Brackish water treatment by ceramic TiO$_2$ low-pressure nanofiltration membranes

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Graphical abstract

Abstract

The application of ceramic Nanofiltration (NF) membrane in brackish water treatment is increasing due to strict water quality standards. The Ceramic NF membrane process is widely used nowadays and has replaced reverse osmosis (RO) membranes in many applications due to lower energy consumption and higher flux rates. This work aims to study tubular ceramic TiO$_2$ nanofiltration membrane separation performance for single and tertiary salt mixtures at low-pressure (2 bar). The results showed that the ions rejection is ordered as follows: R (Cl$^-$) ion size > R calcium (Ca$^{2+}$) ion size > R sodium (Na$^+$) ion size > R magnesium (Mg$^{2+}$) ion. The highest Cl$^-$ rejection was 46%, the highest Ca$^{2+}$ and Mg$^{2+}$ rejection was 44%, and the highest Na$^+$ rejection was 37%. The pure water flux through the investigated ceramic NF membrane for distilled water was calculated to be 5.29 x 10$^{-7}$ m$^3$ s$^{-1}$ m$^{-2}$ bar$^{-1}$, and the flux was dropped down for the salt mixture feed to 2.56 x 10$^{-7}$ m$^3$ s$^{-1}$ m$^{-2}$ bar$^{-1}$ due to the presence of salt particles in the feed.

Keywords: Ceramic membranes, nanofiltration, water purification, desalination, brackish water

1. Introduction

Water is one of the vital elements for plants, animals, and human beings; therefore, water availability for food production and residential uses is of paramount concern to governments and researchers, especially under the pressure of the significant increase in the world population. Using alternative non-conventional water resources is the desalination researcher’s goal worldwide (Hilal et al., 2007). Water can be categorized according to salinity into four types: seawater, brackish water of high salinity, brackish water of low salinity, and potable water. Brackish water’s salinity ranges between 1,000 and 15,000 ppm (Mohsen, 2007). Brackish water salinity is more than freshwater, but not as much as seawater. In principle, brackish water contains between 0.5 and 30 grams of salt per liter (Mohsen, 2007).

Jordan is classified among the ten poorest world countries in water, with less than 120 m$^3$ estimated yearly water consumption per person. According to the water and irrigation ministry, the water deficits were about 419 million cubic meters (MCM) in 2019 and will reach about 490 MCM by 2025. Jordan’s available total brackish water amount is about 219 MCM per year, with an average salinity of 1000-3000 ppm. The estimated ground brackish water is about 78 MCM per year, with an average salinity ranging from 1300 to 8028 ppm (Mohsen et al., 1999).

Mohsen et al. (Mohsen et al., 2007) stated that the available amount of brackish water in Jordan valley is about 80 MCM, which can be a potential water resource for irrigation if treated by economically feasible technologies. Alsarayreh et al. (Alsarayreh et al. 2017) investigated using a renewable energy source, such as solar energy, for brackish water desalting in Jordan Valley. They reported that applying photovoltaic (PV) energy systems for brackish water desalting in a single small-scale plant at the current electricity price is relatively high cost and not affordable for the farmers; however, it might be feasible for larger-scale desalination plants to supply reclaimed water for more than ten farms with governmental funding. The economic evaluation studies indicated that using PV-powered desalination systems for brackish water treatment from the government’s perspective is economically feasible even for small-scale desalination plants.

In the same direction, Taha and Al-Sa’ed (Taha & Al-Sa’ed, 2017) investigated the viability of applying reverse...
osmosis (RO) desalination system powered by solar energy to treat saline brackish water at the Marj Naajeh desalination plant in the Jordan Valley. An economic evaluation was benchmarked to traditional energy sources such as diesel generators and network electricity. The environmental impacts were also considered. The results recommended upgrading agricultural wells of different water capacities and quality. They also suggested that conducting additional research studies on the design, capacity, and efficiency of future desalination plants would be more beneficial.

Mohsen and Al-Jayyousi (Mohsen et al., 1999) used multi-criteria analysis to evaluate various desalination methods. The evaluation criteria implemented were based on technical, economic, and environmental aspects. In addition, both the quality and quantity of different brackish waters were assessed. The investigation showed that RO desalination technology is ranked as the highest suitable for treating brackish waters, followed by electrodialysis (ED) as the second most successful desalination technology.

1.1. Desalination using nanofiltration membrane

Membrane technology has been extensively applied over the last decade for brackish water, wastewaster, and seawater desalination due to its separation performance and low energy requirement (Schäfer et al., 1998). Nanofiltration (NF) is a well-known process to separate salts from saline brackish waters. It uses membranes manufactured from organic materials and polymers with pore sizes ranging from 0.1 nm to 10 nm. Contrary to RO membranes that operate at high pressure and can reject all solutes in water, NF membranes are characterized by low operating pressures and the rejection of particular solutes according to their size and charge (Hilal et al., 2004). The membrane system generally consists of a pre-treatment process of the feed water before the primary NF process, which can be followed by a post-treatment process. Using NF membrane separation technology, purifying and desalting brackish waters is currently used as an alternative to traditional salt separation processes (Al-Qadami et al., 2020; Al-Zoubi et al., 2007; N Hilal et al., 2004) to supply clean and secure water resources for agriculture. Recently, NF membranes were successfully used to treat brackish groundwater water found in South Tunisia (Kammoun et al., 2020).

NF membranes are applied to separate salts from saline waters, surface water, and groundwater (Al-Harashsheh et al., 2017). Surface water treatment using NF is a good option; however, the surface water usually has variable salt concentration due to rain dilution. Moreover, NF is an appropriate choice for eliminating organic matter (Hilal et al., 2004). On the other hand, NF is more appropriate than RO for lime softening and separating naturally occurring organic carbon materials such as color and harmful side products (Van der Bruggen & Vandecasteele, 2003). Retentions higher than 90% using NF systems were found for multivalent ions, whereas monovalent ions retentions were about 60–70% (Hilal et al., 2005; Schaep et al., 1998). Another study implemented NF membranes as a pre-treatment step for real seawater desalination using RO process in a pilot plant (Kaya et al., 2020). The integrated system gave a reasonable rejection of seawater salts (97%) at 30 bar. According to the recent study by Suhalim et al. (Suhalim et al., 2022), the separation mechanisms of ionic compounds in the polymeric NF membranes are due to size exclusion, Donnan exclusion, and dielectric exclusion.

NF membranes' application for water softening was studied and compared to other traditional techniques, such as pellet softening and granular activated carbon (Sombekke et al., 1997). The results revealed that NF membranes have several advantages in terms of lower investment costs and productivity. A significant advantage of applying NF is its ability to remove all hardness cations. Hence, they are applicable when treating unnecessary side streams to decrease hardness to low levels. This can be achieved by splitting only a part of the feed water to be softened by the NF membranes, and the permeate is then mixed with the bulk flow stream. In contrast, when applying the precipitated lime softening process, the hardness could not be reduced to levels below 50 mg/L CaCO₃. This side stream softening is practically not attainable (Hilal et al., 2004).

1.2. Ceramic nanofiltration membranes

Ceramic NF membranes are regarded as a good choice in many applications (Condom et al., 2004; Radeva et al., 2021; Zhu et al., 2018); the separation of solutes from saline waters using ceramic Nanofiltration (CNF) membranes were extensively investigated (Sondhi et al., 2003). Recently, CNF membranes have become a gradually favored option when treating saline water. CNF membranes have several advantages over other organic and polymeric NF membranes; chemical stability and the ability to operate at different extreme pH levels and high temperatures and pressures make these membranes more attractive for water treatment. Moreover, CNF membranes can resist vigorous backwashing and chemical cleaning, which help to control membrane fouling. The rejection of ions through CNF membranes is influenced by the concentration of salts and the feed water's pH. Correspondingly, trans-membrane pressure primarily affects ionic rejection (Skluzaček et al., 2007; Van Gestel et al., 2002b).

Many additives were used to modify NF membranes, including Al₂O₃, TiO₂, ZrO₂, and SiO₂. Of these, TiO₂ attracted attention due to its unique characteristics, e.g., higher chemical and physical stability, withstanding high temperatures, and long service life (Bhave, 1991; Guo et al., 2018; Koutsonikolas et al., 2010; Sekulić et al., 2004; Zhang et al., 2006).

Qi et al. (Qi et al., 2012) used γ-Al₂O₃ modified CNF membranes, and the results showed less chemical stability in severe conditions, thus limiting their industrial applications. Benfer et al. (Benfer et al., 2001) showed that the ZrO₂ NF membrane showed a rejection rate as high as 66.3% for SO₄²⁻. Cai et al. (Cai et al. 2015) modified
Pb/TiO₂ membranes via the colloidal sol-gel process and attained higher rejections of salts than undoped TiO₂ membranes.

Gestel et al. (Van Gestel et al., 2002a, 2006) studied the stability of γ-Al₂O₃, TiO₂, and ZrO₂-modified CNF membranes in both acid and alkaline solutions. The γ-Al₂O₃ membranes results present less stability in an aqueous solution with (pH 3-11); however, the TiO₂ and ZrO₂ membranes showed high stability for pH values from (1 to 13)

TiO₂ is gaining more popularity in ceramic NF membrane improvement owing to its unique characteristics like super-hydrophilicity and high chemical stability (Lu et al., 2016; Sekulić et al., 2004). Many researchers have focused on the development of TiO₂-modified ceramic NF membranes. Titanium dioxide (TiO₂) nanoparticles are also employed in membranes to increase salt removal from water (Sakarkar et al., 2021; Ursino et al., 2018). The retention of charged and uncharged solutes and solvent permeability is the most crucial characteristic of membrane performance (Luo & Yinhu, 2013; Schäfer et al., 1998; Wahab Mohammad & Sobri Takriff, 2003). More details about the fabrications of CNF membranes and their applications in the desalination process using membrane distillation techniques were reviewed by Tai et al. (Tai et al., 2019).

Liang investigated the viability of using TiO2 CNF membranes to treat an artificial brine experimentally (Liang, 2018). A laboratory-scale apparatus of two flat disc CNF membranes with different molecular weight cut-offs were applied to filtrate artificial brines at different concentrations of single salt or NaCl/Na₂SO₄ salts mixtures. The results indicated the low rejection of sulphate, chloride, and sodium ions. It was also realized that the sulphate rejection could be improved by decreasing ionic strength and adding sodium chloride according to the Donnan effect and co-ion competing. In addition, fouling problems were measured due to the drop in water permeability during the experiment. Moreover, the difference in molecular weight cut-off before and after brine filtration was considered an indication of membrane fouling.

Hudaib et al. (Hudaib et al., 2019) investigated the separation performance of low-pressure tubular TiO₂-CN membranes operating at 2 bar. The effect of the valences, type of ions, transmembrane pressure (TMP), and the Zeta potential of the membrane on the separation of ions was studied experimentally. A single and three salt mixtures of sodium chloride, sodium nitrate, and sodium sulphate were tested in the study. It was found that the rejection of ions is arranged in the order: sulphate ions > nitrate ions > chloride ions > sodium ions. The highest rejection ratios for sulphate, nitrate chloride and sodium ions were 62%, 51%, 42%, and 37%, respectively. In another study (Pérez-González et al., 2015), a polymeric NF270 membrane was investigated to remove the SO₄²⁻ / Cl⁻ mixture. It was found that SO₄²⁻ rejection was in the 75-96% range and chloride rejections between 2% and 11%. However, when NaCl concentration is increased, the rejection of SO₄²⁻ was reduced due to the decrease in the Donnan effect. Similar findings were reported by Krieg et al. (Krieg et al. 2005).

In another study by Reig et al., the NF270 membrane was used to reject NH₄⁺ /NO₃⁻ in a cross-flow apparatus, applying a spacer-filled feed channel of a rectangular geometry (Reig et al., 2016). The membrane permeability to sodium ions was found to be more significant than to chloride ions because of an electric field that caused negative rejections of the trace nitrate anion. The existing electric field highly accelerated this. Another study shows that solute transport through an NF membrane pore depends on the sieving mechanism and the surface force interaction between the solute and the pore (Xu & Lebrun, 1999).

1.3. Irrigation water treatment

Although irrigation water contains ions such as Na⁺, Ca²⁺, and Mg²⁺, which are essential for the plant’s growth, excessive amounts are harmful and can cause severe damage to the plants. High Na⁺ replaces Ca²⁺ and Mg²⁺ ions at the exchange complex, which carry a negative charge. This will negatively affect the soil’s quality leading to soil structure deterioration, and the soil aggregates will be dispersed. In the long run, this will reduce the soil permeability for infiltration, initiating unfavorable growing conditions for plants. A critical factor that should be considered is the sodium absorption ratio (SAR) which is a sum of the Ca²⁺ and Mg²⁺ ions concentration divided by the concentration of Na⁺ ions. The higher the SAR, the lower the Na⁺ hazard. The soil infiltration rate can be decreased if the irrigation water is treated with a high SAR value and the salinity level is reduced to low or moderate. (Ayers & Westcot, 1985)

Unfortunately, magnesium ions concentration in the brackish water in Jordan is high and can cause nutrition disorders for the plants. The high magnesium (Mg²⁺) ions concentrations constrain plant growth and decrease crop harvest amounts. If Mg²⁺ dominates in the irrigation water, i.e., the concentration ratio Ca²⁺/Mg²⁺ < 1, the possible influence of sodium might be augmented, and plant nutrition might be disordered (Ayers & Westcot, 1985).

On the other hand, most of the brackish water in the Jordan Valley resources has a high chloride concentration, more than 10 meq/L, which is harmful to the fruit crops. For irrigation purposes, the chloride content should be less than 3 meq/L using sprinkler irrigation equipment. Although chloride has minor adverse impacts on soil, it can toxify the crops and accumulate in the leaves of lemon and other crops irrigated by high chloride contents brackish water. The adverse effects of chloride begin at 10 meq/L concentration, and higher concentrations of more than 3 meq/L in water using sprinkler irrigation will create a burn in the plants’ leaves (Ayers & Westcot, 1985).

Therefore, this study aims to use Ceramic TiO₂ low-pressure NF membrane for single and tertiary salt mixtures in view of chloride removal of the brackish water desalination in Jordan. The efficiency and separation
performance for single and multivalent ions at low pressure, the effect of surface charge (at fixed pH), and its influence on the separation performance of the membrane were investigated to get the best conditions to enhance the separation performance of the NF membrane. Moreover, the effect of ionic strength on ionic rejection was tested.

2. Materials and methods

2.1. Saline water

In this experiment, samples of the saline water containing single salt and mixed salts of sodium chloride, magnesium chloride, and calcium chloride, with a concentration of 0.1 M, were used. All salts are obtained from Sigma-Aldrich, Germany. Both sodium hydroxide and hydrochloric acid (0.1 M) are used to control the solution pH. The membrane is cleaned by sodium hydroxide.

2.2. Ceramic nanofiltration membrane

A ceramic tubular TiO$_2$ Nanofiltration membrane (7.0 mm ID, 10.0 mm OD, 190 mm length and 0.9 nm mean pore radius) was obtained from inopor GmbH- Germany and used for water desalination. The hydrophilicity (contact angle) of the membrane measured by the supplier is about 42±2 °C.

The morphological structure of the used ceramic membrane was investigated using SEM analysis. Figure 1 shows the SEM image of the used membrane obtained by SEM (FEI Quanta 200, Purge, Czech Republic) and EDXS equipment (Amertek Inc., Paoli, PA, USA). The membrane layers (active and supporting) are clearly shown in Figure 1. A thicker skin layer is observed; therefore, higher rejection and lower permeate flux are expected.

![Figure 1 Ceramic NF membrane SEM image showing the active layer and support layer. (Scale bar: 100 μm)](image)

2.3. Experimental set-up

The main components of the bench-scale membrane rig used in this experiment, as shown in (Figure 2) are a peristaltic pump with variable speed (type 603S, Watson-Marlow, UK), magnetic stirrer (RW20, IKAMAG, UK), tubular membrane module, glass container, pressure-relief valve, PVC-reinforced flexible piping, flexible neoprene piping for the pump (Watson-Marlow, UK), flow-meter (Gemü Gebr Müller, Germany), pH/ORP controller (Oakton), Accumet pH/Ion/Conductivity meter (Fisher Scientific, Model 50), balance and stopwatch. The resultant solutions anion’s concentration was measured using ion chromatography (Dionex DX600 Ion Chromatograph), and the cation concentration was measured by inductively coupled plasma atomic emission spectroscopy (ICP-AES).

![Figure 2 The bench-scale set-up used in this study](image)

2.4. Experimental procedure

The separation performance of the investigated CNF membrane was studied by filtration of pure distilled water, then the single or tertiary salt solution, and finally, the distilled water to check the effect of the precipitation of the salt on the membrane surface. First, Distilled water permeated through the ceramic membrane at a constant inlet volumetric flow rate of 3.056×10$^{-5}$ m$^3$/s (110 L/h). Then, the pressure inlet was increased from 0.3 bar to 2.0 bar, which gives transmembrane pressure (TMP) values between 0.2 to 1.9 bar. Next, the permeation experiments for salts were carried out for single and mixed salts. Single salt solutions of NaCl, MgCl$_2$, and CaCl$_2$ were prepared with 0.1 M concentration at a pH of around 7, while mixed solutions were prepared from the three salts together with 0.1 M concentration for each salt. For each experiment, the pressure was increased by 0.2 intervals every 30 minutes, and the collected permeate was continued for 25 minutes. After each experiment, the membrane was cleaned with distilled water many times; after that, it was cleaned with 0.1M NaOH solution for 1 hour and washed with distilled water continuously for 6 hours. The test was repeated three times for all permeation experiments, and the average was recorded.

2.5. TMP measuring and membrane rejection

The permeate flux through a porous membrane is described as the transmembrane pressure driving force (TMP), divided by (R) the resistance to mass transfer, and the permeate viscosity, $\mu$ (Field et al., 1995).

\[ J = \frac{TMP}{R} \quad (1) \]

For distilled water filtration, R will represent the resistance to mass transfer associated with the clean membrane. During an experiment, the resistance to
permeation increases because of different mechanisms, like pore block, cake formation, and concentration polarization. In constant TMP operation, as R increases, the permeate flux declines, while, at constant flux operation, TMP increases as R increases (Tаниguchi et al., 2003). Characterizing nanofiltration membrane can be done by determining its critical flux. Below critical flux, no flux decline occurs with time, while above it, fouling can be observed (Miller et al., 2014). There are two types of critical flux: strong critical flux and weak critical flux. TMP starts to deviate from the distilled water flux in strong critical flux, but the solution and distilled water membrane resistance are the same.

Moreover, in weak critical flux, flux values are lower than distilled water flux, where the membrane resistance for the solution is different from that of the distilled water, and the membrane resistance changes with the increasing flux after the critical flux is reached (Bacchin et al., 2006; Field et al., 1995). Field et al. (Field et al., 1995) research provided specific observations of constant TMP and flux measurements. They observed low membrane fouling and total mass transfer resistance in constant flux experiments and high fouling and total mass transfer resistance in constant TMP experiments, possibly due to the rapid fouling at the beginning of constant TMP experiments.

The rejection (R) of the ceramic NF membrane of ion (i) is calculated by the following formula (Geraldes et al., 2008):

\[
R = 1 - \frac{C_{i,p}}{C_{i,f}}
\]

\(C_{i,p}\) is the ion concentration in the permeate (mol/L), and \(C_{i,f}\) is the ion concentration in the feed (mol/L).

While the TMP is estimated as follows (Piry et al., 2008):

\[
TMP = \left( \frac{P_{\text{inlet}} + P_{\text{outlet}}} {2} \right) - P_{\text{permeate}}
\]

The pressure at the permeate side is usually assumed to be equal to zero, and hence the TMP would be as follows:

\[
TMP = \left( \frac{P_{\text{inlet}} + P_{\text{outlet}}} {2} \right)
\]

The water permeability, WP (L.h\(^{-1}\) m\(^2\) bar\(^{-1}\)), was estimated as follows (N Hilal et al., 2005):

\[
WP = \frac{V_r} {A.t.MTP}
\]

Where \(V_r\) is the Volume of water permeate (L), \(A\) is effective membrane area (m\(^2\)), \(t\) is time (h), and \(MTP\) is the applied pressure (bar).

3. Results and discussion

3.1. Zeta potential measurement

The membrane surface charge was measured using the electrophoretic mobility of TiO\(_2\) membrane powder derived from the membrane top layer (Guizard et al., 1999). The experiments were run at pH ranging from 3 to 10 at room temperature (25±0.50 °C). Sodium chloride (NaCl) salt with 0.1M concentration was used. The pH of the solutions was adjusted by using 0.1M HCl and 0.1M NaOH. The crushed membrane was added to the prepared solution, and the zeta potential was measured. Different pH was used to understand the membrane properties and the salts’ effect on membrane charge.

Figure 3 shows the results of membrane zeta potential at different pH values, which indicates that the membrane zeta potential is decreased as the pH increases. The isoelectric point (ISP) determines the membrane charge, which is considered the main parameter in justifying the ion's rejection behavior at a pH of about 4.5. Moreover, Figure 3 shows that as the pH increases, the membrane zeta potential decreases, and the higher the pH, the more negative membrane zeta potential is obtained. Luo & Yinhua, Mänttäri et al., and Szoke et al. further investigated the effect of pH on NF performance (Luo & Yinhua, 2013; Mänttäri et al., 2006; Szoke et al., 2003).

\[ \text{Figure 3 Surface zeta potential for TiO2 NF membrane} \]

3.2. Single salt rejection

Figure 4 shows that the rejection of Na\(^{+}\) increases as TMP increases until it reaches the maximum value of 8% at 1.5 bar; then, it decreases and becomes negative, which means it passes freely through the membrane. The rejection of Cl\(^{-}\)ions from NaCl solution is increased as the TMP increases; this may be due to the electrostatic repulsion interaction between the negative charges of the membrane and the Cl\(^{-}\) ion. These results are consistent with many studies (Luo & Yinhua, 2013; Szoke et al., 2003; Teixeira et al., 2005). Figure 4 indicates that rejection of the Cl\(^{-}\) sharply increases after 1.5 bar and reaches the highest rejection of Cl\(^{-}\) which is about 22% at a TMP of 1.9 bar. This sharp increase may result from the negative rejection of Na\(^{+}\), and the effects of the membrane charge and the Cl\(^{-}\) cause more repulsion as no cations can neutralize part of the negative membrane charge. The electro-neutrality condition can clarify the low rejection of both Cl\(^{-}\) and Na\(^{+}\) ions. Both ions are diffused through the membrane to neutralize the charge on the permeate side.

In MgCl\(_2\), the rejection of Cl\(^{-}\) is higher than that of Mg\(^{2+}\), as illustrated in Figure 5. This result may be explained by the electrostatic interaction between the membrane charge and ion charge or the so-called Donnan exclusion. The attraction between negative membrane charge and positive Mg\(^{2+}\) leads to the passage of Mg\(^{2+}\) ions more freely through the membrane than Cl\(^{-}\). While both
negative charges of $\text{Cl}^{-}$ and membrane surface cause repulsion between the membrane and $\text{Cl}^{-}$ to reject $\text{Cl}^{-}$ and then back to the solution, other researchers obtained similar results (Labban et al., 2017; Yaroshchuk, 2008).

The rejections of $\text{Mg}^{2+}$ are negative (except at the lowest TMP value, 0.3 bar). Similar results were reported by Yaroshchuk (Yaroshchuk, 2008). Negative rejection does not mean that mass ($\text{Mg}^{2+}$ solute in this case) is being created; neither the mixture now has more $\text{Mg}^{2+}$ ions than it initially started with. Negative rejection only means that the system has a higher concentration of $\text{Mg}^{2+}$ in the permeate relative to the feed. In other words, negative rejection for a given ion species only occurs when more of that ion is in the smaller permeate volume relative to the much larger feed volume. In the same line, a membrane with negative rejection enhances the transport of that solute or ion across it (Yaroshchuk, 2008; Labban et al., 2017, Hudaib et al., 2019).

Besides, the bigger size of the $\text{Mg}^{2+}$ ion compared to the $\text{Cl}^{-}$ ion size plays a role in rejection, causing $\text{Mg}^{2+}$ to pass more freely through the membrane (Pontalier et al., 1997). The small rejection threshold might be due to TMP, which can overcome the charge force between the membrane and the ion charges and force the ions to pass through the membrane.

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The rejection of $\text{Cl}^{-}$ is the highest, followed by $\text{Ca}^{2+}$, then $\text{Mg}^{2+}$ and $\text{Na}^{+}$, which is negative after (0.3) TMP. All the rejection values sharply decrease at lower TMP, i.e., below 0.3 bar, almost constant or slightly varies after increasing the TMP over 0.3 bar with the same ranking. The higher rejection values of $\text{Cl}^{-}$ are due to the repulsion of the negative charges of both $\text{Cl}^{-}$ and the membrane, while the other three ions have a positive charge opposite to the membrane charge, which causes attraction between the membrane and the cations, allowing those
positive ions to permeate more freely through the membrane. In addition to the size of the cations playing a role in rejection, the sizes are arranged as follows (Cl\(^-\) ion size > calcium (Ca\(^{2+}\)) ion size > sodium (Na\(^{+}\)) ion size > magnesium (Mg\(^{2+}\)) ion size. The sodium (Na\(^{+}\)) ion size > magnesium (Mg\(^{2+}\)) ion size, but the charge of Mg\(^{2+}\) is double the charge of Na\(^{+}\), so the repulsion of magnesium is double that of Na. This filtration rejection (efficiency) will decrease with increasing the salt concentration. This conclusion was approved in our previous study (Al-Zoubi & Waid, 2009).

The low rejection values of the three investigated cations and Cl\(^-\) might be due to the electro-neutrality condition at both sides of the membrane, where ions permeate from the high concentration (feed side) to the low concentration (permeate side) to achieve the charge equilibrium condition.

A comparison of the ion’s filtration from a single salt solution and a mixed salt solution shows that the rejection of Mg\(^{2+}\) from a mixed salt solution is higher than that of a single salt solution; it showed negative as the TMP increases more than 0.3 bar. The rejection of Ca\(^{2+}\) from a mixed salt solution is higher than that of a single salt solution, with positive rejection in both cases. The mixed solution makes the rejection of Na\(^{+}\) negative and lower than that of the single salt solution. The rejection of the CI\(^-\) ion in all cases is still the highest of all ions, while the rejection of CI-1 from calcium chloride is higher than that of magnesium chloride and sodium chloride mono solutions.

3.4. Permeate flux behavior

Pure water flux for the investigated membrane was found to increase linearly with the transmembrane pressure, as shown in Figure 8. The permeate flux vs. pressure was obtained initially for distilled water (distilled water 1), then it was calculated for the salt mixture (NaCl, MgCl\(_2\), and CaCl\(_2\)). After the salty water experiment, the membrane was cleaned using the following steps: Initially, the membrane was washed with distilled water many times to remove the salt ions from the membrane surface. After that, the membrane was cleaned with 0.1M NaOH solution for 1 hour, and finally, the membrane was washed with distilled water for 4 hours until the ions were removed. The permeate flux is estimated again for the distilled water (distilled water 2). The water permeabilities estimated from Eq.4 were 5.29 x 10\(^{-7}\), 4.72 x 10\(^{-7}\) and 2.56 x 10\(^{-7}\) m\(^3\) s\(^{-1}\) m\(^{-2}\) bar\(^{-1}\) for distilled water 1, salty water, and distilled water 2, respectively. The permeate flux increases with pressure due to the increase in solvent flux. Transport through the NF membrane can be explained in terms of convection and diffusion (N Hilal et al., 2004); the water flux is higher at a higher pressure, and the contribution of diffusion becomes less important relative to convection.

On the other hand, the permeate flux of salty water was relatively low due to the salt particles’ presence in the feed water, which obstructs the diffusion of permeate water through the surface of the NF membrane. The flux of distilled water 2 was lower than that of distilled water 1 due to the precipitation of some of the salt particles over the membrane surface (concentration polarization), which occurred after carrying out a salt mixture’s filtration of NaCl, MgCl\(_2\), and CaCl\(_2\) (N Hilal et al., 2004).

Figure 8 The permeate flux vs. pressure for distilled water 1, salty mixture (NaCl, MgCl\(_2\), and CaCl\(_2\)), and distilled water 2.

4. Conclusion

The objective of this study is to investigate the separation performance of the ceramic NF membrane of saline water solutions containing single and mixed salts of NaCl, MgCl\(_2\), and CaCl\(_2\) under low pressure. The results showed that the rejections of Cl\(^-\) ions are always the highest due to the negative charge of both the chloride and the membrane. The rejection of Na\(^{+}\) increased as TMP increased until it reached the maximum value of 8% TMP at 1.5 bar, then started to decrease and become negative, which means it passed freely through the membrane. While the rejection of calcium cations is positive, on the contrary, the rejection of magnesium cations is negative; that is, the membrane freely allows the cations to go through. As a result, the Cl\(^-\) ion has a higher rejection than the Ca\(^{2+}\) ion. The highest Ca\(^{2+}\) rejection was 43.1%, the highest Mg\(^{2+}\) rejection was 42.1%, and the highest Na\(^{+}\) rejection was 33.0%. The highest Cl\(^-\) rejection was 44.6%. The sequence of rejection is (Cl\(^-\) ion size > calcium (Ca\(^{2+}\)) ion size > sodium (Na\(^{+}\)) ion size > magnesium (Mg\(^{2+}\)) ion size. The ion’s size and the charge, and the value of the charges play an essential role in the rejection process. Based on the obtained results, a CNF membrane is potentially recommended to filtrate salty water with high Cl\(^-\) concentration.

Many combined processes were investigated for brackish water treatment to achieve the best rejections, such as using NF technology with RO to get optimal brackish water treatment (Cai et al., 2020). Chen et al. used a dual-stage NF system to treat brackish water in the Binhai Area of Tianjin (with TDS > 12,000 mg/L); A quartz sand filter activated carbon and UF were used as a pre-treatment unit (Chen et al., 2013).

However, in this study, authors investigated the use of a low-pressure NF process to study the way of rejecting different types of ions. Thus in future work, a combined cheap pre-treatment process will be investigated with the NF membrane used in this study.
References


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