- Optimization Approach to Evaluate the Solar, UV and LED Visible Light Fenton Processes
 for Pollutant Degradation in Landfill Leachate
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- 9 GRAPHICAL ABSTRACT

Fenton Dosage

Graphical Abstract: Optimization Approach to Evaluate the Solar, UV and LED Visible Light Fenton Processes for Pollutant Degradation in Landfill Leachate

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- 11 Abstract

In the current study, three different photo-assisted sources were used to evaluate the photo-Fenton oxidation process for the pollutant degradation of mature landfill leachate. Solar, UV, and LED visible light photo sources were adopted in this Fenton process. Fenton dosages (FeSO₄:H₂O₂) 15 1:30, 1.5:30, 2:30, and 2.5:30g/L were used in the investigation, which was conducted under a variety of pH settings (2 to 3.5) and at experimental reaction duration of 30, 60, 90, and 120 minutes. To analyze and compare the efficiency of the solar, UV, and LED visible light photo Fenton processes for pollutant degradation, Taguchi experimental design orthogonal arrays (L16) was employed. To analyse the significance of this experimental design for landfill leachate treatment, Larger the Better was selected. The highest levels of pollutant degradation in the Solar Photo Fenton method was identified when compared to UV, and LED visible light photo Fenton processes. The maximum removal of 95% color, 83% COD, 89% TSS, 94% Cr, 86% Cd, and 93% Cu was achieved in Solar Photo Fenton process at pH-3, Fenton dosage 1.5:30 g/L and reaction period 60 minutes.

25 Key words: Landfill leachate, UV, LED visible light photo Fenton process, Solar, Taguchi method.

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27 **1. Introduction**

The production of wastewater was mostly a result of human activity in the earlier days, it was not 28 considered as a serious issue. However, as urbanisation and population grew, these wastewaters 29 later had a significant negative impact on the environment (Giusti, 2009). The increase in trash 30 output on both developed and developing countries caused serious environmental, economic, and 31 social issues (Dharmarathne and Gunatilake, 2013). Solid wastes in India have emerged as the 32 major environmental issue because of its high population density. The landfill leachates were 33 regarded as a significant environmental concern since they have a major impact on soil, surface and 34 35 ground water (Mohan and Gokul, 2022). Leachate wastewater is highly contaminated, foulsmelling, dark-brown colored liquids that are formed when rainwater seeps through landfill layers 36 and when organic wastes decompose (Ghaffariraad and Ghanbarzadeh Lak, 2021). The environment 37 and aquatic organisms are harmed by the different organic chemicals, heavy metals, ammonia 38 nitrogen, soluble salts, BODs, CODs, suspended particles, etc. that are present in landfill leachate 39 (Christensen et al., 2001). The extremely concentrated ammonia nitrogen concentration of leachates 40 creates a variety of issues, including rapid eutrophication, algae growth, and dissolved oxygen 41 depletion (Niza et al., 2021). Landfill age, kind of trash disposed of, percolation, climatic 42

43 conditions, and other variables all influence the composition of the leachate. Based on their age
44 leachate from landfills are divided into three categories: young (5 years), intermediate (5–10 years),
45 and mature or stable leachates (>10 years) (Amor et al., 2015).

Treatment of leachate is necessary to control the ground water pollution, soil pollution and air 46 pollution; however it is challenging because of leachate's intricate structure. The numerous 47 available procedures for the treatment of landfill leachates include biological treatment methods, 48 membrane processes, coagulation, flocculation, and AOP (Avsar et al., 2007). Biological treatments 49 were considered as an efficient method for treating young leachates because of its higher 50 BOD_5/COD ratio. However, it was not preferred for treating mature leachate (Li et al., 2017) since it 51 contains high molecular weight contaminants, higher concentration of N₂ and other organic matter 52 (Klein et al., 2017) which are not easily biodegradable. One of the efficient methods to treat mature 53 leachates was the physicochemical treatment process. Among the various physicochemical 54 processes, AOP method was most preferred since it was cost effective and has good removal 55 efficiency. Biodegradability of mature leachate was improved by AOP (Deng and Englehardt, 56 2008). AOP uses Reactive Oxygen Species (ROS) for the mineralization of organic contaminants 57 (Ma et al., 2021). By using reactive chemicals species like hydroxyl radicals (OH⁻), the most 58 recalcitrant molecules were degraded into bio-degradable compounds (Gogate and Pandit, 2004). 59 The various technologies of AOP include photo catalytic oxidation, Fenton and Fenton-like 60 oxidation, ozonation, electrochemical oxidation and SR-AOPs. In general, combined AOP 61 technologies were preferred over individual AOP because it exhibits higher oxidation efficiency. 62

The ROS was efficiently generated by different AOP combinations that include TiO₂/UV, H₂O₂/ UV, UV/Fe²⁺/H₂O₂, UV/O₃, Fenton/ Photo Fenton, Fenton/ Electro Fenton and so on (Abdelhaleem and Chu, 2020; Rocha et al., 2011). Henry J. Fenton was discovered the Fenton process in 1894 (Vorontsov, 2019). One of the efficient technologies used in AOP is the Photo Fenton method. The Fenton technique was frequently enhanced by using a variety of photo sources. Based on the applied photo sources, the Photo Fenton process is classified as Solar photo Fenton, Visible light

photo Fenton and UV light photo Fenton. Hydroxyl radicals are used as the main oxidizing agent in 69 this process. The combination of ferrous ions and H₂O₂ forms the Fenton reagent. The Fenton's 70 reagent always uses a catalyst for the reaction process. The effective removal of contaminants was 71 72 achieved by combining high concentrated H₂O₂ and low concentrated ferrous ions. However, it cannot be directly applied to the contaminated surface since it kills the surface microorganisms 73 (Hartmann et al., 2010). The regeneration of ferrous ions and exhibiting of more hydroxyl radicals 74 in the Photo Fenton process Eq.(1) and (2) highly enhances the degradation of organic contaminants 75 when compared to any other conventional Fenton methods (Primo et al., 2008 and Bautitz et al., 76 2007). To produce the hydroxyl radicals, Hydrogen peroxide was being decomposed by ferric ions 77 under the presence of UV irradiation and under acidic conditions (Poblete and Perez 2020). 78

79
$$Fe^{2+}+H_2O_2 \rightarrow Fe^{3+}+OH^-+OH$$
 (1)

80
$$Fe^{3+}+H_2O+hv \rightarrow Fe^{2+}+OH+H^+$$
 (2)

In these reaction cycles complexes and hydroxides of iron ions were the major contributors. This 81 reaction can be driven by visible and UV light because of the broad absorption bands of these 82 complexes. Generally, Design of experiments (DOE) was used to examine the association of 83 multiple input factors and key output responses. There are several methods available to analyze the 84 design in DOE. The most commonly used methods are Response Surface Methodology and 85 Taguchi OA method of analysis. The orthogonal array experimental optimisation design was put 86 forwarded by Taguchi. In this design the effect of various factors on the performance characteristic 87 were examined in a condensed set of experiments. The parameters which affects a process and the 88 levels at which these factors should be fluctuated can be found by Taguchi method. Selection of the 89 proper orthogonal array was the most important process in TOA methodology. It was selected by 90 number of parameters and their levels (Fraley et al., 2006). Compared with other DOE methods, 91 Taguchi was quiet easy and convenient to use. Results obtained from Taguchi analysis closely 92 match the actual values. Taguchi analysis was performed using many software's like design expert 93

and Minitab. However, Minitab statistical software has many graphical outputs and user friendly
tools to analyze. Data transformation, analysis of variance, correspondence analysis, regression
analysis, multivariate analysis of variance, analysis of covariance, factor analysis, neural network,
correlation was also done by minitab software (Okagbue et al., 2021). The present study focusses on
solar, UV, and LED visible light photo Fenton processes used to treat landfill leachate. It also
compares and examines the pollutant degradation effects of each process by using Taguchi
statistical experimental method in Minitab software (Version 21.1.0).

101 **2. Material and Methods**

102 *2.1. Materials*

103 2.1.1. Raw leachate wastewater collection and chemicals used

The matured landfill leachate samples were obtained from a dumpsite in Dindigul, Tamil Nadu, at 104 various time periods between the months of March and April at same point of source. In order to 105 stop the material from decomposing, it was kept at 4°C. Every chemical utilised in analytical grade 106 was acquired from Merck. Chemicals utilised in this study included diluted HCl (1.0 N), sodium 107 chloride solution (0.1 N), Ferrous sulphate powder (FeSO₄·7H₂O=278.01), buffer solution, 108 hydrogen peroxide (30% W/V), and NaOH (1.0 N). Treated supernatant wastewater samples were 109 filter through Wattman filter paper No.40 and pollutant concentrations in leachate samples before 110 and after treatment were determined, and the effectiveness of the three methods (Solar Photo 111 Fenton, UV Photo Fenton and LED visible Fenton process) for removing pollutants was compared. 112

113 *2.1.2 Analytical procedures*

Physico-chemical characteristics of collected leachate sample were studied before and after treatment according to the guild lines of standard methods (APHA, 2013). Turbidity was measured using Portable turbidity meter (model: 331, make: Deep Vision) and values were recorded in nephelometric turbidity units (NTU). pH meter(model:111120306,make:Deep vision) used to measure and also adjust the pH value, UV-visible spectrophotometer(Model: DR6000, make: HACH, CO., USA) was used to calculate color removal, gravimetrical method to determine total solids, total suspended solids and total dissolved solids, Heavy metal concentration in before and after the treatment of leachate sample was studied using Atomic absorption spectroscopy (AAS) (model name – Shimadzu Atomic Absorption Spectrophotometer AA6300;ROM version-1.03)

123 2.2. Experimental setup

124 2.2.1. Solar Photo Fenton, UV Photo Fenton and LED visible Fenton process

Solar photo Fenton study was carried out in the open sunlight at terrace of main building, 125 University college of Engineering, Dindigul. The study was performed with 1000ml glass beaker 126 with magnetic stirrer to achieve even mixing (Pignatello et al., 2006). 500ml wastewater sample was 127 taken for each study. UV and visible light Fenton process carried out with artificial light sources of 128 Super 395nm UV light (size of UV lamp -3.5 cm \times 8.5 cm) and Wipro LED B22 Bulb Base with 129 50 lumen coverage. Stable radiation was maintained in UV and LED visible light Fenton process at 130 room temperature in laboratory. pH level was varied from 2 to 3.5, FeSO₄ concentration varied 131 from 1 to 2.5 g/L while H₂O₂(30% W/V) concentration was maintained at 30 g/L, reaction times of 132 30, 60, 90, and 120 were chosen for this study. Based on the findings of the preliminary 133 investigation, the control variable limitations were fixed. Bubbles rising during the advanced 134 oxidation process served as indicators that the oxidation process had begun in a solution. Leachate 135 wastewater sample was combined with prepared Fenton Solution, and then the sample was taken 136 into open atmosphere to conduct solar Photo Fenton experiment. The UV and visible light photo 137 Fenton processes were identical to the solar photo Fenton experiment, but artificial photo sources 138 were used. In place of sunlight, Wipro LED B22 Bulb Base with 50 lumen coverage and Super 139 395nm UV light (3.5 cm x 8.5 cm in size) were utilised in the UV and LED Visible Light Fenton 140 procedure. Using the following equation, the percentage elimination of the response variables 141 Color, COD, TSS, Cr, Cd, and Cu was computed (3). 142

Prepared Fenton Solution was mixed with leachate wastewater sample, then the sample was taken out to open place to triggering the oxidation process. UV and visible light photo Fenton process were same as solar photo Fenton process, but UV radiation was artificially used. Super 395nm UV light (size of UV lamp – $3.5 \text{ cm} \times 8.5 \text{ cm}$) and Wipro LED B22 Bulb Base with 50 lumen coverage light sources were used in UV and LED Visible Light Fenton process as a replacement for sun light. The percentage removal of response variables Color, COD, TSS, Cr, Cd and Cu was calculated using the following Eq.(3).

150

 $R = (Initial Value of response variable - Final value of response variable) \times 100$ (3)

151

(Initial value response variable)

152 Where,

153 R(%) - Removal efficiency in percentage

154 2.2.2. Taguchi experimental design

The experiments were examined and optimized by Taguchi experimental design method. It was an orthogonal array design. It is an efficient method when compared to other design methodologies since it enables desired parameter optimization to be less than or equal to 90%. In this present study three different type of photo Fenton process (Solar, UV, and Visible light) were individually analyzed by Taguchi method. Control factors such as pH, dosage, Time and their levels were fixed in this Taguchi experimental design. It was furnished in Table 1. From the findings of the preliminary one-time classical investigation, each level of the control parameters was fixed.

162 2.2.3. Signal-to-Noise Ratio S/N

163 The S/N ratio clearly identifies uncontrollable variables or undesirable noise in the experimental 164 input and output parameters. According to Minitab's S/N ratios there are three types of 165 performance characteristics: Larger the better, smaller the better and nominal the better. To 166 evaluate the experimental design for the treatment of leachate for all three experimental167 procedures, larger the better choice was selected from Eq(4).

168
$$S/N = -10 x \log (\Sigma (1/Y^2)/n)$$
 (4)

169 Where Y = responses for the given factor level combination

n = number of responses in the factor level combination.

Analysis of variance (ANOVA), which is used to confirm the significance of the model, describes 172 the importance of the experimental modal and provides extensive information on the most 173 influencing factor for this Photo Fenton process. The findings of an ANOVA are widely applied to 174 analyse the interaction behaviour among the control variables and responses. Each factor underwent 175 a qualitative significance analysis using the F and P values. The design is considered significant if 176 the probability value (p value) is less than or equal to 0.05. F value greater than Fcr value for each 177 factor confirm that the adequacy of the modal. The total degrees of freedom (DF), Adjusted mean 178 squares (adj MS) Adjusted sums of squares (adj SS), S value and R² value also can obtained from 179 ANOVA; it is easy to predict the significance of the experimental design. 180

181 **3. Result and Discussion**

182 *3.1. Physio-chemical Characteristics of matured landfill leachate wastewater*

The obtained matured landfill wastewater samples had the following physio-chemical properties as follows- pH 8.5, dark brown color with high colloidal nature, foul odour, 562 mg/L of BOD₅, 960 NTU of turbidity, 2521 mg/L of COD, 5513 mg/L of Total solids, 4327 mg/L of Total dissolved solid , 1214 mg/L of Total suspended solids, the heavy metal concentration was 1.198 mg/L of chromium, 0.2 mg/L of cadmium, 0.77 mg/L of copper. 188 3.2 Comparative analysis of Photo Fenton oxidation process: Solar, UV and LED visible light
189 photo

In this present investigation, the control factor levels of the pH, Fenton Dosage, and reaction time 190 were chosen as the same values for solar, UV, and LED visible light photo Fenton experiment. All 191 192 the three photo Fenton experiments showed the best pollutant reduction efficiency in pH 3. Fenton reaction was often quite active at pH 3 or lower. Solution in an acidic environment (2-4) promotes 193 the formation of more OH⁻ ions, which results in a high rate of oxidation (Pignatello et al., 2006; 194 Silva et al., 2015 and Kavitha et al., 2005). From the study, it was observed that Fenton process 195 heavily depends on the rate of OH ions scavenged because when pH values greater than 3 196 demonstrated less pollutant degradation potential. Higher pH causes the reaction between ferric ions 197 and H₂O₂ to produce more ferric hydroxo species, which reduces the effectiveness of pollutant 198 degradation (Karale et al., 2014; Ipek Gulkaya et al., 2006 and Badawy et al., 2006). The pH is a 199 200 significant factor in molar fraction of iron -inorganic, iron-organic and iron-water species (Silva,et al., 2015), so the effectiveness of pollutant removal was greatly influenced by pH. 201

In this solar, UV, and LED visible light Fenton investigation, the dosage concentration of FeSO4 202 was adjusted from 1 to 2.5 g/L while 30 g/L of H₂O₂ dosage remained constant. Pollutant 203 degradation effectiveness was excellent up to a dosage of 2 g/L, but as the dosage increased, iron 204 precipitation formed, which slowed down the reaction. Higher concentration of FeSO₄ dosage 205 induced the self - hindrance of OH[.] radical which reduce the pollutant degradation efficiency, OH[.] 206 + $Fe^{2+} \rightarrow Fe^{3+} + OH^-$ (Du and Qiu 2013), because both H₂O₂ and Fe²⁺ can serve as radical 207 scavengers. Several researches believed that the ideal dosage of these two compounds was crucial 208 to achieving high pollutant degradation efficiency (Tang and Tassos, 1997; Kochany and 209 Lugowski, 1998). In the photo Fenton experiment H₂O₂ initiate superior effect in oxidizing organic 210 complexes and producing of more oxygen ions in the presence of Fe²⁺ ion (Panizza and Cerisola, 211 2001). Fenton experiment is also activated at lower Fe^{2+} ion concentrations, but the removal is 212 inefficient. The photo Fenton method is more preferable than classical Fenton method since it 213

results in less iron sludge formation by decreasing the dosage of catalyst and Solar sources (Solar or
UV) increase the utilization of H₂O₂ (Bandala et al., 2009; Giri and Golder, 2014 and Xiao et al.,
2014).

There were four different time intervals used for the solar, UV, and LED visible light photo Fenton 217 processes: 30, 60, 90, and 120 minutes. Maximum results for the Solar Fenton procedure at 60 218 minutes of reaction time were 95% of color, 83% of COD, and 89% of TSS. At 120 minutes of 219 reaction time, the UV and LED visible Fenton processes demonstrated maximum of color 93% and 220 90%, COD 79% and 81% and TSS 87% and 84% removal efficiencies respectively. In comparison 221 to UV and LED visible Fenton processes, the solar Fenton process demonstrated good removal 222 efficiency in a shorter reaction time. According to (Jyoti Katara and Reshma Patel 2018), the solar 223 Fenton process outperformed the UV Fenton method in terms of pollutant degrading efficiency. 224 Temperature and UV radiation intensity may be to causes for the pollutant's fast deterioration in the 225 solar Fenton reaction. The study was done in the morning at 10:30 am - 12.30 pm, when the outside 226 temperature varied from 34 to 36°C and the inside room temperature was between 31 and 33°C. 227 High temperature boosts the effectiveness of pollutant degradation by increasing OH' production 228 and H₂O₂ consumption (Zazo et al., 2011). The production of more hydroxyl radicals (OH[•]) by the 229 ferrous ions and hydrogen peroxide (H₂O₂) combination in the treatment method, high COD 230 removal was achieved (Lucas et al., 2007; Gallard and De Laat 2000). During the Fenton process, 231 hydroxyl radicals convert all the minerals in wastewater into CO₂, water, and inorganic compounds. 232 The study was carried out on Fenton process for landfill leachate using two standard high-pressure 233 234 mercury-vapor immersion lamps (100 W and 450 W) and a specially made 8 W 365 nm UVA-LED lamp (Tejera et al., 2021). With an 8 W 365 nm UVA-LED lamp, COD could be removed 90%, 235 whereas high pressure mercury vapour immersion lamps demonstrated poor COD removal. UV 236 237 radiation improves the usage of H₂O₂ concentration and speeds up the photolysis process on each small organic molecule, but the visible light photo Fenton process requires a long processing time 238 and higher energy consumption. Using 14 W UV lamps at 1.1 g/L Fe ion and 5.5 g/L H₂O₂ 239

concentration, 84.43% of COD and 92.54% of total PAHs were removed from landfill leachate 240 wastewater (Singa et al., 2018). The three treatment methods virtually always used the ideal Fenton 241 dose concentration. Optimum Fenton dosage concentration was almost common in the all three 242 243 treatment method. In comparison to UV radiation from artificial sources, solar sources was higher in pollutant degradation because more active photons from high light wave length of solar induced 244 high reaction rate in comparison to the other two artificial photo sources. In the current 245 investigation, it was found that pH and UV radiation both had a significant impact on the pollutant 246 degradation in landfill leachate wastewater in all three Fenton processes. In comparison to the 247

standard Fenton method, the solar photo Fenton procedure demonstrated greater pollutant degrading
efficiency (Vilar et al., 2012).

250 *3.3. Statistical Experimental Result*

In Taguchi experimental optimization techniques, OA L16 (4^{3}) was used to optimize the process 251 variables using minitab statistical software (Version 21.1.0). pH, Dosage concentration and Time 252 were taken as a control variables and color, COD, TSS, Cr, Cu and Cd were taken as a responses. 253 All the experimental values were analyzed in three times and average values noted. Experimental 254 results of solar, UV and LED visible light photo Fenton process corresponding to L16 design were 255 presented in Table 2, Table 3 and Table 4 respectively. Totally 16 number of runs were performed 256 for each solar, UV and LED visible light Photo Fenton method. Solar Fenton process achieved 257 maximum amount of color 95%, COD 83%, TSS 89%, Cr 94%, Cd 86% and Cu 93% at control 258 factors level 3 pH (pH 3), level 2 dosage (Dosage 1.5:30 g/L) and level 2 (60 minutes) reaction 259 time. From Table 3 and Table 4 it was observed that UV and Visible light photo Fenton method 260 needed higher reaction time to show the better removal efficiency. Photo Fenton mechanism highly 261 depends on the emission frequency of the light source, availability and absorption rate of 262 photoactive species (Pignatello et al., 2006). 263

Figure 1 illustrates the S/N ratio, which describes how each control factors affects each response. The control variables pH, dose concentration (Fe^{2+}/H_2O_2), and reaction time were plotted in the S/N ratio graph's X axis, and each response (Color, COD, TSS, Cr, Cd, and Cu) was plotted in the graph's Y axis. S/N ratio graph clearly demonstrate that pH played a major role in the elimination of color, COD, TSS, Cr, Cd, and Cu in all three of the treatment processes—solar, UV, and LED.

Tables 5 and 6 display the ANOVA results for the solar, UV, and LED visible light photo Fenton 269 processes. For all three processes (Solar, UV, and LED Fenton process), almost all of the p values 270 were less than 0.05, and larger F values indicate that the designs were significant. In the ANOVA 271 results, pH had a very high SN value, demonstrating that pH was an important factor in the 272 breakdown of pollutants throughout all treatment stages. Table 7 gives the detailed summary report 273 of Taguchi experimental design results of these three types of Fenton process. From Table.7, it was 274 observed that R value more than 99% for all the control variables for all the three methods, it is 275 evident that the good correlation between the each responses and each operating variables. Almost 276 all the methods achieved more 97% of adj R² value; it also confirms the significance of this design. 277 Less residual error (less than 50%) in total run indicates the results to be reliable (Pourjafar et al., 278 2013 and Reyhani, et al., 2015). F value should be greater than Fcr value otherwise the design was 279 not significant (Sousa, et al., 2020). In this robust design DF value for each factor is 3 a residual 280 error is 6 so the confidence level is more than 95% for all the factors in these three methods. Fcr < 281 F, it confirms that this design is significant. 282

Figures 2 and 3 illustrated interaction effect of pH and dosage concentration on color and Cr removal in 2D Contour plots. The well-conditioned elliptical form is strong indication that the pH and dose concentration for the elimination of color and Cr interact more. Solar Fenton technology removed up to 95% of the color and 94% of the Cr, UV Fenton technology removed up to 93% of both the color and Cr, and LED visible light technology removed up to 90% of the color and 91% of the Cr. In this study the heavy metals concentration of before and after the treatment was analysed with AAS. The Fenton process was a suitable choice for removing phenol, cyanide, and Cr (VI).

Compared to TiO2/UV, the H2O2/UV method demonstrated a quicker decrease of chromium 290 (Golbaz et.al., 2013). Municipal wastewater can be successfully treated with the solar photo Fenton 291 technique to remove heavy metals (Cr 92%, Pb 100%, Cu 72.4%, Cd 100%, Ni 36%, Fe 94%, Zn 292 293 58%) (Chaudhary et.al., 2012 and Barwal et.al., 2015).

Figure.4 a, b and c furnished the probability plot of removal efficiency of Solar, UV and LED 294 visible light Fenton process. It is confirmed that all of the control variables and responses had a 295 positive interaction impact on one another because all of the anticipated and observed values are 296 followed in a straight line and the 95% confidence limit is displayed. Solar photo Fenton 297 technology has been described as being both affordable and appropriate for complex industrial 298 wastewater (Malato et al., 2013). 299 20

4. Conclusion 300

According to the findings of this investigation, with one hour of reaction time at an acidic pH, the 301 solar-assisted Fenton process successfully removed color, COD, TSS, Cr, Cu, and Cd from landfill 302 leachate wastewater, whereas the other two methods required longer reaction times. Using the 303 Taguchi technique, the most accurate data and relationships between controls and responses may be 304 gathered and analysed with the fewest number of experimental designs. The findings of this 305 experiment demonstrate that generation and rate of scavenging of hydroxyl radicals in the process 306 highly depends on pH and UV radiation. The three techniques' R² values were greater than 99%, 307 which demonstrated the model's good suitability. The solar, UV, and LED visible light Fenton 308 processes are all effective at removing pollutants from landfill leachate wastewater while generating 309 a less amount of sludge as compared with other Physico- chemical methods. However, solar photo 310 Fenton process has higher removal efficiency and uses a naturally occurring energy source in 311 comparison to the other two artificial photo sources. Organic pollutant and heavy metal reduction in 312 landfill leachate wastewater can be accomplished using the competitive and promising solar photo 313 Fenton process. 314

Future directions

Economical contribution of these three processes has to be investigated. There have also been other studies done on the efficiency of micro pollutant degradation. Investigating various UV sources and wave lengths may improve the efficiency of pollutant degradation.

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- 320

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325 **Conflict of Interest**

326 The authors declare that there are no competing interests associated with this research work.

327 Author contributions

First author collected wastewater samples, conducted all the experimental studies, analysed theresults and prepared the manuscript.

330 Second author analysed the result finding, validated the results and reviewed the manuscript

331 References

Abdelhaleem, A., & Chu, W. (2020). Prediction of carbofuran degradation based on the 332 radical's generation using the FeIII impregnated dopedhydroxyl Ν 333 TiO2/H2O2/visible LED photo-Fenton-like process. Chemical Engineering 334 Journal, 382, 122930. 335

336	Amor, C., De Torres-Socías, E., Peres, J. A., Maldonado, M. I., Oller, I., Malato, S., & Lucas,
337	M. S. (2015). Mature landfill leachate treatment by coagulation/flocculation combined
338	with Fenton and solar photo-Fenton processes. Journal of Hazardous Materials, 286,
339	261-268.
340	APHA, Standard Methods for the examination of water and wastewater,23nd ed.,American
341	Public Health Association (APHA), New york, 2017
342	Avsar, Y., Kurt, U., & Gonullu, T.(2007). Comparison of classical chemical and
343	electrochemical processes for treating rose processing wastewater. Journal of
344	hazardous materials, 148(1-2), 340-345.
345	Badawy MI, Ghaly M, Gad-Allah T (2006) Advanced oxidation processes for the removal of
346	organophosphorus pesticides from wastewater. Desalination 194:166-175.
347	doi:10.1016/j.desal. 2005.09.027
348	Bandala ER, Brito L, Pelaez M. Degradation of domoic acid toxin by UV-promoted Fenton-
349	like processes in seaeater. Desalination. 2009;45:135-145
350	Barwal, A., & Chaudhary, R. (2015). Optimization of operational parameters in moving bed
351	biofilm reactor with low cost polystyrene biocarrier by the response surface
352	method. Water Quality Research Journal, 52(1), 26-41.
353	Bautitz, I. R., & Nogueira, R. F. P. (2007). Degradation of tetracycline by photo-Fenton
354	process-Solar irradiation and matrix effects. Journal of Photochemistry and
355	Photobiology A: Chemistry, 187(1), 33-39.
356	Chaudhary, R., & Thakur, R. S. (2012). Photocatalytic treatment of industrial wastewater
357	containing chromium as a model pollutant-effect on process parameters and
358	kinetically studies. Journal of Renewable and Sustainable Energy, 4(5), 053121.
359	Christensen, T. H., Kjeldsen, P., Bjerg, P. L., Jensen, D. L., Christensen, J. B., Baun, A., &
360	Heron, G. (2001). Biogeochemistry of landfill leachate plumes. Applied
361	geochemistry, 16(7-8), 659-718.

- Deng, Y., & Englehardt, J. D. (2008). Hydrogen peroxide-enhanced iron-mediated aeration
 for the treatment of mature landfill leachate. *Journal of hazardous materials*, *153*(1-2),
 293-299.
- 365 Dharmarathne, N., & Gunatilake, J. (2013). Leachate characterization and surface
 366 groundwater pollution at municipal solid waste landfill of Gohagoda, Sri
 367 Lanka. International Journal of Scientific and Research, 3(11), 1-7.
- Du, Y., & Qiu, M. (2013). Comparative study of advanced oxidation for textile
 wastewater. *Desalination and Water Treatment*, *51*(31-33), 5954-5958.
- Fraley, S., Oom, M., Terrien, B., & Date, J. Z. (2006). Design of experiments via Taguchi
 methods: orthogonal arrays. *The Michigan chemical process dynamic and controls open text book, USA, 2*(3), 4.
- Gallard, H., & De Laat, J. (2000). Kinetic modelling of Fe (III)/H2O2 oxidation reactions in
 dilute aqueous solution using atrazine as a model organic compound. *Water Research*, 34(12), 3107-3116.
- Ghaffariraad, M., & Ghanbarzadeh Lak, M. (2021). Landfill leachate treatment through
 coagulation-flocculation with lime and bio-sorption by walnut-shell. *Environmental Management*, 68(2), 226-239. <u>https://doi.org/10.1007/s00267-021-01489-4</u>
- Giri AS, Golder AK. Chloramphenicol degradation in Fenton and photo-Fenton: Formation of
 Fe²⁺-chloramphenicol chelate and reaction pathways. Industrial and Engineering
 Chemistry Research. 2014;53:16196-16203
- 382 Giusti, L. (2009). A review of waste management practices and their impact on human
 383 health. *Waste management*, 29(8), 2227-2239.
- Gogate, P. R., & Pandit, A. B. (2004). A review of imperative technologies for wastewater
 treatment I: oxidation technologies at ambient conditions. *Advances in environmental research*, 8(3-4), 501-551.

387	Golbaz, S., Jafari, A. J., & Kalantari, R. R. (2013). The study of Fenton oxidation process
388	efficiency in the simultaneous removal of phenol, cyanide, and chromium (VI) from
389	synthetic wastewater. Desalination and Water Treatment, 51(28-30), 5761-5767.
390	Gulkaya, I., Surucu, G. A., & Dilek, F. B. (2006). Importance of H2O2/Fe2+ ratio in Fenton's
391	treatment of a carpet dyeing wastewater. Journal of Hazardous Materials, 136(3),
392	763-769.
393	Hartmann, M., Kullmann, S., & Keller, H. (2010). Wastewater treatment with heterogeneous
394	Fenton-type catalysts based on porous materials. Journal of Materials
395	<i>Chemistry</i> , 20(41), 9002-9017.
396	Jyoti Katara and Reshma L. Patel(2018)" Treatment of textile wastewater by photo fenton
397	process with UV light and solar radiation", International Journal of Creative Research
398	Thoughts, Volume 6,
399	Karale, R. S., Manu, B., & Shrihari, S. (2014). Fenton and photo-Fenton oxidation processes
400	for degradation of 3-aminopyridine from water. APCBEE procedia, 9, 25-29.
401	Kavitha, V., & Palanivelu, K. (2005). Degradation of nitrophenols by Fenton and photo-
402	Fenton processes. Journal of Photochemistry and Photobiology A: Chemistry, 170(1),
403	83-95.
404	Klein, K., Kivi, A., Dulova, N., Zekker, I., Mölder, E., Tenno, T & Tenno, T. (2017). A pilot
405	study of three-stage biological-chemical treatment of landfill leachate applying
406	continuous ferric sludge reuse in Fenton-like process. Clean Technologies and
407	Environmental Policy, 19(2), 541-551.
408	Kochany, J., & Lugowski, A. (1998). Application of Fenton's reagent and activated carbon for
409	removal of nitrification inhibitors. Environmental technology, 19(4), 425-429
410	Li, G., Chen, J., Yan, W., & Sang, N. (2017). A comparison of the toxicity of landfill leachate
411	exposure at the seed soaking and germination stages on Zea mays L.(maize). Journal
412	of Environmental Sciences, 55, 206-213 https://doi.org/10.1016/j.jes.2016.06.031.

- Lucas, M. S., Dias, A. A., Sampaio, A., Amaral, C., & Peres, J. A. (2007). Degradation of a
 textile reactive Azo dye by a combined chemical–biological process: Fenton's reagentyeast. *Water research*, 41(5), 1103-1109.
- Ma, Dengsheng, Huan Yi, Cui Lai, Xigui Liu, Xiuqin Huo, Ziwen An, Ling Li et al. "Critical
 review of advanced oxidation processes in organic wastewater
 treatment." *Chemosphere* 275 (2021): 130104.
- 419 Malato, S., Fernández-Ibánez, P., Maldonado, M. I., Oller, I., Polo-López, M. I., & Pichat, P.
- 420 (2013). Solar photocatalytic pilot plants: commercially available
 421 reactors. *Photocatalysis and water purification: From fundamentals to recent*422 applications.
- Mohan, S., & Gokul, D. (2022). Treatment of leachate from open dumpsite of municipal solid
 waste by ozone based advanced oxidation process. *Ozone: Science* & *Engineering*, 44(3), 250-264.
- Niza, N. M., Yusoff, M. S., Zainuri, M. A. A. M., Emmanuel, M. I., Shadi, A. M. H., Hanif,
 M. H. M., & Kamaruddin, M. A. (2021). Removal of ammoniacal nitrogen from old
 leachate using batch electrocoagulation with vibration-induced electrode
 plate. *Journal of Environmental Chemical Engineering*, 9(2), 105064.
- Okagbue, H. I., Oguntunde, P. E., Obasi, E. C., & Akhmetshin, E. M. (2021). Trends and
 usage pattern of SPSS and Minitab Software in Scientific research. In *Journal of Physics: Conference Series* (Vol. 1734, No. 1, p. 012017). IOP Publishing.
- Panizza, M., & Cerisola, G. (2001). Removal of organic pollutants from industrial wastewater
 by electrogenerated Fenton's reagent. *Water research*, *35*(16), 3987-3992.
- Pignatello, J. J., Oliveros, E., & MacKay, A. (2006). Advanced oxidation processes for
 organic contaminant destruction based on the Fenton reaction and related
 chemistry. *Critical reviews in environmental science and technology*, *36*(1), 1-84.

- Poblete, R., & Pérez, N. (2020). Use of sawdust as pretreatment of photo-Fenton process in
 the depuration of landfill leachate. *Journal of environmental management*, 253,
 109697.
- 441 Pourjafar, S., Jahanshahi, M., & Rahimpour, A. (2013). Optimization of TiO2 modified poly
 442 (vinyl alcohol) thin film composite nanofiltration membranes using Taguchi
 443 method. *Desalination*, *315*, 107-114.
- Primo, O., Rivero, M. J., & Ortiz, I. (2008). Photo-Fenton process as an efficient alternative
 to the treatment of landfill leachates. *Journal of hazardous materials*, *153*(1-2), 834842.
- Primo, O., Rivero, M. J., Ortiz, I., & Irabien, A. (2007). Mathematical modelling of phenol
 photooxidation: Kinetics of the process toxicity. *Chemical Engineering Journal*, 134(1-3), 23-28.
- Reyhani, A., Sepehrinia, K., Seyed Shahabadi, S. M., Rekabdar, F., & Gheshlaghi, A. (2015).
 Optimization of operating conditions in ultrafiltration process for produced water
 treatment via Taguchi methodology. *Desalination and Water Treatment*, 54(10), 26692680.
- 454 Rocha, E. M., Vilar, V. J., Fonseca, A., Saraiva, I., & Boaventura, R. A. (2011). Landfill
 455 leachate treatment by solar-driven AOPs. *Solar Energy*, *85*(1), 46-56.
- 456 Sajjadi, S. A., Afsharnia, M., Azrah, K., Javan, N. S., & Biglari, H. (2015). Humic Acid
 457 Degradation via Solar Photo-Fenton Process in Aqueous Environment. *Iranian*458 *Journal of Health, Safety and Environment, 2*(3), 304-312.
- Silva, T. F., Ferreira, R., Soares, P. A., Manenti, D. R., Fonseca, A., Saraiva, I.& Vilar, V. J.
 (2015). Insights into solar photo-Fenton reaction parameters in the oxidation of a
 sanitary landfill leachate at lab-scale. *Journal of environmentalmanagement*, *164*,3240 http://dx.doi.org/10.1016/j.jenvman.2015.08.030.

463	Singa, P. K., Isa, M. H., Ho, Y. C., & Lim, J. W. (2018). Treatment of hazardous waste
464	landfill leachate using Fenton oxidation process. In E3S Web of conferences (Vol. 34,
465	p. 02034). EDP Sciences.
466	Sousa, M. R. S., Lora-García, J., López-Pérez, M. F., Santafé-Moros, A., & Gozálvez-
467	Zafrilla, J. M. (2020). Operating conditions optimization via the Taguchi method to
468	remove colloidal substances from recycled paper and cardboard
469	productionwastewater. Membranes, 10(8),170. <u>https://doi.org/10.3390/membranes1207</u>
470	<u>0698</u>
471	Tang, W. Z., & Tassos, S. (1997). Oxidation kinetics and mechanisms of trihalomethanes by
472	Fenton's reagent. Water Research, 31(5), 1117-1125.
473	Tejera, J., Gascó, A., Hermosilla, D., Alonso-Gomez, V., Negro, C., & Blanco, Á. (2021).
474	Uva-led technology's treatment efficiency and cost in a competitive trial applied to the
475	photo-fenton treatment of landfill leachate. Processes, 9(6), 1026.
476	Vilar, V. J., Moreira, J. M., Fonseca, A., Saraiva, I., & Boaventura, R. A. (2012). Application
477	of Fenton and solar photo-Fenton processes to the treatment of a sanitary landfill
478	leachate in a pilot plant with CPCs. Journal of Advanced Oxidation
479	<i>Technologies</i> , <i>15</i> (1), 107-116.
480	Vorontsov, A. V. (2019). Advancing Fenton and photo-Fenton water treatment through the
481	catalyst design. Journal of Hazardous Materials, 372, 103-112.
482	https://doi.org/10.1016/j.jhazmat.2018.04.033
483	Xiao D, Guo Y, Lou X, Fang C, Wang Z, Liu J. Distinct effects of oxalate versus malonate on
484	the iron redox chemistry: Implications for the photo-Fenton reaction. Chemosphere.
485	2014;103:354-358
486	Xu, M., Wu, C., & Zhou, Y. (2020). Advancements in the Fenton process for wastewater
487	treatment. Adv. Oxid. Process, 61.

- Zazo, J. A., Pliego, G., Blasco, S., Casas, J. A., & Rodriguez, J. J. (2011). Intensification of
 the Fenton process by increasing the temperature. *Industrial & Engineering Chemistry Research*, 50(2), 866-870.
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494 Figure 1. S/N ratio Plot for color removal with each control factors for Solar (a), UV (b) and LED
495 visible light (c)



Figure 2. 2D Contour plot for removal of color by using Solar (1a), UV (1b) and LED visible light

498 photo Fenton (1c)





Figure 3. 2D Contour plot for removal of Cr by Solar (2a), UV (2b) and LED visible light photo
Fenton (2c)



Figure 4. Probability plot for removal Efficiency of Solar (a), UV (b) and LED visible light (c)



 Table 1. Operating factors and their levels

Factors	Level 1	Level 2	Level 3	Level 4	
рН	2	2.5	3	3.5	
Dosage(g/L)	1:30	1.5:30	2:30	2.5:30	
$(FeSO_4 : H_2O_2)$)				
Time (min)	30	60	90	120	



Table 2. Experimental results corresponding to L 16 design of solar Photo Fenton process

Run	Run pH	Dosage	Time	Color	Color COD	Total	Cr	Cd	Cu
	(Level)	(Level)	(Level)	(%)	(%)	Suspended	(%)	(%)	(%)
						Solids			
						(%)			
1	1	1	1	22	13	19	35	30	35
2	1	2	2	27	15	25	37	32	37
3	1	3	3	25	13	21	35	29	33
4	1	4	4	23	10	18	31	27	30
5	2	1	2	60	45	49	58	57	73
6	2	2	1	48	37	43	55	54	69
7	2	3	4	53	40	45	50	49	70
8	2	4	3	46	39	43	53	55	68
9	3	1	3	85	76	82	92	84	85
10	3	2	4	87	79	87	90	82	83
11	3	3	1	72	70	86	90	80	88
12	3	2	2	95	83	89	94	86	93
13	4	1	4	75	65	67	86	70	68
14	4	2	3	73	63	69	88	73	70
15	4	3	2	78	68	69	88	74	74

Table 3. Experimental results corresponding to L 16 design of UV Photo Fenton process

Run	рН	Dosage	Time	Color	COD	Total	Cr	Cd	Cu
	(Level)	(Level)	(Level)	(%)	(%)	Suspended	(%)	(%)	(%)
						Solids			
						(%)			
1	1	1	1	19	11	16	30	28	33
2	1	2	2	25	14	23	32	30	33
3	1	3	3	21	14	19	37	33	38
4	1	4	4	20	10	15	39	34	40
5	2	1	2	46	39	39	51	51	64
6	2	2	1	58	43	47	53	55	71
7	2	3	4	53	40	45	57	61	73
8	2	4	3	43	37	43	61	65	79
9	3	1	3	82	76	80	90	77	81
10	3	2	4	93	81	87	93	82	91
11	3	3	1	88	75	86	93	80	87
12	3	4	2	79	70	78	88	77	85
13	4	1	4	69	62	65	84	71	77
14	4	2	3	73	65	68	88	74	76
15	4	3	2	65	60	63	82	71	70
16	4	4	1	61	59	60	77	66	65

Table 4. Experimental results corresponding to L 16 design of LED Visible light Photo Fenton

Run	рН	Dosage	Time	Color	COD	Total	Cr	Cd	Cu
	(level)	(level)	(level)	(%)	(%)	Suspended	(%)	(%)	(%)
						Solids		\cdot	
						(%)	1		
1	1	1	1	17	11	14	28	26	30
2	1	2	2	24	12	21	29	28	31
3	1	3	3	19	12	17	35	30	37
4	1	4	4	18	8	13	38	32	38
5	2	1	2	44	37	36	48	49	62
6	2	2	1	56	41	45	50	53	68
7	2	3	4	51	38	42	57	51	55
8	2	4	3	48	35	40	54	49	51
9	3	1	3	80	74	78	88	75	79
10	3	2	4	90	79	84	91	80	87
11	3	3	1	86	75	83	90	78	85
12	3	4	2	80	73	79	87	75	82
13	4	1	4	65	61	67	81	69	74
14	4	2	3	68	64	70	86	72	77
15	4	3	2	61	59	65	83	68	70
16	4	4	1	59	54	61	74	64	66

548 Table 5. ANOVA for SN ratios of Color removal by Solar, UV, and LED Visible light Photo549 Fenton process

Source	DF	Seq SS	Adj SS Adj MS		F	Р			
Analysis of Variance for SN ratios of Color removal by solar photo Fenton									
pН	3	835.506	835.506	311.835	460.7	0.000			
Dosage	3	7.704	7.704	2.568	3.79	0.039			
Time	3	2.831	2.831	0.944	1.39	0.033			
Residual Error	6	6.061	6.061	0.677					
Total	15	852.102							
Analysis of Varia	nce for SN ra	tios of Color rem	oval by UV	photo Fentor	<u>n</u>				
		0							
pН	3	345.799	345.799	181.933	207.2	0.000			
Dosage	3	9.459	9.459	3.153	3.59	0.036			
Time	3	1.701	1.701	0.567	0.65	0.013			
Residual Error	6	7.267	7.267	0.878					
Total	15	364.226							

Analysis of Vari	ance for SN	ratios of Color r	emoval by LEI) Visible Lig	nt photo Fen	<u>ton</u>
nH	3	307 833	307 833	69 2777	1303 20	0.000
Dosage	3	3.877	3.877	1.2924	24.31	0.000
Time	3	1.368	1.368	0.4559	8.58	0.014
Residual Error	6	0.319	0.319	0.0532		
Total	15	313.397		(

Table 6. ANOVA for SN ratios of Cr removal by Solar, UV, and LED Visible light Photo Fenton

Source	DF	Seq SS	Adj SS	Adj MS	F	Р					
			0								
Analysis of Varia	Analysis of Variance for SN ratios of Cr removal by Solar Photo Fenton										
pH	3	338.724	338.724	79.5745	834.96	0.000					
Dosage	3	4.135	4.135	1.3783	14.46	0.004					
Time	3	1.773	1.773	0.5909	6.20	0.029					
Residual Error	6	0.189	0.189	0.0630							
Total	15	344.821									
Analysis of Varia	ance for	SN ratios of C	r by UV photo	Fenton							
рН	3	407.833	407.833	69.2777	1303.20	0.000					
Dosage	3	3.877	3.877	1.2924	24.31	0.001					
Time	3	1.368	1.368	0.4559	8.58	0.014					
Residual Error	6	0.319	0.319	0.0532							
Total	15	413.397									
Analysis of Variand	ce for Sl	N ratios of Cr b	<u>y Visible light</u>	photo Fenton	<u>l</u>						
pH	3	216.536	216.536	58.8453	708.87	0.000					
Dosage	3	2.821	2.821	0.9403	11.33	0.007					

	Time	3	0.524	0.524	0.1747	2.10	0.021
	Residual Error	6	0.498	0.498	0.0830		
	Total	15	220.379				
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Table 7. Summary of contribution of each factor for Solar, UV and LED visible light Fenton
process.

Treatment Method	Solar			UV light			LED visible light		
Performance	S	R-sq	R-sq	S	R-sq	R-sq	S	R-sq	R-sq
Criteria		(%)	Adj		(%)	Adj		(%)	Adj
		S	(%)			(%)			(%)
Color	0.1652	99.90	99.69	0.9752	99.30	98.99	0.9752	99.30	98.99
COD	0.6854	99.29	98.67	0.4754	99.56	99.27	0.4754	99.56	99.27
TSS	0.3140	99.55	97.75	0.5240	99.45	99.07	0.5240	99.45	99.07
Cr	0.2511	99.92	99.60	0.2313	99.82	99.65	0.2313	99.82	99.65
Cd	0.7196	99.18	98.90	0.9186	99.00	98.47	0.9186	99.00	98.47
Cu	0.3464	99.80	99.00	0.3534	99.85	99.26	0.3534	99.85	99.26