 10-12 meant that organic matter derived from phytoplankton, while 383 the value of the C/N ratio over 20 meant that mixtures of algae and higher plants (Lei et al., 384 2018). 385

Moreover, studies had also shown that rapid population growth and climate warming had also effectively accelerated the eutrophication of Lake Xian'nv (Liao et al., 2008), Fortunately, these problems have attracted enough concern from the Chinese government and some non-government organizations, and a series of comprehensive water protection measurements had been conducted to reduce the environmental impact of various behaviors (Zou, 2015). Among them, one of these representative measures was to completely ban cage fishing in 2012 (Luo, 2017), which effectively reduced the nutrient content in Xian'nv Lake.



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Figure 6 Distribution of dominant diatom species in surface sediments and sediment core (A);
 TDI Box diagram of diatom for surface sediment and four diatom sedimentary zones (B)

Consistent with the eutrophication trend, the diatom community in Lake Xian'nv was mainlycomposed of benthic diatoms and showed a regular community shift. The abundance of

eutrophic species including Cyclostephanos dubius, Cyclotella meneghiniana, 399 and Achnathidium minutissimum increased gradually since 1961. These species were common in 400 some eutrophic lakes in Europe and North America due to higher nutrient level (Smol and 401 Stoermer, 2010; Yang et al., 2010; Lotrer et al., 1997). As illustrated in Figure 6(A), the main 402 dominant species with relative abundance greater than 5% included Achnathidium 403 minutissimum and Cyclostephanos dubius in the sediment core were also the dominant species 404 of surface sediment diatoms. The TDI calculation results in Figure 6(B) also indicated that the 405 TDI value of diatom community in surface sediment was significantly higher than that in 406 combination zone I, combination zone II, and combination zone IV, and slightly lower than 407 that in combination zone III, indicating that the water nutrient level in the lake area has been 408 controlled after years of improvement. The abundance and species composition of diatoms in 409 the sediment core showed obvious differences at different depths, which was consistent with 410 the change trend of TDI index. 411

Before the construction of the dam from 1955 to 1961, phytoplankton was dominated by 412 benthic diatoms and the quantity of diatoms was low with lower productivity. The TDI value 413 in the same period was maintained at a low nutritional level. After the completion of the 414 415 reservoir, the number of benthic diatoms suitable for flowing water decreased significantly, and the number of planktonic diatoms with easy growth in still water increased gradually. 416 Diatom species increased significantly, and diatom productivity increased greatly. During this 417 period, the TDI value increased, and nutrient-tolerant species appeared. Aulacoseira 418 granulata, an indicator for clean water, gradually decreased after the construction of the dam, 419 meaning a transfer of the ecosystem (Lei et al., 2021). Generally, Aulacoseira granulata was a 420

siliceous genus with a higher sedimentation rate and a preference for disturbed currents (Chen
et al., 2013; Schmieder, 2009; Hötzel et al., 1996). *Aulacoseira granulata* has been associated
with physical alterations, such as depth variation, turbulence, and mixing regime
(Costa-Böddeker et al., 2012), and tended to be more pronounced under increased flow
conditions in Jurumirim Reservoir (Nogueir et al., 2000).

At the initial stage of reservoir construction, nutrient indexes TP, TN, and TOC were 426 maintained ata low level, diatoms were dominated by benthic communities including 427 Cocconeis placentula, Gomphonema gracile and Fragilaria ulna, and planktonic diatom 428 Aurancoseira grannulata. The study on the eutrophication process of lakes in the middle and 429 lower reaches of the Yangtze River indicated that the content of Aulancoseira grannulata was 430 maintained ata higher ratio when the lake nutrients were low and dominated by benthic 431 diatoms (Chen et al., 2014). After the reservoir transferred from river to lake, the species and 432 biomass of planktonic diatoms increased, while the epiphytic and benthic diatoms couldn't 433 reproduce in large numbers in the deep-water environment (Wang, 2016). The contents of 434 Cyclostephanos dubius, Cyclotella meneghiniana, Aulacoseira alpigena and Achnathidium 435 minutissimum were gradually increased, and the TDI index value was also increased. A study 436 on diatoms in 45 eutrophication lakes of the Yangtze River revealed that Cyclotella 437 meneghiniana, Achnathidium minutissimum, and Aulancoseira grannulata were usually 438 indicators for water eutrophication (Dong et al., 2006). 439

In order to more clearly understand the structural change characteristics of sedimentary diatoms, principal component analysis (PCA) was performed on diatom data with relative abundance greater than 2%, as shown in Figure 7. The results showed that the first axis could explain 74.1% of the changes in the diatom community, and the relationship was significant.
The second axis only explained 6.9% of the changes in the diatom community, with little
impact. There were significant differences between the two ends of main axis 1, in which the
positive direction was mainly composed of benthic diatom species, such as *Navicula cincta*, *Stephanodiscus minutulus*, *Fragilaria ulna*, *Synedra ulna*, *Gomphonema gracile*, and the
negative direction was mainly composed of nutrient tolerant planktonic diatom species, such
as *Cyclostephanos dubius*, *Cyclotella radiosa*, *Aulacoseira granulata*, *Cyclotella*

450 *meneghiniana*. The change of diatom community could reflected in the transformation from 451 the positive direction of main axis 1 to the negative direction, which was consistent with the 452 succession from benthic diatom to planktonic diatom and the change of nutrient level.





Figure 7 Ordination plot of diatom assemblage with principal component analysis for Lake
 Xian'nv

456 Nutrient indicators accumulation in the sediment usually corresponded to the eutrophication 457 process of a lake (Zhang et al., 2018). The evolution process of nutrients in Lake Xian'nv was 458 reconstructed by using diatom community combination and nutritional indicators in this

research. The study revealed that the diatom communities were consistent with sedimentary 459 nutrition TDI index records. For more than 60 years after the operation of the reservoir, the 460 nutrient species including Aulancoseira grannulata, Cyclotella meneghiniana, Achnathidium 461 minutissimum, Cyclostephanos dubius, Aulacoseira alpigena had been competed with the 462 indicator species of clean water, and occupied a certain advantage between combination zone 463 II to IV and an absolute advantage in combination zone III. Due to the intensification of 464 human activities, such as cage fish culture, tourism, and domestic sewage discharge, the 465 nutrients in water had been increased. However, the water nutrition level had also been 466 decreased after the ban on cage fish culture in 2012. 467

#### 468 **5. Conclusion**

The diatom community succession after the formation of reservoir Lake Xian'nv was 469 reconstructed using multiple sediment biogeochemical proxies. Before the dam was built from 470 1955 to 1961, the phytoplankton were mainly benthic diatoms, and the number of diatoms 471 was small. After the completion of the reservoir, the number of benthic diatoms suitable for 472 flowing water decreased significantly, such as Aulacoseira granulata, and the number of 473 planktonic diatoms easy growth in still water increased gradually, diatom species increased 474 significantly.At the initial stage of reservoir construction, nutrient indexes TP, TN and TOC 475 maintain in a low level, diatoms were dominated by benthic communities including 476 Cocconeis placentula, Gomphonema gracile and Fragilaria ulna, and planktonic diatom 477 Aurancoseira grannulata. After the reservoir transferred from river to lake, the species and 478 479 biomass of planktonic diatoms increased, such as Cyclostephanos dubius, Cyclotella meneghiniana, Aulacoseira alpigena and Achnathidium minutissimum while the epiphytic and 480

benthic diatoms couldn't reproduce in large numbers in deep-water environment. To sum up,
the change of diatom community could reflected in the transformation from benthic diatom,
such as Navicula cincta Stephanodiscus minutulus Fragilaria ulna Synedra ulna
Gomphonema gracile, to nutrient tolerant planktonic diatom species, such as Cyclostephanos
dubius Cyclotella radiosa Aulacoseira granulata Cyclotella meneghiniana.

The acceleration of nutrient input had significantly increased the nutrient levels of the water 486 body and mainly contributed to the continuous increase of algal biomass and shift of the 487 diatom community structure. The hydrological dynamic conditions changed by the dam 488 construction, and accelerated human activities synchronously enhanced the ecosystem 489 changes. At the same time, the improvement of environmental protection management, 490 prohibition of some human activities, and the construction of sewage plants after 2008 have 491 effectively reduced the nutrient level of water. The supervision of industrial sewage discharge 492 should be strengthen to help reduce the discharge of industrial sewage by relevant 493 departments. In addition, it is essential to promote the concept of water conservation, which 494 can minimize the impact of human activities on the lake. 495

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