

1 **Effects of dimethoate, an organophosphate insecticide, on photosynthesis of five selected**
2 **phytoplankton species**

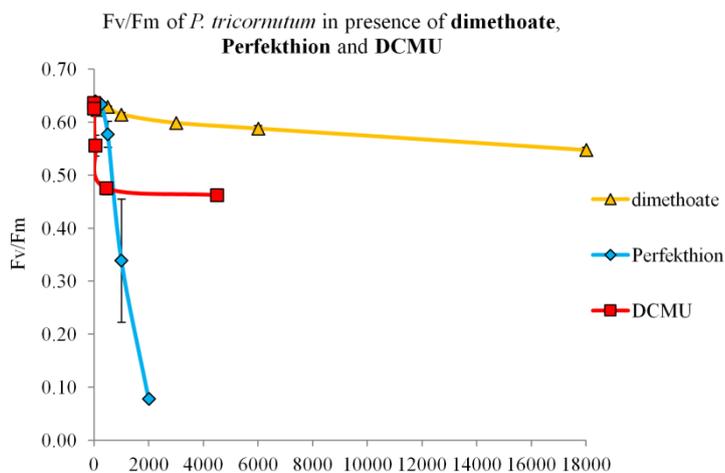
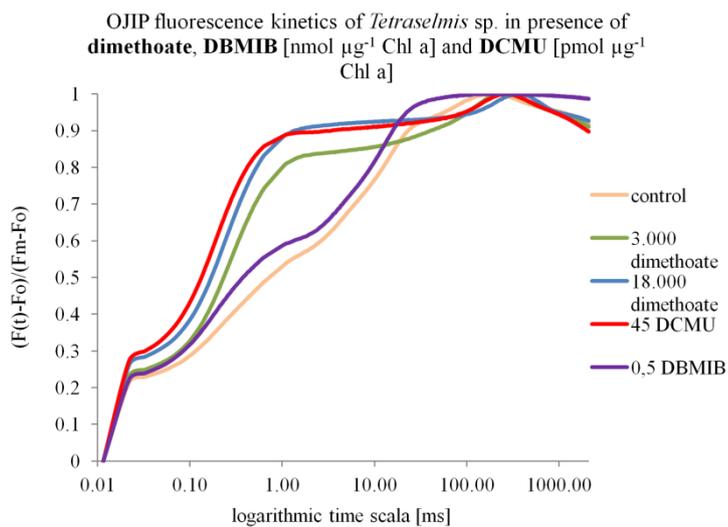
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9
10 **Graphical abstract**



concentrations: insecticides [nmol μg^{-1} Chl a] & DCMU [pmol μg^{-1} Chla]

12 **Abstract**

13

14 Several studies in the late eighties have shown that the insecticide dimethoate is not only toxic to
15 insects due to the inhibition of the acetylcholinesterase, but also to phototrophs by the inhibition of
16 photosynthesis. Although the use of dimethoate and also its commercial formulations are not any
17 longer permitted in the EU, it is still under application mostly in private hands or small farms in the
18 Mediterranean area and in legal use in many non-EU countries. The mode of action in the aquatic
19 environment is not clear until now. In the present study we have extended the aquatic toxicity test via
20 photosynthetic oxygen production by the analysis of Chl a - *in vivo* fluorescence and xanthophyll
21 cycle pigments, which indicate the capacity to resist against photoinhibition. The study presents a
22 data set including three different green algae and two diatoms to test the variety of responses from
23 the major algal taxa in the aquatic environment. The data show that dimethoate as well as the
24 commercial product Perfekthion inhibits PSII in a similar manner as DCMU (3-(3,4-Dichlorophenyl)-
25 1,1-dimethylurea), a PSII inhibitor that decreases the photosynthetic oxygen evolution rate. The
26 formulated product shows a higher toxicity than the pure chemical compound, which indicates that
27 the formulation cocktail increases stability/permeability as key elements controlling toxicity, leading
28 probably to photoinhibition.

29 **Keywords:** Dimethoate, insecticide, Perfekthion, phytoplankton, photosynthesis, oxygen rate,
30 fluorescence, photochemical efficiency, non-photochemical quenching, xanthophylls

31

32 **1. Introduction**

33 Phytoplankton play a key role in aquatic ecosystems as primary producers of the aquatic food chain.
34 The planktonic algae enhance the rapid nutrient recycling as well as the removal of soluble organic
35 substances. Although the phytoplankton biomass in the world's oceans amounts to not more than 1-

36 2% of the total global plant carbon, it can fix up to almost 50% of the photosynthetic carbon every
37 year (Falkowski et al., 1998; Field et al., 1998).

38 In recent years, new challenges in environmental management have arisen with a variety of
39 environmental pollutants establishing their presence (Musmarra et al., 2019). Among these, pesticides
40 are increasingly used in the modern society, posing serious environmental hazards. They may enter
41 water ecosystems through spray, drift, leaching, runoff, or accidental spills (Van der Werf, 1996),
42 affecting non-target organisms such as phytoplankton (Kungolos et al., 2009). The use of
43 organophosphate (OP) insecticides increased drastically in the past decades because of their low
44 persistence in the environment (Epa, 2006; Petsas et al., 2007; Vagi et al., 2005) and high efficiency
45 against pests. Dimethoate is an important OP insecticide with foliar application used to kill insects
46 by contact and stomach action (Tomlin, 1997). It is an acetyl cholinesterase inhibitor that has been
47 used in the Greek agriculture primary against the fruit fly (*Bactrocera oleae*, *Dacus oleae*), which
48 lays its eggs into the olive fruit. However, besides its efficiency in killing insects, several studies have
49 shown that photosynthesis is the main target of dimethoate in phytoplankton species (e.g. Wong &
50 Chang, 1988; Kobbia et al., 1991; Perona et al., 1991a; Mohapatra & Schiewer, 1998; Chen et al.,
51 2007; Moermond et al., 2008). The same is also true for Rogor, which has been the commercial
52 product of dimethoate used in agriculture and which after some years was renamed to Perfekthion,
53 containing 37% dimethoate as active ingredient (a.i.) (Adhikary, 1989; Panda, 1998; Surekha, 1999;
54 Mohapatra & Schiewer, 2000; Jena et al., 2012;).

55 According to guidelines from the Federal Office of Consumer Protection and Food Safety of Germany
56 (BVL, 2019) and the European Food and Safety Authority (EFSA, 2022), the EU commission
57 prohibited the use of plant protection products containing dimethoate after the 30th of June 2020 (EU,
58 2019 a,b). However, these products are still in use in many other non-EU countries. Furthermore, the
59 use of non-registered plant protection products is still significant in many rural parts of Greece.

60 Little is known about the precise target of dimethoate within the photosynthetic apparatus of
61 phytoplankton cells. Different effects have been described, such as inhibition of the electron transport
62 between photosystem I (PSI) and photosystem II (PSII) of *Nostoc* cells (Chen et al., 2007), increase
63 of PSII fluorescence (Mohapatra et al., 1997), and inhibition of PSII activity in *Synechocystis* sp.
64 (Mohapatra et al., 1996). For Rogor it has been reported that it inhibits the PSII-PSI electron-flow in
65 *Chlorella vulgaris* (Jena et al., 2012) and in *Synechocystis* sp. (Mohapatra, 2000).

66 The measurement of photosynthetic oxygen production is routinely performed by using a Clark
67 electrode. While this measurement allows quantification of the total efficiency of the photosynthetic
68 light reaction, fluorescence-based measurements as OJIP transients and NPQ measurements enable
69 mechanistic insights in single steps of the photosynthetic electron transport chain (ETC). The OJIP
70 fluorescence reflects the time course of energy transfer from PSII to PSI. O, J, I and P fluorescence
71 levels are induced by different redox states of the ETC (Stirbet & Govindjee, 2011). At O-level, the
72 PSII reaction centers are open, i.e. Q_A is oxidized, hence fluorescence is minimal (F_o). At P-level, all
73 PSII reaction centers are in a closed state and fluorescence is maximal (F_m). The J- and I-level are
74 intermediate levels, whereas the J-level is a reflection of the exchange of a reduced plastoquinone-
75 molecule (Pq) for an oxidized Pq at the Q_B -site, while the I-level reflects the rate limitation imposed
76 by the re-oxidation of plastoquinol molecules at the cytochrome (cyt) b_6f -complex (Schansker et al.,
77 2005). The variable fluorescence is the difference between F_m and F_o . The maximum photochemical
78 efficiency F_v/F_m describes the maximum efficiency of a photon trapped in the antenna to induce
79 photochemistry at PSII reaction centers (Schreiber et al., 1994).

80 Plants and algae employ a mechanism to protect themselves from the adverse effects of high light
81 (HL) intensity, where light energy absorption exceeds the capacity for light utilization in
82 photosynthesis, by dissipating the excessive energy as heat (NPQ) (Müller et al., 2001; Horton et al.,
83 2005; Baker, 2008). HL conditions increase the transthylakoidal proton gradient (low pH in the
84 lumen), which activates the de-epoxidizing enzymes of the xanthophyll cycles. By interacting with

85 antenna proteins, the newly formed pigments of the xanthophyll cycles induce a conversion of
86 excessively absorbed light energy into heat. Two major photoprotective xanthophyll cycles are
87 known: the violaxanthin cycle (Vx cycle), primarily of the green algae (Chlorophyceae) and land
88 plants (Hager, 1980), and the diadinoxanthin (Dd) cycle of the diatoms (Bacillariophyceae). Under
89 HL conditions violaxanthin (Vx) is converted via antheraxanthin (Ax) to zeaxanthin (Zx) in the Vx
90 cycle (Sapozhnikov et al., 1957; Yamamoto et al., 1962) and diadinoxanthin (Dd) to diatoxanthin
91 (Dt) in the Dd cycle (Hager & Stransky, 1970). The de-epoxidation state (DES) of each cycle can be
92 determined by measuring the respective xanthophyll pigment concentrations by HPLC.

93 NPQ can be composed of the high-energy state (qE), the state transition (qT) and the photoinhibitory
94 (qI) quenchings (Müller et al., 2001; Goss & Lepetit, 2015). qI is either caused by the inactivation or
95 damage of PSII reaction centers or by stable quenching in the PSII antenna. In diatoms, qT does not
96 exist and its NPQ mainly relies on qE, a quenching mechanism, which is controlled by the build-up
97 of a transthylakoidal proton gradient (ΔpH), the xanthophyll cycle and the presence of specific
98 polypeptides of the light-harvesting complex (LHC) antenna, named LhcX (Lavaud, 2007; Bailleul et
99 al., 2010; Buck et al., 2019). The requirements of NPQ in green algae (violaxanthin cycle, proton
100 gradient and LHCII aggregation) are in principle comparable to those of higher plants (Goss &
101 Lepetit, 2015), but instead of PsbS, green algae rely on LhcSR proteins (related to LhcX) (Bassi
102 Dall'Osto, 2021) and the importance of the xanthophyll cycle for qE is species dependent (Quaas et
103 al., 2015).

104 The aim of the study was to investigate the impact of dimethoate on the photosynthesis of
105 phytoplankton species from different taxonomical classes. Furthermore, this research may improve
106 our understanding of the connection between the ETC, the xanthophyll pigment de-epoxidation and
107 NPQ and enrich our knowledge on the ecotoxicology of OPs.

108 **2. Materials and methods**

109 *2.1 Reagents*

110 All reagents were of analytical grade and purchased from Sigma-Aldrich Chemie GmbH (Steinheim,
111 Germany), unless otherwise stated. Dimethoate (2-dimethoxyphosphinothioylsulfanyl-*N*-
112 methylacetamide, CAS No. 60-51-5) with a purity $\leq 100\%$ was used in all experiments by Clark
113 electrode, PAM and HPLC, while the PSII inhibitor DCMU (3-(3,4-Dichlorophenyl)-1,1-
114 dimethylurea, CAS-No. 330-54-1) was used as a positive control. The effect of 3 μmol dimethoate
115 μg^{-1} Chl *a* and 45 pmol DCMU μg^{-1} Chl *a* on photosynthesis of the tested phytoplankton species was
116 compared in all experiments, as a similar effect of both pesticides was observed in the OJIP
117 fluorescence kinetics of the tested species (except in *Tetraselmis* sp.). DBMIB (2,5-Dibromo-6-
118 isopropyl-3-methyl-1,4-benzoquinone, CAS-No. 29096-93-3), an inhibitor of cyt *b*₆*f* complex, was
119 used as a positive control in the OJIP fluorescence experiments. Perfekthion (BASF SE company,
120 Ludwigshafen, Germany), the commercial formulation of dimethoate, is a blue emulsifiable liquid
121 that contains 37,2% (w/w) dimethoate as active ingredient (a.i.), 43.5-48% cyclohexanone (CAS-No.
122 108-94-1), 4.2-5.2% solvent naphtha (CAS-No. 64742-94-5) and 4.2-5.2% acetic anhydride (CAS-
123 No. 108-24-7). All reagent concentrations are referred to final concentrations in the phytoplankton
124 samples and are dissolved in ethanol. The control of each reagent was 100% ethanol.

125 2.2 Phytoplankton species and culture media

126 The three Chlorophyceae species *Chlamydomonas reinhardtii*, *Dunaliella tertiolecta* and
127 *Tetraselmis* sp., and the two Bacillariophyceae species *Phaeodactylum tricornutum* and
128 *Thalassiosira pseudonana* were grown in 250 ml flasks (20°C), in a day/night rhythm of 16/8 h, with
129 a white light intensity of 15-45 $\mu\text{E m}^{-2} \text{s}^{-1}$ (LL), on shakers (120-130 rpm). The cells were counted
130 per Coulter Counter instrument (Multisizer 3, Beckman, Indianapolis, USA), as only phytoplankton
131 cells in exponential growing phase were used in the experiments.

132 The type of the culture media and the origin of the phytoplankton species are described in Table 1.
133 The pH of all used culture media was adjusted with hydrochloric acid to 7.0. In *f*₂ medium the organic
134 compound Tris (pH 8) was added as a buffer.

135 $f2_{Si/Se}$ medium was prepared for *T. pseudonana* using $f2$ medium according to Guillard and Ryther
 136 (1962) and Guillard (1975). 4.5 ml $f2$ nutrient solution, 1 ml of a 0.82 M Na_2SiO_3 and 11.6 μ M
 137 H_2SeO_3 solution were added to 1 L artificial seawater (ASW) via a sterilised filter syringe.
 138 The recipe for *C. reinhardtii* culture medium was obtained from Pringsheim and Koch (1964).

139 **Table 1:** Phytoplankton species used in the study with their respective origin, culture medium, salinity and pH.

species	origin	culture medium	salinity	source
<i>Tetraselmis</i> sp.	SAG (Culture Collection of Algae, Goettingen University, Germany)	$f2$	35‰	Guillard & Ryther (1962) and Guillard (1975)
<i>D. tertiolecta</i>	SAG 183.80	$f2$	35‰	Guillard & Ryther (1962) and Guillard (1975)
<i>C. reinhardtii</i>	SAG 11-32a	<i>Chlamydomonas</i> medium	0‰ (bidistilled Water)	Pringsheim & Koch (1964)
<i>P. tricornutum</i>	CCAP 1052/6 (UTEX 646, SAG 1090-6), Finland	$f2$	35‰	Guillard & Ryther (1962) and Guillard (1975)
<i>T. pseudonana</i>	CCMP 1335 Bigelow Laboratory for Ocean Sciences, West Boothbay Harbor	$f2_{Si/Se}$	35‰	Guillard & Ryther (1962) and Guillard (1975)

140

141 2.3 Cell lysis and chlorophyll a determination

142 Chlorophyll a (Chl a) of the algal culture was determined in order to use comparable pesticide
 143 concentrations (μ g⁻¹ Chl a) in the experiments. For this purpose, 3 ml of each phytoplankton species
 144 was centrifuged for 4 min at 4000 x g and 20°C (Beckman Coulter, Allegra 25R Centrifuge, Krefeld,
 145 Germany). The supernatant was discarded, methanol (100 μ l) and glass beads were placed on the cell
 146 pellet and vortexed (Vortex-Genie 2, Scientific Industries, New York, USA). Acetone (900 μ l) was
 147 added and the mixture was vortexed, then centrifuged for 1 min at 4000 x g and 20°C. The absorbance
 148 of the supernatant (2 replicates) was measured (Ultrospec 2100 pro, UV/Visible Spectrophotometer,
 149 Biochrom, Cambridge, England). The Chl a concentration of the tested Bacillariophyceae and

150 Chlorophyceae species was calculated according to the formula of Jeffrey und Humphrey (1975), as
151 adapted by using the measurements of Ritchie (2006) (formulas 1 - 2).

152 Bacillariophyceae:

$$153 \text{ Chl a} = 11.49 \times (E_{664} - E_{750}) - 0.45 \times (E_{630} - E_{750}) \text{ } [\mu\text{g mL}^{-1}] \quad (1)$$

154 Chlorophyceae:

$$155 \text{ Chl a} = 11.87 \times (E_{664} - E_{750}) - 1.79 \times (E_{647} - E_{750}) \text{ } [\mu\text{g mL}^{-1}] \quad (2)$$

156 Chl a: chlorophyll a; E_x : absorbance at x nm wavelength

157 *2.4 Oxygen evolution rate measurements by Clark electrode*

158 The photosynthetic oxygen evolution rate and the respiration of the tested phytoplankton species was
159 tested with the Clark electrode in presence of 0.125-18 $\mu\text{mol } \mu\text{g}^{-1}$ Chl a dimethoate and 12.5-2000
160 $\text{nmol } \mu\text{g}^{-1}$ Chl a Perfekthion. The concentration of the phytoplankton species was adjusted to 2 μg
161 Chl a mL^{-1} and after acclimatisation (1 h) to low light (LL), 2 ml of the sample was placed into the
162 reaction chamber of a Clark electrode (Hansatech Instruments, Oxy-Lab, Helmut Saur, Laborbedarf,
163 Reutlingen, Germany). Algal respiration was measured as negative oxygen evolution rate under dark
164 conditions (4 minutes). Immediately afterwards, under light condition ($150 \mu\text{E m}^{-2} \text{ s}^{-1}$ light intensity),
165 the net photosynthetic oxygen evolution rate of the algae was calculated, first in absence (4-6 minutes
166 illumination) and then in presence of dimethoate/Perfekthion (for another 4-6 minutes). 4.5-4500
167 $\mu\text{mol DCMU } \mu\text{g}^{-1}$ Chl a was used as a positive control. Under the following dark condition (4
168 minutes) the algal respiration was measured in presence of the insecticide.

169 *2.5 Fluorescence measurements by PAM instrument*

170 The fluorescence measurements were performed by an Aqua Pen instrument (Aqua Pen-C, AP-C 100,
171 Photon Systems Instruments, Drasov, Czech Republic). Each phytoplankton sample was adjusted to
172 $1.0 \mu\text{g Chl a mL}^{-1}$ and acclimatised (1 h) to LL. The insecticide was added to 1 ml of each algae and
173 dark acclimated (10 minutes). Afterwards, it was pipetted into a cuvette and OJIP fluorescence and
174 NPQ were measured under blue excitation light (450 nm) .

175 The final concentrations of dimethoate, Perfekthion, DCMU, and DBMIB of the samples are listed
 176 in Table 2, according to the tested phytoplankton species and the experiments.

177 **Table 2:** Final dimethoate, Perfekthion, DCMU, and DBMIB concentrations of the fluorescence experiments.

measurements	phytoplankton species	concentrations [nmol μg^{-1} Chl a]			
		dimethoate	Perfekthion	DCMU	DBMIB
OJIP	3 Chlorophyceae (i.e. <i>Chlamydomonas reinhardtii</i> , <i>Dunaliella tertiolecta</i> and <i>Tetraselmis</i> sp.) & 2 Bacillariophyceae (i.e. <i>Phaeodactylum tricornutum</i> and <i>Thalassiosira pseudonana</i>)	125-18000	12.5-2000	0.045	-
	<i>Tetraselmis</i> sp.				0.5
F_v/F_m	3 Chlorophyceae & 2 Bacillariophyceae	125-18000	12.5-2000	0.0045-4.5	-
NPQ _L	3 Chlorophyceae & 2 Bacillariophyceae	125-18000	25-2000	0.045	-
NPQ _{DR}	3 Chlorophyceae & 2 Bacillariophyceae	3000	500	0.045	-

178
 179 Specifically, for OJIP fluorescence measurements, the intensity of the measuring light was 3.3 nE m⁻²
 180 s⁻¹ and of the saturating flash 2100 $\mu\text{E m}^{-2} \text{s}^{-1}$.

181 For NPQ measurements, dark-acclimated algae were exposed to actinic irradiance (700 $\mu\text{E m}^{-2} \text{s}^{-1}$,
 182 200 s) to elicit a transient Kautsky effect. Moreover, a sequence of saturating flashes was applied on
 183 top of the actinic light to probe NPQ in the light adapted state (NPQ_L) (for details see Table 3). NPQ_L
 184 was automatically calculated using the Stern-Volmer parameter (formula 3):

185
$$\text{NPQ}_L = F_m/F_{m'} - 1 \quad (3)$$

186 F_m : maximum fluorescence in the dark; F_m : maximum fluorescence in the light
 187 After exposure to continuous illumination, the relaxation of NPQ was determined by means of
 188 saturating light pulses applied in the dark (390 s) (Table 3). NPQ dark recovery (NPQ_{DR}) of
 189 dimethoate/Perfekthion- and DCMU-treated phytoplankton samples (Table 2) was automatically
 190 calculated and compared to the last measured NPQ_L of the control in order to examine the maximum
 191 recovery condition of each phytoplankton species.

192 **Table 3:** NPQ protocol with duration and saturated light pulse details for the NPQ measurements with the
 193 Aqua Pen instrument (PSI, online).

	Phase	Duration	Nr. of pulses	1st pulse	Pulse interval
NPQ	Light (NPQ _L)	200 s	10	10 s	20 s
	Dark Recovery (NPQ _{DR})	390 s	7	20 s	60 s

194 2.6 Xanthophyll pigment measurements by HPLC

195 Specific xanthophyll pigments of the phytoplankton samples were quantified by reversed phase
 196 HPLC (VWR-Hitachi LaChrom Elite) to determine the DES of the xanthophyll cycle. In order to
 197 induce the de-epoxidation, the samples in the Clark electrode were illuminated with 500 $\mu\text{E m}^{-2} \text{s}^{-1}$
 198 light intensity. 0.125-3 $\mu\text{mol dimethoate } \mu\text{g}^{-1}$ Chl a, 0.5 $\mu\text{mol Perfekthion } \mu\text{g}^{-1}$ Chl a and 45-4500
 199 pmol DCMU μg^{-1} Chl a were added to each sample at the beginning of the measurements under dark
 200 conditions (4 minutes). After 10 minutes of illumination, samples were filtered through 1.2 μm
 201 Membrane Isopore Polycarbonate filters (Millipore, USA), the filters flash-frozen in liquid nitrogen
 202 and stored at -80°C until further HPLC analysis.

203 Pigment extraction and HPLC separation were performed according to the protocol of Jakob et al.
 204 (1998). The pigments of the filtered cells were extracted with a mixture (700 μl) of 81% methanol /
 205 9% 0.2 M ammonium acetate / 10% ethyl acetate and after adding a minor amount of glass beads,
 206 vortexed and centrifuged (Eppendorf Centrifuge 5415 D, Hamburg, Germany) at 13200 rpm for 2.5
 207 min. 400 μl of the supernatant from each sample tube was injected onto a calibrated HPLC system

208 equipped with a 10°C-cooled autosampler (L-2200), a photodiode array (model L-2455) and a
209 Nucleosil 120-5 C18 column (Macherey-Nagel, Düren, Germany). Pigments were separated using a
210 linear gradient system consisting of eluent A (90% methanol / 10% 0.5 M ammonium acetate, vol/vol)
211 and eluent B (90% methanol / 10% ethyl acetate, vol/vol). The flow rate over the column was 1.0 ml
212 min⁻¹.

213 The individual peaks of the HPLC diagram were identified based on the respective absorption spectra
214 and retention times. Pigments were calculated in pmol using specific calibration factors and the DES
215 ($D_t/(D_t+D_d)$) of the Dd cycle was calculated..

216 The DES of A_x/V_x and not of $(A_x+Z_x)/(V_x+A_x+Z_x)$ was measured under HL for the green algae,
217 as Z_x could not be distinguished from lutein in the absorption spectrum of the chromatogram. A_x/V_x
218 was calculated according to the area of each pigment in the chromatogram, without using a conversion
219 factor. The DES of $(A_x+Z_x)/(V_x+A_x+Z_x)$ was indirectly calculated as $(V_x \text{ dark control} - V_x$
220 $\text{light})/V_x \text{ dark control} = 1 - (V_x \text{ light})/(V_x \text{ dark control})$, as the control value of V_x in the dark roughly
221 corresponds to the total quantity of $V_x+A_x+Z_x$ in the light, assuming there is no *de novo* V_x synthesis
222 during the 10 min of light stress exposure and most A_x and Z_x are epoxidised to V_x in dark
223 acclimation conditions. $V_x\text{-light}$ is the control value of V_x or the V_x -value after insecticide treatment,
224 both after 10 min of light exposure to $500 \mu\text{E m}^{-2} \text{ s}^{-1}$.

225 **3. Results and discussion**

226 *3.1 Oxygen evolution measurements*

227 Absolute net oxygen evolution rates for the tested species (Table S1) are in a comparable range as
228 usually reported for these species (Geel et al., 1997; Ruffle et al., 2001; Bailleul et al., 2010; Chen et
229 al., 2021). Net oxygen evolution rate inhibition of the tested algae in presence of $3 \mu\text{mol dimethoate}$
230 $\mu\text{g}^{-1} \text{ Chl a}$ is about 50-60%, except for *C. reinhardtii* with 25% (Table 4 and Table S1). Apparently
231 the last species is less sensitive to dimethoate than the other tested species. $45 \text{ pmol DCMU } \mu\text{g}^{-1} \text{ Chl}$
232 a inhibits the oxygen evolution rate of the selected species similar or stronger compared to the

233 mentioned dimethoate concentration, although the DCMU concentration was several orders of
 234 magnitude lower than dimethoate (Table 4).

235 The lowest observed effect concentration (LOEC) of dimethoate is 0.125 $\mu\text{mol } \mu\text{g}^{-1}$ Chl a for *D.*
 236 *tertiolecta* and *Tetraselmis* sp., 0.250 $\mu\text{mol } \mu\text{g}^{-1}$ Chl a for *T. pseudonana*, 0.5 $\mu\text{mol } \mu\text{g}^{-1}$ Chl a for *P.*
 237 *tricornutum* and 2 $\mu\text{mol } \mu\text{g}^{-1}$ Chl a for *C. reinhardtii*. Mohapatra et al. (1997) reported a significant
 238 effect of dimethoate on net photosynthetic oxygen production and photosynthetic carbon fixation of
 239 *Synechocystis* sp. PCC 6803 at concentrations higher than 50 μM .

240 **Table 4:** Net oxygen evolution rate inhibition of the tested phytoplankton species under 150 $\mu\text{E m}^{-2} \text{ s}^{-1}$ light
 241 intensity, in presence of dimethoate and DCMU (positive control). 3 technical replicates (t.r.) for each
 242 dimethoate-treated sample , 1-3 t.r. for each DCMU-treated sample.

Net O ₂ evolution rate inhibition [%]	<i>C.</i> <i>reinhardtii</i>	<i>D.</i> <i>tertiolecta</i>	<i>Tetraselmis</i> sp.	<i>P. tricornutum</i>	<i>T. pseudonana</i>
3 μmol dimethoate μg^{-1} Chl a	25 ± 6	53 ± 2	52 ± 1	58 ± 4	52 ± 3
45 pmol μg^{-1} Chl a DCMU	49 ± 10 (interpolated value)	-	79 ± 4	48 ± 5	71

243 Dimethoate does not remarkably affect the respiration of the tested phytoplankton species, except of
 244 *D. tertiolecta* and *Tetraselmis* sp., decreasing the respiration at concentrations ≥ 6 μmol dimethoate
 245 μg^{-1} Chl a (Table 5). In similar studies it has been reported that low concentrations of dimethoate
 246 increased the respiration of the tested phytoplankton species. Piska and Waghary (1991) observed
 247 that about 50 μM dimethoate increased the respiration of the primary producers of a lake ecosystem
 248 and decreased the net photosynthetic production. Respiration was also found to be increased in
 249 *Anabaena* (Perona et al., 1991b), *Nostoc* (Chen et al., 2007) and *Synechocystis* (Mohapatra et al.,
 250 1997) cells after adding dimethoate concentrations higher than 100 μM .

251 **Table 5:** Respiration rate of *D. tertiolecta* and *Tetraselmis* sp. measured in the dark, in presence of increasing
 252 dimethoate concentrations.

Respiration rate [$\mu\text{mol mg}^{-1} \text{Chl a h}^{-1}$]	control	3 μmol dimethoate $\mu\text{g}^{-1} \text{Chl a}$ (3. t.r.)	6 μmol dimethoate $\mu\text{g}^{-1} \text{Chl a}$ (1 t.r.)	18 μmol dimethoate $\mu\text{g}^{-1} \text{Chl a}$ (1 t.r.)
<i>D. tertiolecta</i>	53 \pm 2 (3 t.r.)	63 \pm 3	39	30
<i>Tetraselmis sp.</i>	104 \pm 15 (2 t.r.)	96 \pm 12	57	39

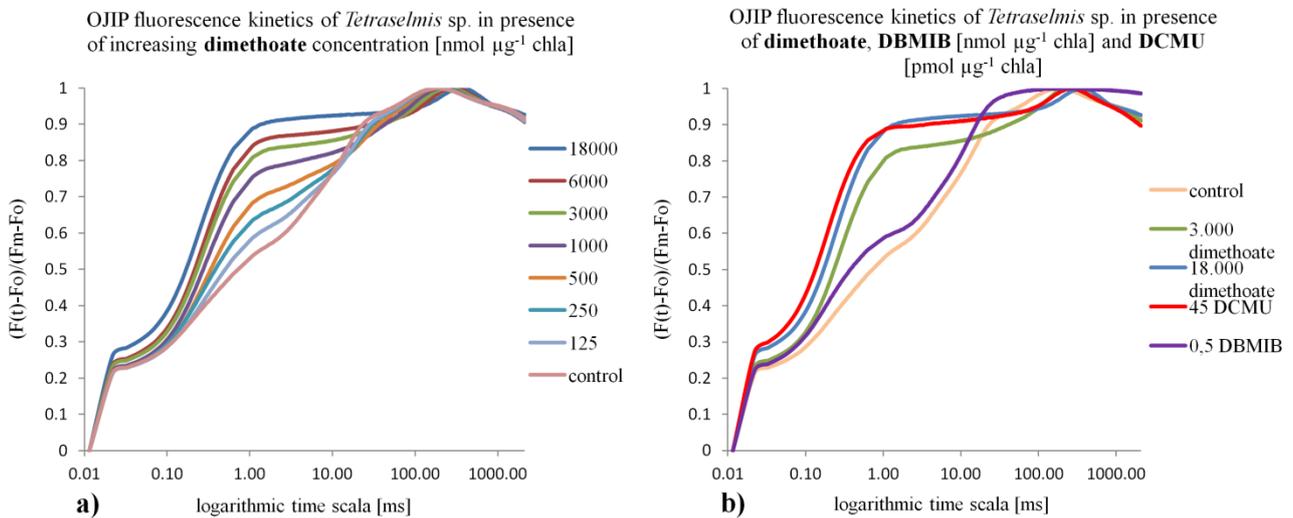
253 3.2 Chlorophyll fluorescence measurements

254 3.2.1 OJIP measurements

255 OJIP fluorescence measurement of *Tetraselmis sp.* (Fig. 1a) depicts a representative trace of all OJIP
 256 measurements of the selected phytoplankton species in presence of dimethoate. The results show an
 257 increasing PSII inhibition with increasing dimethoate concentration (Fig. 1a). However, even high
 258 concentrations like 18 $\mu\text{mol dimethoate } \mu\text{g}^{-1} \text{Chl a}$ do not inhibit PSII completely.

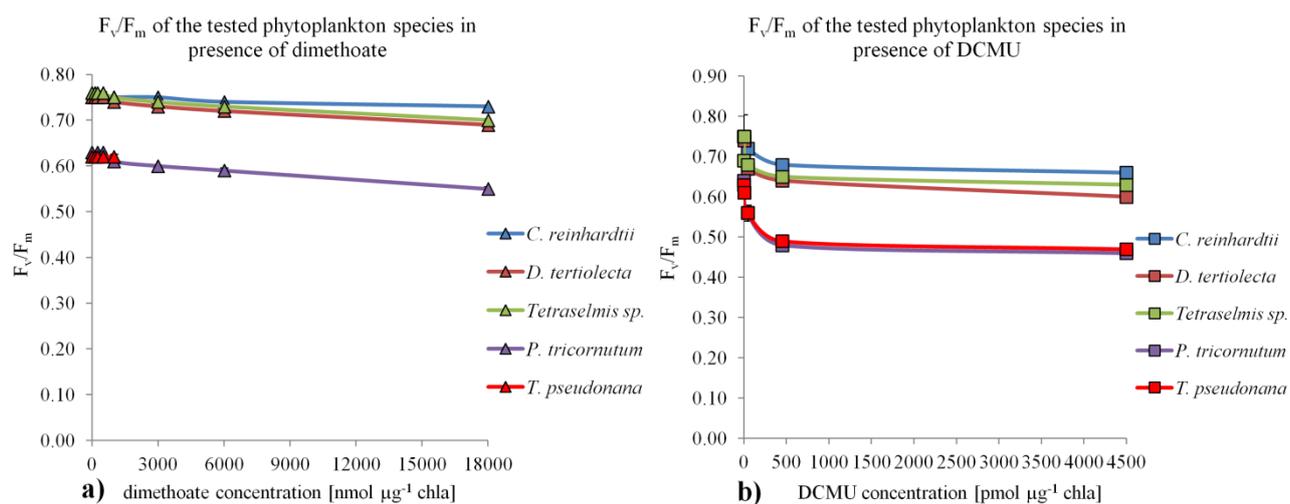
259 Dimethoate is acting like DCMU in all five tested phytoplankton species, disrupting the electron
 260 transport at the side of PSII beyond the primary acceptor (Q_A), as the O-J rise is related to the
 261 reduction of Q_A in PSII (Boisvert et al., 2006). These findings are in agreement with Sridevi et al.
 262 (2012), who examined the effect of dimethoate as a.i. of Rogor on photosynthetic pigment
 263 fluorescence of *Chlorella vulgaris*. In our study, 45 $\mu\text{mol DCMU } \mu\text{g}^{-1} \text{Chl a}$ shows a similar PSII
 264 inhibition like 18 $\mu\text{mol dimethoate } \mu\text{g}^{-1} \text{Chl a}$ in *Tetraselmis sp.* (Fig. 1b), and like 3 $\mu\text{mol dimethoate}$
 265 $\mu\text{g}^{-1} \text{Chl a}$ in the other four tested phytoplankton species. Clearly, dimethoate does not act like
 266 DBMIB (Fig. 1b), which binds at the $\text{cyt } b_6f$ complex at low concentrations (0.5 $\text{nmol DBMIB } \mu\text{g}^{-1}$
 267 Chl a), which is manifested by a maximum of fluorescence at the I-level and with no distinguishable
 268 I-P transition (Lepetit & Dietzel, 2015). Mohapatra et al. (1996) demonstrated that dimethoate affects
 269 the PSII activity and phosphorylation of *Synechocystis* at all tested concentrations (10-3000 μM). In
 270 contrast to the studies showing that dimethoate inhibits PSII, Chen et al. (2007) reported that
 271 dimethoate inhibits the ETC between PSII and PSI or even the dark reaction of the Cyanophyceae
 272 species *Nostoc*. According to this author, dimethoate (2 mM) significantly increases the electron

273 transport activity from water to methylviologen by 193% compared to the control and the PSI activity
 274 by more than 400% compared to non-treated cultures. Our data indicate a binding of dimethoate at
 275 the PSII core, partially preventing the electron transfer from Q_A towards the cyt b_6f complex.



276 **Fig. 1:** OJIP fluorescence kinetics of *Tetraselmis* sp. after double normalisation in presence of **a)** increasing
 277 dimethoate concentrations [$\mu\text{mol } \mu\text{g}^{-1} \text{ Chl a}$] and **b)** dimethoate compared to DBMIB [$\mu\text{mol } \mu\text{g}^{-1} \text{ Chl a}$] and
 278 DCMU [$\mu\text{mol } \mu\text{g}^{-1} \text{ Chl a}$]. Dimethoate inhibits the PSII beyond Q_A like DCMU (positive control), while
 279 DBMIB (positive control) the cyt b_6f complex. F_v : variable fluorescence; F_0 : minimum fluorescence; F_m :
 280 maximum fluorescence. Two technical replicates (t.r.) were taken per reagent concentration.
 281

282 The effect of dimethoate (Fig. 2a) and DCMU (Fig. 2b) on maximum photochemical efficiency
 283 (F_v/F_m) of the tested phytoplankton species is similar. The three green algal species show similar F_v/F_m
 284 values under different dimethoate concentrations and remarkable higher values than the two diatoms.
 285 The control values of F_v/F_m for the used green algae and diatoms are 0.74-0.75 and 0.64, respectively
 286 (Table 7), which is in accordance with the literature (Hofstraat et al., 1994; Bonenta et al., 2012;
 287 Lepetit et al., 2017; Qin et al., 2021). In the five phytoplankton species, F_v/F_m is slightly decreasing
 288 with increasing dimethoate concentrations, and clearly different from the control at concentrations
 289 higher than 3-6 $\mu\text{mol dimethoate } \mu\text{g}^{-1} \text{ Chl a}$, depending on the species. Dimethoate apparently does
 290 not decouple antennas from PSII. Otherwise, the free moving antennas would fluoresce stronger in
 291 the ground fluorescence (higher F_0), resulting in a much lower F_v/F_m compared to the control.



292

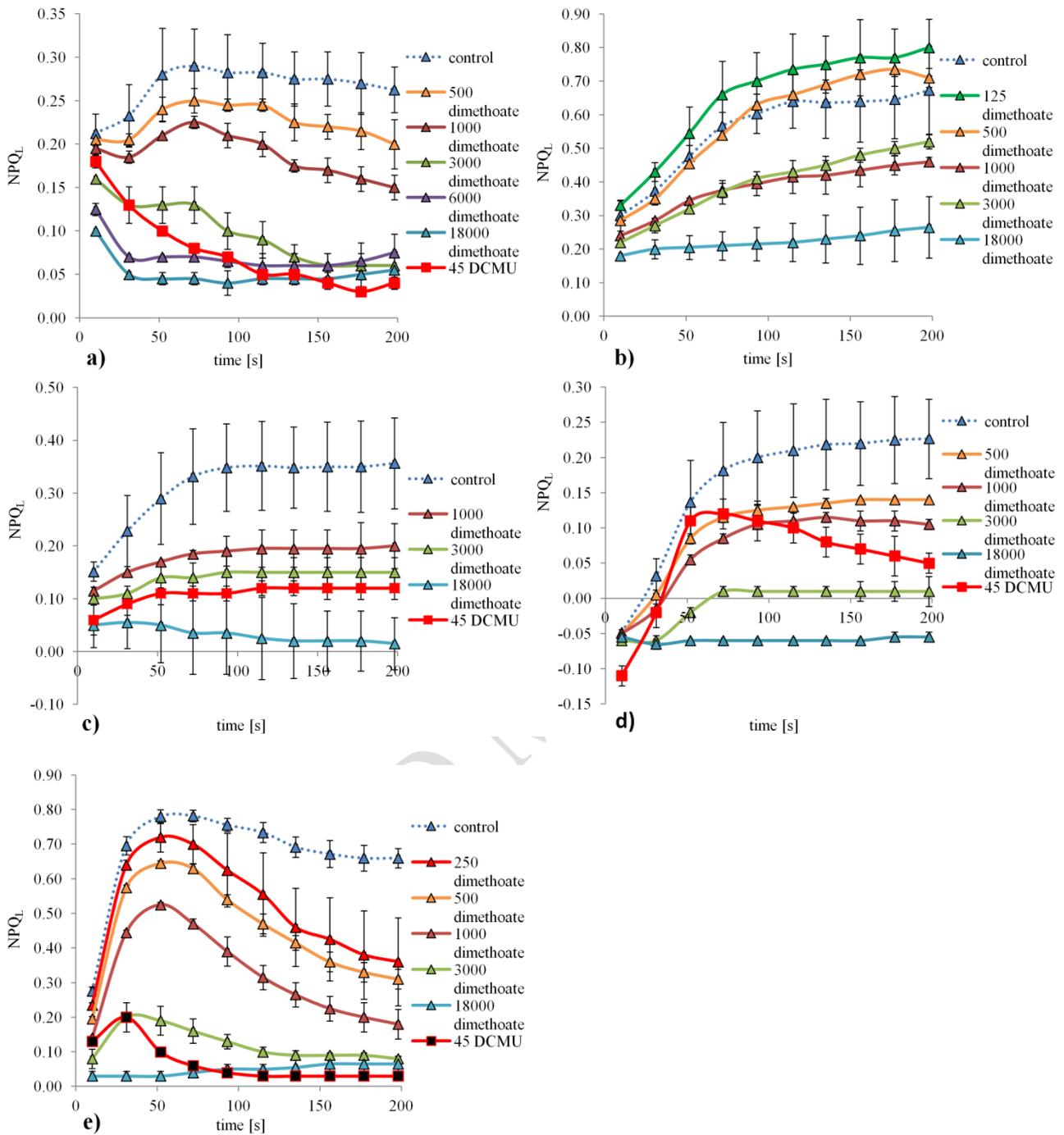
293 **Fig. 2:** Effect of increasing **a)** dimethoate and **b)** DCMU concentrations on maximum photochemical
 294 efficiency (F_v/F_m) of the five selected phytoplankton species. Two technical replicates (t.r.) per F_v/F_m value.

295 3.2.2 NPQ measurements

296 As outlined before, NPQ is a complex phenomenon that relies on different mechanisms, classically
 297 qE, qT and qI, though recently other factors have been added, such as qH (Malnoe, 2018) and qZ
 298 (Nilkens et al., 2010). Thus, NPQ analyses alone do not provide major insights into the respective
 299 mechanisms. However, as on a short-term scale NPQ processes are usually triggered directly or
 300 indirectly by the delta pH (which is also connected to the redox state of the plastoquinone pool, the
 301 trigger for state transitions), we performed NPQ analyses to further corroborate the effect of
 302 dimethoate on the photosynthetic ETC. NPQ in the light (NPQ_L) decreases with increasing
 303 dimethoate concentrations in the five tested phytoplankton species compared to the control (Fig. 3a-
 304 e), probably because dimethoate inhibited PSII and thus the establishment of a transthylakoidal proton
 305 gradient, the prerequisite for qE. NPQ_L development of *P. tricornutum* is similar to *Tetraselmis sp.*
 306 in presence of dimethoate (Fig. 3d,c). $3 \mu\text{mol dimethoate } \mu\text{g}^{-1} \text{ Chl a}$ showed a similar effect on NPQ_L
 307 development as $45 \text{ pmol DCMU } \mu\text{g}^{-1} \text{ Chl a}$ in *C. reinhardtii*, *Tetraselmis sp.* and *T. pseudonana* (Fig.
 308 3a,c,e).

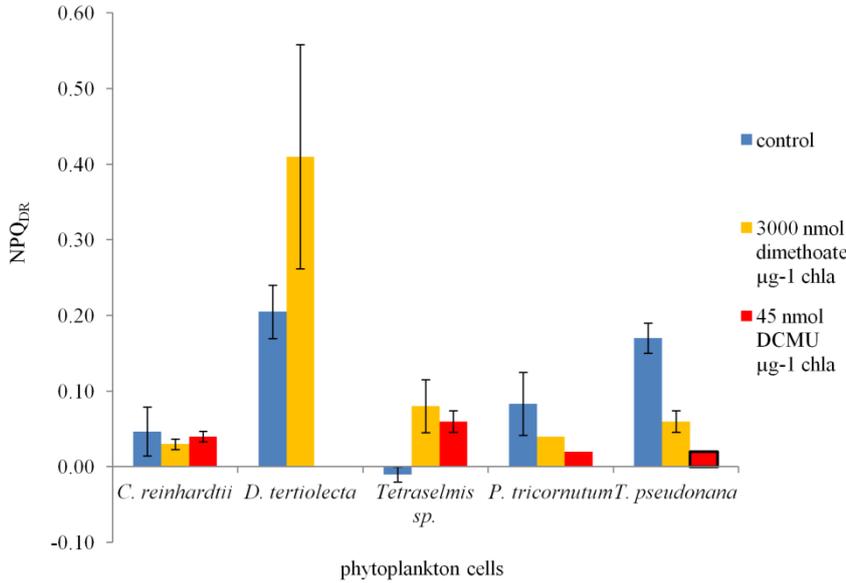
309 NPQ_L measured in *C. reinhardtii* was low, as the LHCSR pigment-binding proteins that induce NPQ_L
 310 are little expressed under the tested light conditions. According to Nawrocki et al. (2020), the

311 expression of these complexes depends on prior HL exposure of the cells. A negative development
312 of NPQ_L was observed in presence of the photosynthetic inhibitors over time, potentially induced by
313 state transition, which is very prominent in *C. reinhardtii* (Nawrocki et al., 2016).
314 In *D. tertiolecta*, NPQ_L increases over time (Fig. 3b). In the presence of dimethoate, this increase is
315 lowered as expected, with high concentrations of dimethoate completely suppressing NPQ
316 development.
317 NPQ_L of *T. pseudonana* steeply increases during the first 50 seconds and strongly decreases
318 afterwards at concentrations up to 1 $\mu\text{mol dimethoate } \mu\text{g}^{-1} \text{ Chl a}$ (Fig. 3e). Probably the immediate
319 increase of NPQ is directly triggered by the proton gradient alone and is relatively independent of the
320 xanthophyll cycle. The Calvin cycle needs some time to get activated and thus the proton gradient is
321 especially high directly after light onset. After the activation of the Calvin cycle, NADPH and ATP
322 are consumed, resulting in a partial proton gradient relaxation, followed by an NPQ_L decrease. A
323 similar NPQ kinetic has been observed in the highly related diatom *Cyclotella meneghiniana*, where
324 such a fast, pH dependent qE component was characterized in depth (Grouneva et al., 2008, 2009)



328 **Fig. 3:** Effect of increasing dimethoate concentrations [nmol µg⁻¹ Chl a] on NPQ development in the light
 329 (NPQ_L) of **a)** *C. reinhardtii*, **b)** *D. tertiolecta*, **c)** *Tetraselmis* sp., **d)** *P. tricoratum* and **e)** *T. pseudonana*.
 330 DCMU: positive control. Two technical replicates were taken for each NPQ_L value.

331 NPQ in the dark (NPQ_{DR}) of the phytoplankton species treated with 3 μmol dimethoate μg⁻¹ Chl a is
 332 similar to the control and to the 45 pmol DCMU μg⁻¹ Chl a-treated species (Fig. 4), which indicate
 333 that no photoinhibition of the algae occurs after pesticide application.



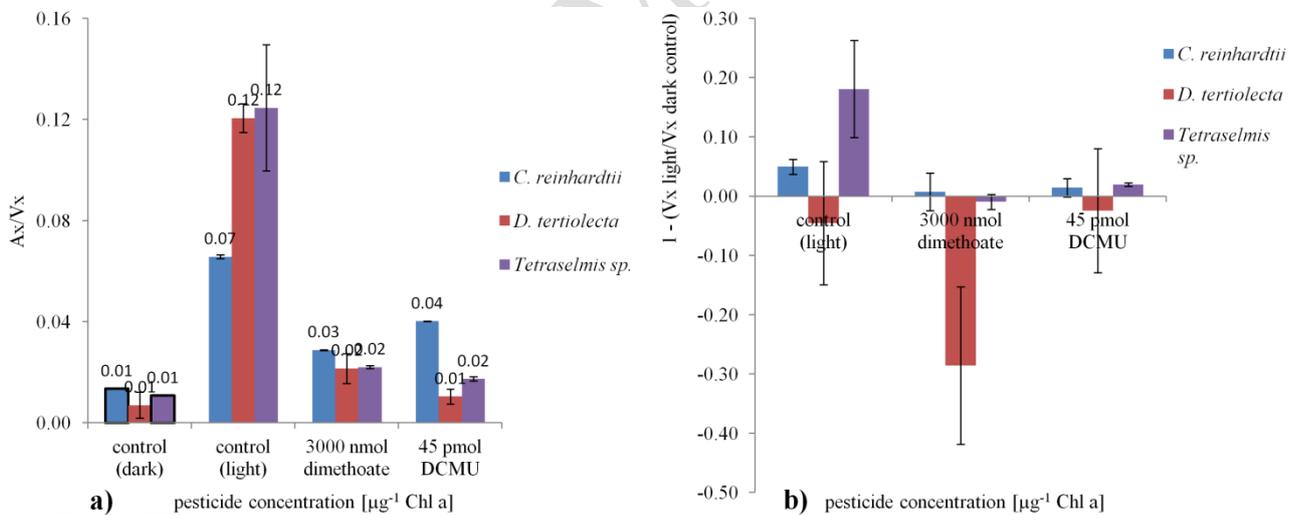
334
 335 **Fig. 4:** Effect of dimethoate and DCMU on NPQ dark relaxation (NPQ_{DR}) of the tested phytoplankton species.
 336 DCMU: positive control. 2-3 technical replicates (t.r.) per control, 2 t.r. per reagent-treated sample, except of
 337 one t.r. for DCMU-treated *T. pseudonana* (black edging). No data for DCMU treatment is available for *D.*
 338 *tertiolecta*.

339 3.3 Xanthophyll pigment measurements

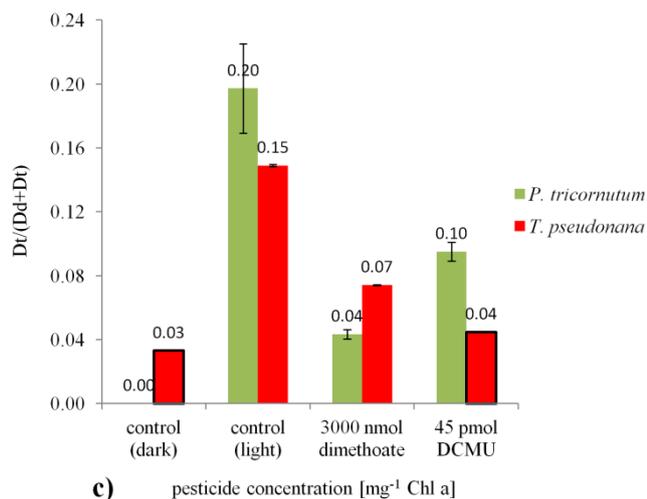
340 Another piece of evidence for the acting site of dimethoate in the photosynthetic apparatus was
 341 obtained through xanthophyll pigment analysis by HPLC. While we could measure the de-
 342 epoxidation state Dt/(Dd+Dt) for the diatoms, in the green algae we could not distinguish Zx from
 343 lutein in our HPLC approach. Hence, we measured Ax/Vx, although Ax/Vx alone does only provide
 344 limited information about the operating xanthophyll cycle, as a low Ax/Vx ratio could indicate a
 345 general low xanthophyll cycle activity, but also hide a fast de-epoxidation of Ax into Zx. However,
 346 the DES (Ax+Zx)/(Vx+Ax+Zx) can be converted into 1 - (Vx light)/(Vx dark control) and here the
 347 relative decrease of Vx reflects the xanthophyll cycle activity (for details see chapter 2.6). The
 348 xanthophyll DES measurements show that dimethoate acts like DCMU. Due to its potential binding

349 at the Q_B site of PSII, dimethoate disrupts the electron flow towards the $cytb_6f$ complex, and thus
 350 decreases the proton gradient, that is responsible for activating the de-epoxidation reaction. A_x/V_x of
 351 the tested green algal species (Fig. 5a) and $Dt/(Dd+Dt)$ of the diatom species (Fig. 5c) decreased
 352 compared to the control in presence of dimethoate and DCMU, while a V_x de-epoxidation was
 353 obvious under HL conditions in absence of the pesticides.
 354 As expected, $1 - (V_x \text{ light})/(V_x \text{ dark control})$ decreased under HL conditions in dimethoate and
 355 DCMU-treated *C. reinhardtii* and *Tetraselmis* sp. cells compared to the control (Fig. 5b).
 356 Interestingly, this parameter even reached negative values in *D. tertiolecta*. Probably, the HL
 357 exposure time was too low in order to start a pronounced de-epoxidation in this species, in line with
 358 results obtained by Casper-Lindley and Björkman (1998), and at the same time, there is some de novo
 359 synthesis of V_x .

360
361



362

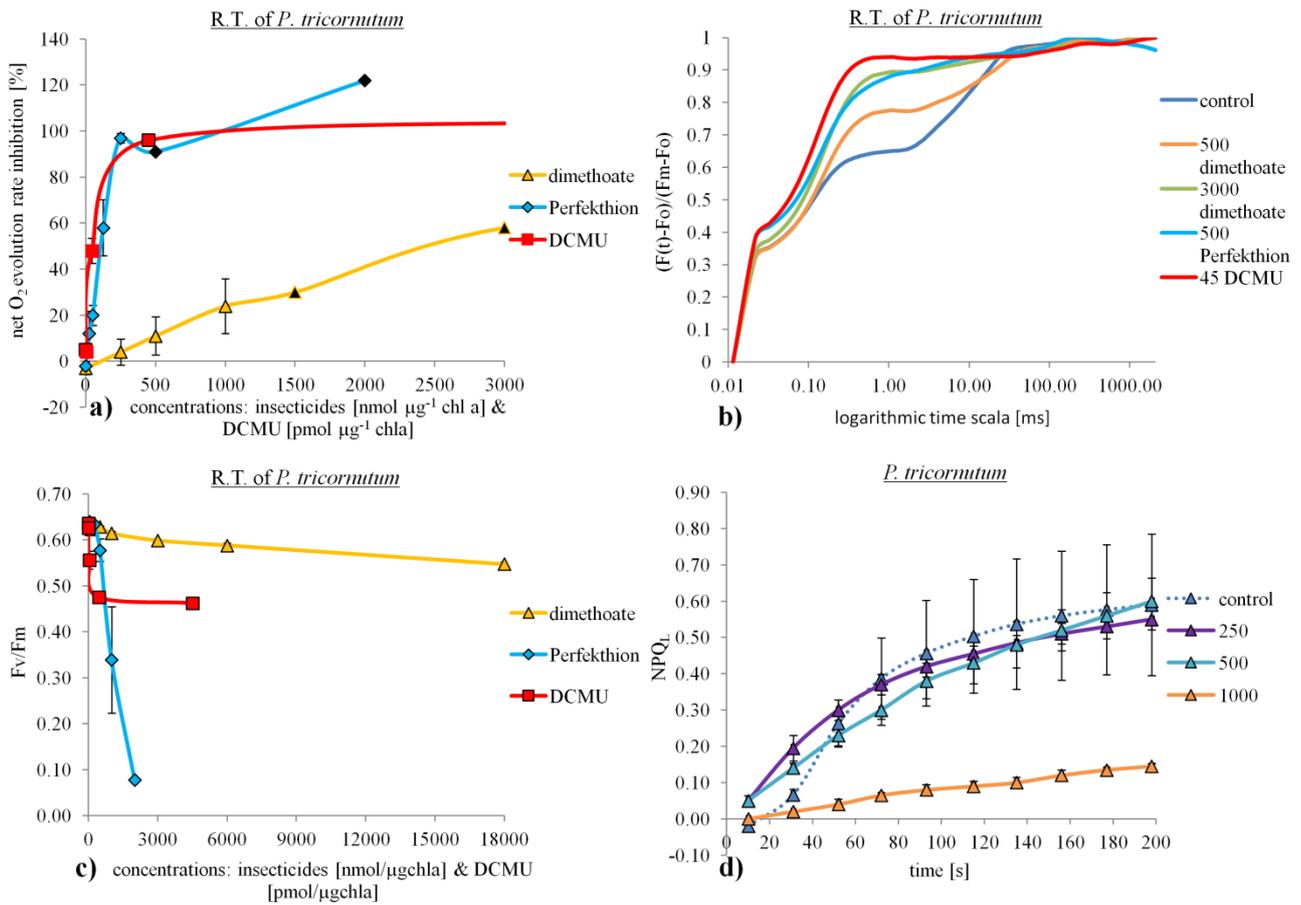


363

364 **Fig.5:** Effect of dimethoate and DCMU under HL conditions on the de-epoxidation state of **a)** A_x/V_x and **b)**
 365 $1 - (V_x \text{ light}/V_x \text{ dark control})$ that corresponds to $(A_x+Z_x)/(V_x+A_x+Z_x)$ of the 3 tested green algal species
 366 and **c)** $D_t/(D_d+D_t)$ of the 2 diatom species. V_x light: V_x value of the control or the insecticide-treated sample
 367 under HL conditions; V_x dark control: control value of V_x in the dark. DCMU: positive control. At least 2
 368 technical replicates (t.r.) for each measurement; columns with black edging shows one t.r..

369 3.4. Effect of Perfekthion on photosynthesis of the five selected phytoplankton species

370 As with dimethoate and DCMU, the net oxygen evolution rate inhibition increases with increasing
 371 concentrations of Perfekthion in the five phytoplankton species. A representative trace (R.T.) for all
 372 net oxygen evolution measurements is depicted for *P. tricornutum* (Fig. 6a). The single net oxygen
 373 evolution rate data of the pesticide-treated *P. tricornutum* is listed in Table S2 (supplement). The
 374 LOEC of the tested phytoplankton cells is $0.05 \mu\text{mol Perfekthion } \mu\text{g}^{-1} \text{ Chl a}$. $0.5 \mu\text{mol Perfekthion}$
 375 $\mu\text{g}^{-1} \text{ Chl a}$ inhibits the net oxygen evolution rate of the tested phytoplankton species by more than
 376 90%, while the same concentration of dimethoate shows an inhibition of 14-35% (the value of *C.*
 377 *reinhardtii* was not remarkably different to the control). Perfekthion inhibits the photosynthesis of
 378 the tested phytoplankton species stronger than dimethoate, probably because of an additional effect
 379 of the solvents. Panda et al. (1998) observed that concentrations of Perfekthion (there named Rogor)
 380 higher than $100 \mu\text{M}$ increase the permeability of the plasma membrane in *Chlorella vulgaris*.



381

382 **Fig. 6:** Effects of different concentrations of dimethoate and Perfekthion [nmol µg⁻¹ Chl a] and DCMU [pmol
 383 µg⁻¹ Chl a] on a) net oxygen evolution rate (b.r.), b) OJIP fluorescence (2 technical replicates (t.r.)), c) F_v/F_m
 384 (t.r.) and d) NPQ_L development (t.r.) of *P. tricornutum*. DCMU: positive control. Black filled symbols are
 385 mean values of 1 t.r./b.r., black framed symbols mean values of 2 t.r./b.r. and symbols with no black colour at
 386 all mean values of 3 t.r./b.r.. F_v: variable fluorescence; F₀: minimum fluorescence; F_m: maximum fluorescence;
 387 R.T.: representative trace; t.r./b.r.: technical/biological replicate.

388 Perfekthion does not affect the respiration of the tested phytoplankton species, except of *Tetraselmis*
 389 sp., where the respiration is decreased in presence of a higher concentration, i.e. 2 µmol Perfekthion
 390 µg⁻¹ Chl a (Table 6). This high concentration was not tested in the other two green algae.

391 **Table 6:** Respiration rate of *Tetraselmis* sp. measured in the dark, in presence of Perfekthion. t.r.: technical
 392 replicate.

	Control (3 t.r.)	0.5 µmol Perfekthion µg ⁻¹ Chl a (3 t.r.)	2 µmol Perfekthion µg ⁻¹ Chl a (2 t.r.)
Respiration rate	60 ± 18	57 ± 9	5 ± 6

$[\mu\text{mol mg}^{-1} \text{ Chl a h}^{-1}]$			
--	--	--	--

393 According to the OJIP fluorescence measurements, 0.5 $\mu\text{mol Perfekthion } \mu\text{g}^{-1} \text{ Chl a}$ inhibits PSII of
 394 the tested algae similar to 3 $\mu\text{mol dimethoate } \mu\text{g}^{-1} \text{ Chl a}$ and 45 $\mu\text{mol DCMU } \mu\text{g}^{-1} \text{ Chl a}$ (Fig 6b). But
 395 0.5-1.0 $\mu\text{mol Perfekthion } \mu\text{g}^{-1} \text{ Chl a}$ decreases F_v/F_m of the used phytoplankton species even stronger
 396 than 18 $\mu\text{mol dimethoate } \mu\text{g}^{-1} \text{ Chl a}$ (Table 7 and Fig. 6c). At these concentrations, the solvents of
 397 Perfekthion probably make the cell membranes highly permeable to the active ingredient
 398 (dimethoate), and dissociate either the chlorophylls from the proteins or the light harvesting complex
 399 from PSII, which leads to an increased F_0 fluorescence. The traces for *P. tricornutum* (Fig. 6b,c) are
 400 representative for the respective experiments.

401 **Table 7:** F_v/F_m of the tested phytoplankton species after addition of dimethoate and Perfekthion. b.r./t.r.:
 402 biological/technical replicate. Remarkable F_v/F_m -values in bold font.

Insecticide concentrations [$\mu\text{mol } \mu\text{g}^{-1} \text{ Chl a}$]	<i>C. reinhardtii</i>	<i>D. tertiolecta</i>	<i>Tetraselmis</i> sp.	<i>P. tricornutum</i>	<i>T. pseudonana</i>
control (2-3 b.r.)	0.75 ± 0.00	0.75 ± 0.01	0.74 ± 0.04	0.64 ± 0.01	0.64 ± 0.02
18 dimethoate (2 t.r.)	0.73 ± 0.001	0.69 ± 0.01	0.70 ± 0.002	0.55 ± 0.005	0.59 (1 t.r.)
0.5 Perfekthion (2 t.r.)	0.51 ± 0.01	0.31 ± 0.06	0.48 ± 0.05	0.58 ± 0.03	0.38 ± 0.01
1.0 Perfekthion (2 t.r.)	0.30 ± 0.02	0.16 ± 0.02	0.40 ± 0.04	0.34 ± 0.12	0.34 ± 0.01

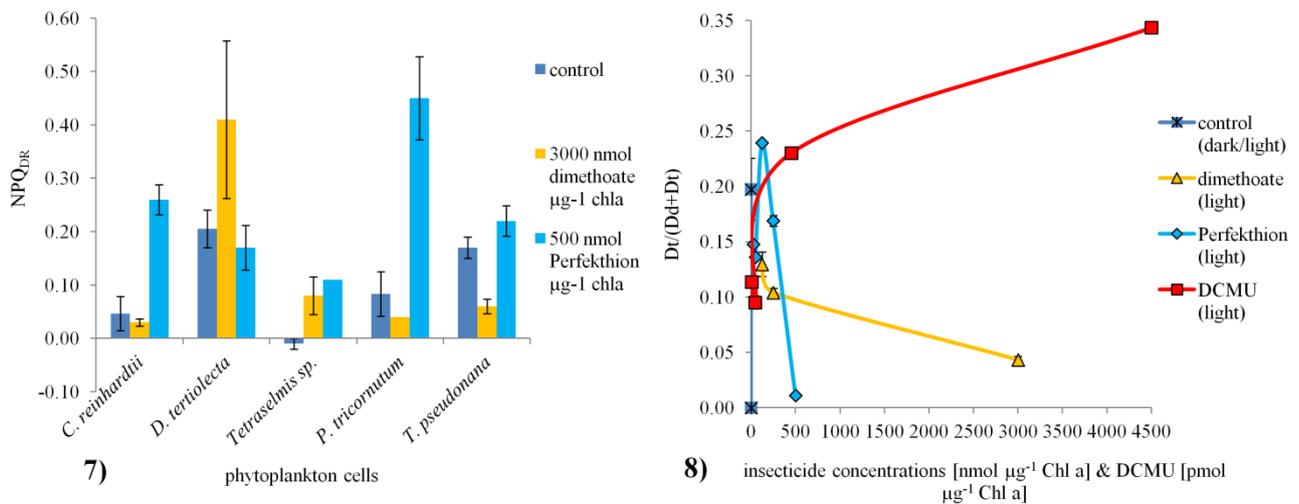
403 The effect of Perfekthion on NPQ_L of *P. tricornutum* (Fig. 6d) was similar to the effect of dimethoate.
 404 1 $\mu\text{mol Perfekthion } \mu\text{g}^{-1} \text{ Chl a}$ suppressed NPQ_L development.

405

406 0.5 $\mu\text{mol Perfekthion } \mu\text{g}^{-1}$ Chl a increases NPQ_{DR} in *C. reinhardtii* and *P. tricornutum* to values 5-10
407 times higher than the control or 3 $\mu\text{mol dimethoate } \mu\text{g}^{-1}$ Chl a (Fig. 7), indicating a photoinhibition
408 in the first species and potentially a Dt-dependent quenching (qZ) with a combination of qI in *P.*
409 *tricornutum*. The photosystems get probably destroyed due to the block of the photochemical
410 quenching channel, resulting in increased triplet chlorophyll species and eventually higher amounts
411 of singlet oxygen. Additionally, Perfekthion may even detach chlorophylls from the pigment binding
412 proteins. However, according to Sridevi (2012), Rogor does not dissociate the LHC in *C. vulgaris*
413 cells (up to 500 μM tested), but the organophosphate insecticides quinalphos and chlorfenvinphos
414 dissociate it. The same author reported that the electron flow between the oxygen-evolving-complex
415 (OEC) and PSII is not significantly inhibited in presence of Rogor. To make a clear statement about
416 photoinhibition longer NPQ_{DR} measurements would be necessary.

417 While the xanthophyll DES measurement of the tested phytoplankton species show that dimethoate
418 acts like DCMU, in *P. tricornutum*, DCMU shows a high de-epoxidation (Fig. 8), well known from
419 other studies (Grouneva et al., 2009; Lepetit et al., 2013). It is possible that the complete inhibition
420 of PSII by DCMU, in contrast to the partial inhibition by dimethoate, leads to additional physiological
421 processes, such as a strong PSI cyclic electron transport that increases the transthylakoidal proton
422 gradient. Future experiments could show if the increase of Dt/(Dd+Dt) in presence of Perfekthion is
423 attributed to the same effect as imposed by DCMU, and whether a probable destruction of the

424 photosystems would be the reason that this effect gets reversed by higher insecticide concentrations.



425

426 **Fig. 7:** Effect of Perfekthion on NPQ_{DR} of the tested phytoplankton species compared to the control and
427 dimethoate treatment. Two technical replicates were taken for each measurement.

428 **Fig. 8:** Effect of increasing dimethoate, Perfekthion and DCMU concentrations on the DES Dt/(Dd+Dt) of *P.*
429 *tricornutum*. DCMU: positive control. Black filled symbols: mean values of 1 t.r.; black framed symbols: mean
430 values of 2 t.r.; symbols with another colour than black: mean values of 3 t.r..

431 Perfekthion contains 37.2% dimethoate (372 g l⁻¹ or 1.62 M). According to Raiffeisen (online) and
432 Profiflor GmbH (online), BVL and the chemical producing company BASF (BASF SE,
433 Ludwigshafen, Germany) respectively, recommended a dilution of Perfekthion in water of 0.83/1000-
434 3.5/1000, before applying on plants on the field. Thus, the concentration to be sprayed on the field
435 should be 1.3-5.7 mM, depending on the kind and size of the plant, and the pest. The LOECs of
436 Perfekthion measured on net oxygen evolution rate of the tested phytoplankton species was 50 nmol
437 μg⁻¹ Chl a (100 μM), demonstrating that the applied field concentration is ca. 10-60 times higher than
438 the LOEC. The sprayed volume of diluted Perfekthion per hectare should be 12-600 l (Raiffeisen,
439 Profiflor, online). If the water volume after a rainfall is estimated to be 10-50 l m⁻², the final
440 concentration of Perfekthion on the field after a rainfall would be 0.03-34.2 μM. After a dry period
441 evaporation could enhance the concentration of dimethoate in the commercial formulation. On the
442 other hand, dilution of Perfekthion in the aquatic environment is expected to decrease its

443 concentration to several orders of magnitude lower than applied on the field. In addition to the low
444 persistence (few days) of Perfekthion in the environment (Petsas et al., 2007), it is unlikely that
445 Perfekthion affects the photosynthesis of the tested algae in the aquatic environment.

446 **4. Conclusions**

447 Net oxygen evolution rate inhibition of the tested phytoplankton species increases with increasing
448 dimethoate concentrations. Dimethoate does not affect the respiration of the algae, except of
449 decreasing the respiration of *D. tertiolecta* and *Tetraselmis* sp. at concentrations higher than 6 μmol
450 μg^{-1} Chl a. OJIP fluorescence measurements show that dimethoate inhibits the acceptor side of PSII
451 beyond Q_A , similarly to DCMU, and does not inhibit PSI. In presence of 3 μmol dimethoate μg^{-1} Chl
452 a, F_v/F_m and NPQ_{DR} is similar to the control and thus, no indication of photoinhibition was observed.
453 Increasing dimethoate concentrations decreased the xanthophyll DES A_x/V_x and $D_t/(D_d+D_t)$ and
454 decreased NPQ_L , an indication of a blocked ETC. In all the experiments 3 μmol dimethoate μg^{-1} Chl
455 a act comparable to 45 pmol DCMU μg^{-1} Chl a.

456 Perfekthion, the commercial formulation of dimethoate, shows a stronger inhibition on the net oxygen
457 evolution rate and on PSII than dimethoate. 2 μmol Perfekthion μg^{-1} Chl a decreased the respiration
458 of *Tetraselmis* sp.. 0.5-1 μmol Perfekthion μg^{-1} Chl a decreased F_v/F_m , increased NPQ_L development
459 and increased NPQ_{DR} in most of the phytoplankton species compared to 3 μmol dimethoate μg^{-1} Chl
460 a, indicating that probably the solvents of Perfekthion dissociated the pigments and/or LHCs from
461 PSII.

462 According to the study, if the recommended concentration of Perfekthion on the field is used, it is
463 unlikely that it affects the photosynthesis of the selected phytoplankton species in the aquatic
464 environment after a rainfall. More specialised studies will help to identify the exact inhibiting location
465 of the insecticide beyond Q_A . Future research will be necessary to examine possible effects of
466 Perfekthion on the oxygen-evolving complex, the LHC, the mitochondria and the membrane
467 permeability of the phytoplankton cells.

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472 **References**

- 473 Adhikary S. P. (1989), Effect of pesticides on the growth, photosynthetic oxygen evolution and
474 nitrogen fixation of *Westiellopsis prolifica*, *J. Gen. Appl. Microbiol.*, **35**, 319-325.
- 475 Allen J.F. (1992), Protein phosphorylation in regulation of photosynthesis, *Biochim Biophys Acta*,
476 **1098**, 275–335.
- 477 Bailleul B., Rogato A., de Martino A., Coesel S., Cardol P., Bowler C., Falciatore A. and Finazzi G.
478 (2010), An atypical member of the light-harvesting complex stress-related protein family
479 modulates diatom responses to light, *Proc Natl Acad Sci USA*, **107**, 18214–18219.
- 480 Baker N.R. (2008), Chlorophyll fluorescence: a probe of photosynthesis in vivo, *Annu Rev Plant Biol*,
481 **59**, 89–113.
- 482 Bassi R. and Dall'Osto L. (2021), Dissipation of Light Energy Absorbed in Excess: The Molecular
483 Mechanisms. *Ann Rev Plant Biol*, **72**, 47-76.
- 484 Boisvert S., Joly D. and Carpentier R. (2006), Quantitative analysis of the experimental O-J-I-P
485 chlorophyll fluorescence induction kinetics. Apparent activation energy and origin of each
486 kinetic step, *FEBS Journal*, **273**, 4770-4777.
- 487 Bonente G., Pippa S., Castellano S., Bassi R. and Ballottari M. (2012), Acclimation of
488 *Chlamydomonas reinhardtii* to different growth irradiances, *J. Biol. Chem.*, **287** (8), 5833-5847.
- 489 Buck J.M., Sherman J., Bártulos C.R., Serif M., Halder M. and Henkel J. (2019), Lhcx proteins
490 provide photoprotection via thermal dissipation of absorbed light in the diatom *Phaeodactylum*
491 *tricornutum*, *Nature Communications*, **10**, 4167.

492 BVL (2019), EU-Genehmigung des Pflanzenschutzmittel-Wirkstoffs Dimethoat nicht erneuert,
493 Bundesamt für Verbraucherschutz und Lebensmittelsicherheit, Korrigierte Version der
494 Fachmeldung vom 1. Juli 2019, 18.09.2019.

495 Casper-Lindley C. and Björkman O. (1998), Fluorescence quenching in four unicellular algae with
496 different light-harvesting and xanthophyll-cycle pigments, *Photosynthesis Research*, **56**, 277-
497 289.

498 Chen Z., Juneau P. and Qiu B. (2007), Effects of three pesticides on the growth, photosynthesis and
499 photoinhibition of the edible cyanobacterium Ge-Xian-Mi (*Nostoc*), *Aquatic Toxicology*, **81**,
500 256-265.

501 Chen B., Jihua L., Xu G. and Li G. (2021), Lowering pO_2 Interacts with Photoperiod to Alter
502 Physiological Performance of the Coastal Diatom *Thalassiosira pseudonana*, *Microorganisms*,
503 **9**, 2541.

504 EFSA (2022), The 2020 European Union report on pesticide residues in food, *EFSA Journal*, **20** (3),
505 7215, 57 p.

506 EPA (2006), Reregistration eligibility decision for dimethoate. US Environmental Protection Agency
507 (EPA) office of pesticide programs. Interim reregistration eligibility decision for dimethoate
508 (12.06.2006).

509 EU (2019a), Commission Implementing Regulation (EU) 2019/1090, *Official Journal of the*
510 *European Union*, L 173, 39-41, 27.06.2019.

511 EU (2019b), Corrigenda, *Official Journal of the European Union*, L 235, 11, 12.09.2019.

512 Falkowski P., Barber R.T. and Smetacek V. (1998), Biogeochemical controls and feedbacks on ocean
513 primary productivity, *Science*, **281**, 200-206.

514 Falkowski P. and Raven J.A. (2007), *Aquatic Photosynthesis*, p. 89-90, Princeton University Press,
515 USA.

516 Field C. B., Behrenfeld M.J., Randerson J.T. and Falkowski P. (1998), Primary production of the
517 biosphere: integrating terrestrial and oceanic components, *Science*, **281**, 237-240.

518 Geel C., Versluis W. and Snel J. F.H. (1997), Estimation of oxygen evolution by marine
519 phytoplankton from measurement of the efficiency of Photosystem II electron flow,
520 *Photosynthesis Research*, **51**, 61-70.

521 Goss R. and Lepetit B. (2015), Biodiversity of NPQ, *Journal of Plant Physiology*, **72**, 13-32 .

522 Grouneva I., Jakob T., Wilhelm C. and Goss R. (2008), A new multicomponent NPQ mechanism in
523 the diatom *Cyclotella meneghiniana*, *Plant Cell Physiol*, **49**, 1217-1225.

524 Grouneva I., Jakob T., Wilhelm C. and Goss R. (2009), The regulation of xanthophyll cycle activity
525 and of non-photochemical fluorescence quenching by two alternative electron flows in the
526 diatoms *Phaeodactylum tricornutum* and *Cyclotella meneghiniana*, *BBA-Bioenergetics*, **1787**
527 (7), 929-938.

528 Guillard R.R.L. and Ryther J.H. (1962), Studies of marine planktonic diatoms. I. *Cyclotella nana*
529 Hustedt and *Detonula confervacea* Cleve, *Can. J. Microbiol.*, **8**, 229-239.

530 Guillard R.R.L. (1975), Culture of phytoplankton for feeding marine invertebrates, In Smith W.L.
531 and Chanley M.H (Eds.) *Culture of Marine Invertebrate Animals*, 26-60. Plenum Press, New
532 York, USA.

533 Hager A. and Stransky, H. (1970), Das Carotinoidmuster und die Verbreitung des lichtinduzierten
534 Xanthophyll-Cyclus in verschiedenen Algenklassen. I. Methoden zur Identifizierung der
535 Pigmente, *Arch. Mikrobiol.* **71**, 132–163.

536 Hager A. (1980), The reversible light-induced conversions of xanthophylls in the chloroplast, In:
537 *Pigments in Plants*, Czygan F.C. (Eds), 57-79, Fischer Press, Stuttgart, Germany.

538 Hofstraat J.W., Peeters J.C.H., Snel J.F.H. and Geel C. (1994), Simple determination of
539 photosynthetic efficiency and photoinhibition of *Dunaliella tertiolecta* by saturating pulse
540 fluorescence measurements, **103**, 187-196.

541 Horton P., Wentworth M. and Ruban A. (2005), Control of the light harvesting function of chloroplast
542 membranes: the LHCII-aggregation model for non-photochemical quenching, *FEBS Lett*, **579**,
543 4201–4206.

544 Jakob T., Goss R. and Wilhelm C. (1998), Activation of Diadinoxanthin De-Epoxidase due to a
545 Chlororespiratory Proton Gradient in the Dark in the Diatom *Phaeodactylum tricornutum*, *Plant*
546 *biol.*, **1**, 76-82.

547 Jeffrey S.W. and Humphrey G.F. (1975), New spectrophotometric equations for determining
548 chlorophylls a, b, c1 and c2 in higher plants, algae and natural phytoplankton, *Biochem. Physiol.*
549 *Pflanzen*, **167**, 191-194.

550 Jena Sridevi, Acharya S. and Mohapatra P. K. (2012), Variation in effects of four OP insecticides on
551 photosynthetic pigment fluorescence of *Chlorella vulgaris* Beij., *Ecotox. and Environ. Safety*,
552 **80**, 111-117.

553 Kobbia I.A., Shabana E.F., Khalil Z. and Zaki F.T. (1991), Growth criteria of two common
554 cyanobacteria isolated from Egyptian flooded soil, as influenced by some pesticides, *Water, Air,*
555 *and Soil Pollution*, **60**, 107-116.

556 Kungolos A., Emmanouil C., Tsiridis V. and Tsiropoulos N. (2009), Evaluation of toxic and
557 interactive toxic effects of three agrochemicals and copper using a battery of microbiotests,
558 *Science of the Total Environment*, **407**, 4610 - 4615.

559 Lavaud J. (2007), Fast regulation of photosynthesis in diatoms: mechanisms, evolution and
560 ecophysiology, *Funct. Plant Sci. Biotechnol.*, **1**, 267–287.

561 Lepetit B., Sturm S., Rogato A., Gruber A., Sachse M., Falciatore A., Kroth P.G. and Lavaud J.
562 (2013), High light acclimation in the secondary plastids containing diatom *Phaeodactylum*
563 *tricornutum* is triggered by the redox state of the plastoquinone pool, *Plant Physiol*, **161**, 853-
564 865.

565 Lepetit B. and Dietzel L. (2015), Light signaling in photosynthetic eukaryotes with 'green' and 'red'
566 chloroplasts, *Environmental and Experimental Botany*, **114**, 30-47.

567 Lepetit B., Gelin G., Lepetit M., Dturm S., Vugrinec S., Rogato A., Kroth P.G., Falciatore A. and
568 Lavaud J. (2017), The diatom *Phaeodactylum tricornutum* adjusts nonphotochemical
569 fluorescence quenching capacity in response to dynamic light via fine-tuned Lhcx and
570 xanthophyll cycle pigment synthesis, *New Phytologist*, **214** (1), 205-218.

571 Malnoe A. (2018), Photoinhibition or photoprotection of photosynthesis? Update on the (newly
572 termed) sustained quenching component qH, *Environ Exp Bot*, **154**, 123-133.

573 Moermond C.T.A., Van Vlaardingen P.L.A., Vos J.H. and Verbruggen E.M.J. (2008), Environmental
574 risk limits for dimethoate. Report 601714001/2008. *National Institute for Public Health and the*
575 *Environment*. RIVM

576 Mohapatra P. K., Schubert H. and Schiewer U. (1996), Short term toxicity effect of dimethoate on
577 transthylakoid pH gradient of intact *Synechocystis* sp. PCC 6803 cells, *Bull. Environ. Contam.*
578 *Toxicol.*, **57**, 722-728.

579 Mohapatra P. K., Schubert H. and Schiewer U. (1997), Effect of dimethoate on photosynthesis and
580 pigment fluorescence of *Synechocystis* sp. PCC 6803, *Ecotoxic. and Environ. Safety*, **36**, 231-
581 237.

582 Mohapatra P. K. and Schiewer U. (1998), Effect of dimethoate and chlorfenvinphos on plasma
583 membrane integrity of *Synechocystis* sp. PCC 6803, *Ecotoxicology and Environmental Safety*,
584 **41**, 269-274.

585 Mohapatra P. K. and Schiewer U. (2000), Dimethoate and quinalphos toxicity: Pattern of
586 photosynthetic pigment degradation and recovery in *Synechocystis* sp. PCC 6803, *Algological*
587 *Studies*, **99**, 79-94.

588 Mueller P., Xiao-Ping Li and Krishna K. N. (2001), Update on Photosynthesis: Non-Photochemical
589 Quenching. A Response to Excess Light Energy, *Plant Physiol*, **125** (4), 1558–1566.

590 Musmarra D., Zafeirakou A., Manakou V. and Emmanouil C. (2019), Efficient and sustainable
591 environmental management as a means of addressing current pollution issues, *Environmental*
592 *Science and Pollution Research*, **26**, 14703 - 14705.

593 Nawrocki W. J., Santabarbara S., Mosebach L., Wollman F.-A. and Rappaport F. (2016), State
594 transitions redistribute rather than dissipate energy between the two photosystems in
595 *Chlamydomonas*, *Nat Plants* **2**, **4**, 16031.

596 Nawrocki W. J., Xin L. and Roberta C. (2020), *Chlamydomonas reinhardtii* exhibits de facto
597 constitutive NPQ capacity in physiologically relevant conditions, *Plant physiology*, **182** (1), 472-
598 479.

599 Nilkens M., Kress E., Lambrev P., Miloslavina Y., Müller M., Holzwarth A.R. and Jahns P. (2010),
600 Identification of a slowly inducible zeaxanthin-dependent component of non-photochemical
601 quenching of chlorophyll fluorescence generated under steady-state conditions in *Arabidopsis*,
602 *BBA-Bioenergetics*, **1797**, 466-475.

603 Panda Sobhana S., Mohapatra P. K. and Mohanty R. C. (1998), Comparative toxicity of two
604 organophosphorus insecticides on membrane integrity of *Chlorella vulgaris*. I. effect on
605 membrane permeability, *Microbiol. Res.*, **153**, 363-368.

606 Panda Sobhana Subhadarshana (1998), Effect of organophosphorus pesticides on the membrane
607 integrity of *Chlorella vulgaris*, Thesis, Utkal university (Botany).

608 Perona E., Marco E. and Orus M. I. (1991a), Effects of dimethoate on N₂-fixing Cyanobacterium
609 *Anabaena* PCC 7119, *Bull. Environ. Contam. Toxicol.*, **47**, 758-763.

610 Perona E., Marco E. and Orus M.I. (1991b), Alteration of dinitrogen fixation and metabolism in
611 cyanobacterium *Anabaena* PCC 7119 by phosphamidon, *Environmental and Experimental*
612 *Botany*, **31** (4), 479-488.

613 Petsas A. S., Vagi M. C., Kostopoulou M. N., Pavlaki M. D., Smaragdaki N. M. and Lekkas T. D.
614 (2007), Acute toxicity and persistence of fenthion and dimethoate in the marine environment.

615 Proceedings of the 10th International Conference on Environmental Science and Technology
616 Kos island, Greece, 5-7.09.2007, 8 p..

617 Piska R. S. and Waghray S. (1991), The toxic effect of dimethoate on primary productivity of lake
618 ecosystem, *Indian J. Environ. Helth.*, **33**, 126-127.

619 Pringsheim E. G. and Koch W. (1964), www.epsag.uni-goettingen.de

620 Qin Z., Xiaomin X., Guangming M., Yehui T. and Gang L. (2021), Differential physiological
621 responses of small *Thalassiosira pseudonana* and large *Thalassiosira punctigera* to the shifted-
622 high light and nitrogen, *Journal of Marine Science and Engineering*, 9, 450, 14 p..

623 Quaas T., Berteotti S., Ballottari M., Flieger K., Bassi R., Wilhelm C. and Goss R. (2015), Non-
624 photochemical quenching and xanthophyll cycle activities in six green algal species suggest
625 mechanistic differences in the process of excess energy dissipation, *Journal of Plant Physiology*,
626 **172**, 92-103.

627 Ritchie R.J. (2006), Consistent sets of spectrophotometric chlorophyll equations for acetone,
628 methanol and ethanol solvents, *Photosynth Res*, **89**, 27-41. Ruffle S.V., Wang J., Johnston H.G.,
629 Gustafson T.L., Hutchison R.S., Minagawa J., Crofts A. and Sayre R.T. (2001), Photosystem II
630 Peripheral Accessory Chlorophyll Mutants in *Chlamydomonas reinhardtii*. Biochemical
631 Characterization and Sensitivity to Photo-Inhibition, *Plant Physiology*, **127**, 633-644.

632 Sapozhnikov D.I., Krasovskaya T.A. and Maevskaya A.N. (1957), Change in the interrelationship of
633 the basic carotenoids of the plastids of green leaves under the action of light, *Dokl Akad Nauk*
634 *USSR* **113**, 465-467.

635 Schansker G., Tóth S.Z. and Strasser R.J. (2005), Methylviologen and dibromothymoquinone
636 treatments of pea leaves reveal the role of photosystem I in the Chl a fluorescence rise OJIP,
637 *Biochim. Biophys. Acta*, **1706**, 250-261.

638 Schreiber U., Bilger W. and Neubauer C. (1994), Chlorophyll fluorescence as a nonintrusive indicator
639 for rapid assessment of *in vivo* photosynthesis, *Ecological studies*, **100**, 49-70.

640 Sridevi J., Acharya S. and Mohapatra P. K. (2012), Variation in effects of four OP insecticides on
641 photosynthetic pigment fluorescence of *Chlorella vulgaris* Beij, *Ecotox. and Environ. Safety*, **80**,
642 111-117.

643 Stirbet A. and Govindjee (2011), On the relation between the Kautsky effect (chlorophyll *a*
644 fluorescence induction) and photosystem II: basics and applications of the OJIP fluorescence
645 transient, *J Photochem Photobiol B*, **104**, 236-257.

646 Surekha P. Rani (1999), Effect of pesticide Rogor on chlorophyll content of *Chlorella vulgaris*.

647 Tomlin C.D.S. (1997), "The Pesticide Manual-A World Compendium", 11th Edition, ed. by Tomlin
648 C.D.S., British Crop Protection Council Publications Sales, Bear Farm, Binfield, Bracknell,
649 Berks RG42 5QE, UK.

650 Vagi M. C., Kostopoulou M. N., Petsas A. S., Lalousi M. E., Rasouli Ch. and Lekkas T. D. (2005),
651 Toxicity of organophosphorous pesticides to the marine alga *Tetraselmis suecica*, *Global NEST*
652 *Journal*, **7** (2), 222-227.

653 Van der Werf (1996), Assessing the impact of pesticides on the environment. *Agriculture, Ecosystems*
654 *& Environment*, **60**, Issues 2-3, 81-96.

655 Wientjes E., van Amerongen H. and Croce R. (2013), LHCII is an antenna of both photosystems after
656 long-term acclimation. *Biochim Biophys Acta*, **1827**, 420-426.

657 Wong P. K. and Chang L. (1988), The effects of 2,4-D herbicide and organophosphorus insecticides
658 on growth, photosynthesis and chlorophyll *a* synthesis of *Chlamydomonas reinhardtii* (mt+),
659 *Environ. Pollution*, **55**, 179- 189.

660 Yamamoto H.Y., Nakayama T.O.M. and Chichester C.O. (1962), Studies on the light and dark
661 interconversions of leaf xanthophylls, *Arch. Biochem. Biophys.*, **97**, 168-173.

662 *Photon Systems Instruments (PSI)*, Drasov, Czech Republic. AquaPen-C AP-C 100, Operation
663 Manual, accessed 09.2015, <www.psi.cz>, online.

664 Profiflor GmbH, Pulheim, Germany, accessed 10.2022, <[http://www.profiflor.de/Produkte/021-](http://www.profiflor.de/Produkte/021-dimethoat-text.htm)
 665 [dimethoat-text.htm](http://www.profiflor.de/Produkte/021-dimethoat-text.htm)>, online.

666 Raiffeisen, accessed 10.2022, <<https://www.raiffeisen.com/pflanzenschutzmittel/detail/024190-68>>,
 667 online.

668 **Supplement**

669 **Table S1:** Net oxygen evolution rates [$\mu\text{mol mg}^{-1} \text{Chl a h}^{-1}$] of each tested phytoplankton species after treatment
 670 with dimethoate and DCMU. 3 technical replicates (t.r.) per measurement, unless otherwise stated. Control 1.
 671 and control 2. were measured on different days.

	<i>C. reinhardtii</i>	<i>D. tertiolecta</i>	<i>Tetraselmis sp.</i>	<i>P. tricornutum</i>	<i>T. pseudonana</i>
Control 1.	119 ± 33	189 ± 17	303 ± 13 (2 t.r.)	193 ± 12	164 ± 44
3 μmol dimethoate μg^{-1} Chl a	93 ± 15	85 ± 8	146 ± 10	77 ± 8	92 ± 9
Control 2.	120 ± 13	-	209 ± 5	156 ± 27	204 (1 t.r.)
45 pmol DCMU μg^{-1} Chl a	60 (interpolated value)	-	44 ± 11	84 ± 10	63 (1 t.r.)

672
 673 **Table S2:** Net oxygen evolution rates [$\mu\text{mol mg}^{-1} \text{Chl a h}^{-1}$] of *P. tricornutum* after treatment with dimethoate,
 674 Perfekthion and DCMU. 3 technical replicates (t.r.) per measurement, unless otherwise stated. b.r. = biological
 675 replicate. Control 1.,2.,3 were measured on different days.

Dimethoate concentrations [$\mu\text{mol } \mu\text{g}^{-1}$ Chl a]	1 b.r.	Perfekthion concentrations [$\mu\text{mol } \mu\text{g}^{-1}$ Chl a]	2 b.r. unless otherwise stated	DCMU concentrations [pmol μg^{-1} Chl a]	1 b.r.
0 (control 1.)	193 ± 12	0 (control 3.)	162 ± 4	0 (control 2.)	156 ± 27
0.25	170 ± 2	0.025	115 ± 18	4.5	162 ± 4 (2 t.r.)
0.5	145 ± 3	0.05	106 ± 4	45	84 ± 10
1.0	123 ± 5	0.125	60 ± 12	450	7 ± 2 (2 t.r.)
1.5	91 ± 5	0.25	6 ± 5	4500	-8 ± 11 (2 t.r.)
3.0	77 ± 8	0.5	15 ± 21		

			(1 b.r., 2 t.r.)		
		2.0	-32 ± 19 (1 b.r., 2 t.r.)		

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ACCEPTED MANUSCRIPT