

1 **Effect of restricted emissions during Covid-19 lockdown on air quality in Rabat – Morocco**

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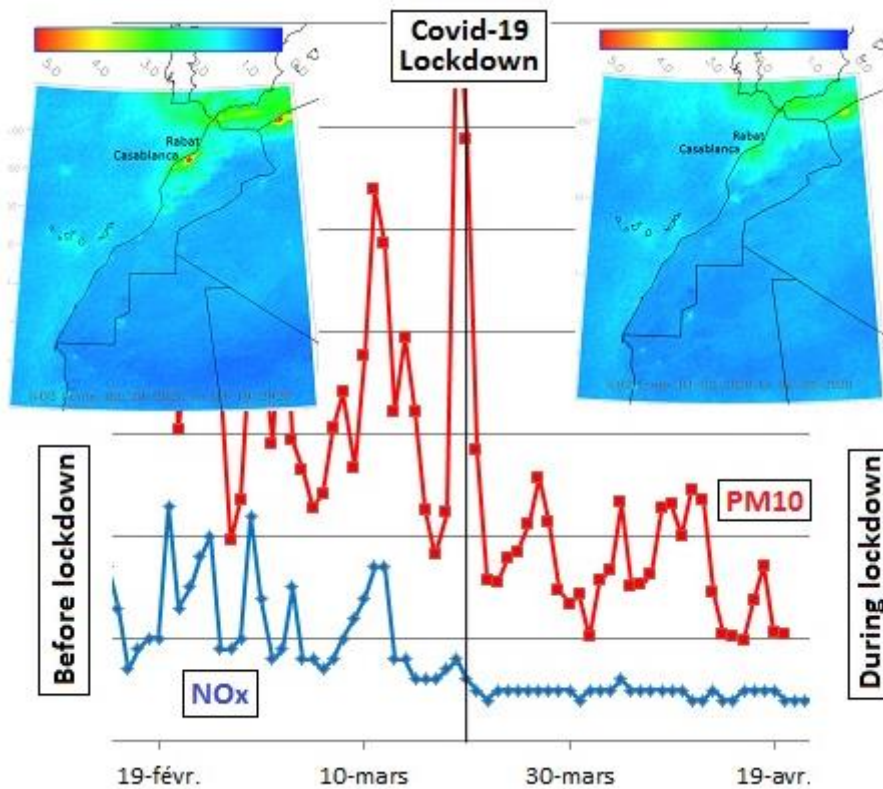
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25

26 **Abstract**

27 The effect of restricted human activities due to the COVID-19 pandemic in Rabat city (Morocco)  
 28 was studied, by analysing data of five criteria pollutants (NO, NO<sub>2</sub>, PM<sub>2.5</sub>, PM<sub>10</sub>, and O<sub>3</sub>) to  
 29 highlight the effectiveness of restricted emissions on air quality. Overall, drastic reductions on NO  
 30 (up to -94%), NO<sub>2</sub> (up to -51%), PM<sub>2.5</sub> (up to -45%), and PM<sub>10</sub> (up to -53%) concentrations were  
 31 observed during lockdown period compared to the first period (before lockdown). Spaceborne NO<sub>2</sub>  
 32 column observations from TROPOMI on board Sentinel-5P, reveal unprecedented NO<sub>2</sub> decreases  
 33 (up to -60%). By contrast, an increase of approximately 20% in ozone mean concentrations was  
 34 observed, probably related to the reduction of nitrogen oxides and also of fine particles (PM<sub>2.5</sub>)  
 35 which leads to less scavenging of HO<sub>2</sub> and as a result greater O<sub>3</sub> production.

36

37 **Keywords:** Coronavirus, Lockdown, Anthropogenic emissions, Air quality, NO<sub>x</sub>, PM, O<sub>3</sub>

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## 40 **1. Introduction**

41 There is significant evidence that air pollution is associated with premature mortality (Lelieveld et  
42 al., 2015) and adverse health effects. According to the World Health Organization, 4.6 million  
43 individuals die annually from diseases and illnesses directly related to poor air quality (Cohen et al.,  
44 2017). Particulate matter with an aerodynamic diameter lower than 2.5  $\mu\text{m}$  and 10  $\mu\text{m}$  (PM<sub>2.5</sub> and  
45 PM<sub>10</sub>), nitrogen oxides (NO<sub>x</sub>) and tropospheric ozone (O<sub>3</sub>) are one of the greatest environmental  
46 problems of developed countries, indeed they are considered among the most threatening air  
47 pollutants, in urban environments, in terms of harmful effects on human health associated with  
48 respiratory and cardiovascular diseases and mortality (Pascal et al., 2013; Cohen et al., 2017).

49 In late December 2019, an unknown infectious disease, later named COVID-19, was identified in  
50 Wuhan, China (Na Zhu et al., 2020). The disease has rapidly spread from Wuhan to other areas and  
51 affected 196 countries worldwide by March 25, 2020, which raised intense attention internationally.  
52 To contain the escalation of the transmission of the virus, many countries have adopted dramatic  
53 measures to reduce human interaction, prohibiting private and public gatherings, restricting all  
54 forms of transportation (both within a city and across cities).

55 As a possible side effect of this unprecedented lockdown, human mobility and relevant production  
56 and consumption activities have since decreased significantly. Transport sector, industrial and  
57 manufacturing sector were heavily affected. But on the other hand, lockdown restrictions are  
58 expected to largely modify anthropogenic emissions of pollutants, both in terms of emitted mass  
59 and time variations (Menut et al., 2020). Several studies on the impact of lockdown measures on air  
60 quality have recently been published. Sicard et al. (2020) studied the effect of lockdown on air  
61 pollution in four Southern European cities (Nice, Rome, Valencia and Turin) and Wuhan (China).  
62 Compared to the same period in 2017–2019, strong reductions in NO<sub>2</sub> mean concentrations were  
63 observed in all European cities, ~53% at urban stations, comparable to Wuhan (57%), reductions in  
64 PM<sub>2.5</sub> and PM<sub>10</sub> at urban stations were overall much smaller both in magnitude and relative change  
65 in Europe (~8%) than in Wuhan (~42%). By contrast, the daily O<sub>3</sub> mean concentrations increased at

66 urban stations by 24% in Nice, 14% in Rome, 27% in Turin, 2.4% in Valencia and 36% in Wuhan  
67 during the lockdown in 2020. The same findings have been reported by [Nakada et Urban \(2020\)](#) in  
68 São Paulo (Brazil), [Tobías et al. \(2020\)](#) in Barcelona (Spain), [Collivignarelli et al. \(2020\)](#) in Milan  
69 (Italy), [Sharma et al. \(2020\)](#) in 22 cities covering different regions in India.

70 On the other hand, [Bauwens et al. \(2020\)](#); [Tobías et al. \(2020\)](#) have used spaceborne NO<sub>2</sub> column  
71 data from two high-resolution instruments, Tropospheric Monitoring Instrument (TROPOMI) on  
72 board Sentinel-5 Precursor and Ozone Monitoring Instrument (OMI) on Aura. Exceptional  
73 decreases in NO<sub>2</sub> columns were observed over widespread areas in China, Europe, Latin America,  
74 and the United States.

75 Morocco was confirmed as one of the countries affected by COVID-19, when the first case was  
76 confirmed in Casablanca on March 2, 2020. The government reacted decisively to the threat of the  
77 pandemic and declared a “health state of emergency” that went into effect on March 20 and  
78 included a nationwide lockdown. In this study, we carry out a rigorous investigation into this issue  
79 and estimate how lockdown affected air quality of Rabat (Capital city of Morocco), by studying the  
80 evolution of the nitrogen oxides (NO and NO<sub>2</sub>), particulate matter (PM<sub>2.5</sub> and PM<sub>10</sub>), and ozone  
81 (O<sub>3</sub>) concentrations determined before lockdown (from February 20 to March 19) and during the  
82 lockdown (from March 20 to April 20).

## 83 **2. Materials and methods**

84 Rabat, capital city of Morocco, is located on the Atlantic coast, with an urban population of  
85 approximately 580,000 (2014) and a metropolitan population of over 1.2 million. Rabat is  
86 characterized by a Mediterranean climate with an annual mean air temperature of 15 °C (from 7 °C  
87 in January to 25 °C in August) and belongs to the sub-humid bioclimatic zone with an average  
88 annual precipitation of 560 mm.

89 During the studied period, the real-time monitoring data were provided by the automatic monitoring  
90 station of the Atmosphere Competence Center (PCMA) installed in the campus of the national  
91 higher school of mining of Rabat, using standard methods and equipment. Thermo scientific model

92 42i NO–NO<sub>2</sub>–NO<sub>x</sub> analyzers, which are based on the most common used technique  
93 (chemiluminescence) for nitrogen oxides measurements. Ozone Analyzer O342e from ENVEA,  
94 using LED-based UV photometry technology, for O<sub>3</sub> measurements. Particulate matter (PM<sub>2.5</sub>,  
95 PM<sub>10</sub>) was determined with PALAS FIDAS® 200E continuous particle monitor , using an optical  
96 particulate monitoring spectrometer.

97 The concentrations of NO, NO<sub>2</sub>, PM<sub>2.5</sub>, PM<sub>10</sub>, and O<sub>3</sub> were obtained at 15-minute intervals.  
98 Experimental data were initially analyzed to identify spurious data and values were organized in  
99 spread sheets as 1-hour means. Daily averages (24 h) have then calculated for the periods before  
100 (February 20th to March 19th) and during the lockdown (March 20th to April 20th), assessing the  
101 variation in the mean concentration ( $\mu\text{g m}^{-3}$ ) between both periods, and their relative change (%).

102 In parallel, remote sensing NO<sub>2</sub> data, obtained by the Copernicus Sentinel-5 Precursor Tropospheric  
103 Monitoring Instrument (S5p/TROPOMI), has been used to assess tropospheric NO<sub>2</sub> background  
104 levels in a high resolution ( $3.5 \times 7 \text{ km}^2$ ) continuous area (Ialongo et al., 2020).

105 In order to be able to interpret the evolution of pollutant concentrations before and during the  
106 confinement, meteorological parameters (temperature, relative humidity, and wind speed) were also  
107 determined. For this, the ERA5 data (<http://cds.climate.copernicus.eu/cdsapp#!/dataset/reanalysis-era5-single-levels?tab=form>)  
108 were collected for the 2 studied periods. These meteorological data are  
109 reanalyses available at a spatial resolution of 0.25° and a temporal resolution of 1 hour.

### 110 **3. Results and discussion**

#### 111 **3.1. Meteorological data analysis for reference period selection**

112 Meteorological data analysis is essential considering that weather phenomena have a massive  
113 influence on air quality. As shown in Table 1, average daily precipitation illustrates a similar  
114 situation between the lockdown and before lockdown periods, with low precipitation during the  
115 lockdown period. The recorded rainfall shows a total of about 2 mm d<sup>-1</sup>. Furthermore, there were  
116 negligible variations in temperature with a variation of less than 1°C between periods. The average

117 wind speeds over the periods are also very similar. Overall, it can be concluded that the conditions  
118 are similar for the two periods, thus preventing an overestimation of air quality.

119 **Table 1.** Averaged meteorological parameters observed in Rabat city before and during the  
120 lockdown.

	Temperature °C	Wind speed M s <sup>-1</sup>	Precipitation mm d <sup>-1</sup>
Mean BL	15,8	2,5	0,2
Mean DL	16,4	2,9	2,1
Difference	<b>0,6</b>	<b>0,4</b>	<b>1,9</b>

121 *BL : Before Lockdown*

122 *DL : During Lockdown*

### 123 3.2. Air quality monitoring

124 Experimental results were obtained from February 20 to April 20, 2020 at PCMA monitoring  
125 station. The obtained results are shown in [table 2](#) and presented in [Figs. 1-4](#), respectively, for  
126 nitrogen oxides (NO and NO<sub>2</sub>), particulate matter (PM<sub>2.5</sub> and PM<sub>10</sub>), and ozone (O<sub>3</sub>). The amplitude  
127 of the concentration variation (ACV) was calculated to assess relative change (%) comparing the  
128 lockdown period to before lockdown, using the equation;

$$129 \quad AVC = \frac{y - x}{x} \times 100\%$$

130 where x and y are the mass concentrations of a substance of interest before and during the control  
131 period for Covid-19, respectively.

#### 132 3.2.1. Nitrogen oxides

133 From the obtained results, as shown in [Fig. 1](#), exceptional decreases in nitrogen oxides levels were  
134 observed in the studied area. A decrease in the concentrations of NO (from 1.7 to 0.1 µg m<sup>-3</sup>) and  
135 NO<sub>2</sub> (from 8.6 to 4.2 µg m<sup>-3</sup>) was recorded, which corresponds to an AVC of -94% and -51%,  
136 respectively. This could be explained by the containment measures against the spread of the Covid-  
137 19, which caused sharp reductions in traffic and industrial activities. The magnitude of changes was  
138 similar to those recorded in European cities (~52%) and Wuhan in China (57%), where automobile  
139 exhausts are the major source of NO<sub>x</sub> ([Sicard et al. 2020](#); [Tobias et al., \(2020\)](#)).

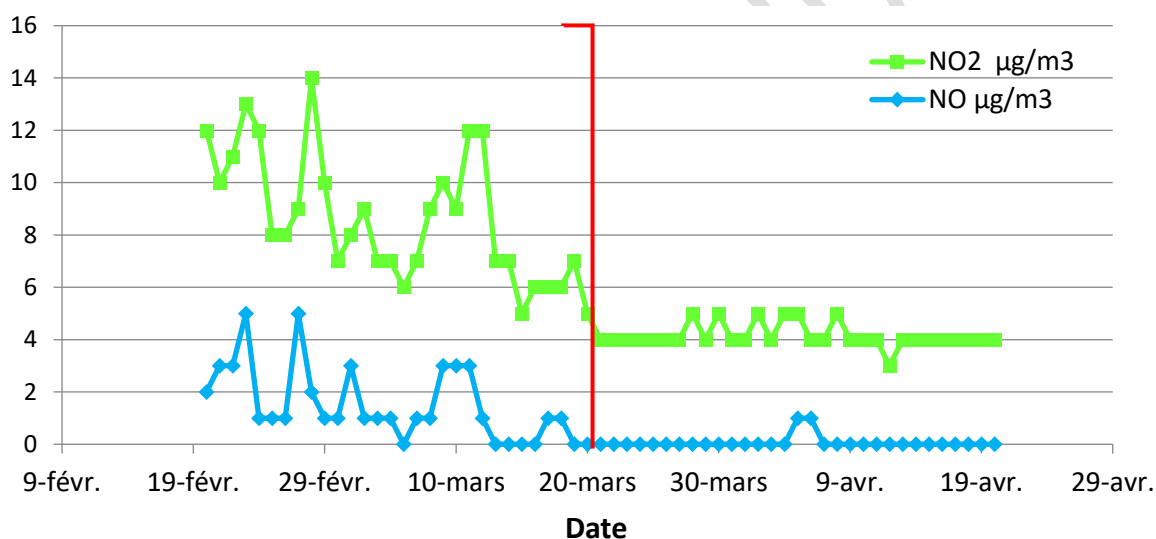
140 **Table 2.** Mean concentrations of NO, NO<sub>2</sub>, PM<sub>2.5</sub>, PM<sub>10</sub>, O<sub>3</sub>, measured in Rabat city before and  
141 during the lockdown.

	<b>NO</b> $\mu\text{g m}^{-3}$	<b>NO<sub>2</sub></b> $\mu\text{g m}^{-3}$	<b>PM<sub>2.5</sub></b> $\mu\text{g m}^{-3}$	<b>PM<sub>10</sub></b> $\mu\text{g m}^{-3}$	<b>O<sub>3</sub></b> $\mu\text{g m}^{-3}$
Mean BL	1,7	8,6	17,1	36,4	35,1
Mean DL	0,1	4,2	9,4	17,1	43,7
Variation in %	<b>-94</b>	<b>-51</b>	<b>-45</b>	<b>-53</b>	<b>+20</b>

142 *BL : Before Lockdown*

143 *DL : During Lockdown*

144 Nitrogen oxides ( $\text{NO}_x = \text{NO}_2 + \text{NO}$ ) are among the main drivers in air quality degradation in urban  
 145 centers, due to their role as catalysts of tropospheric ozone formation, and as precursors of  
 146 secondary inorganic aerosols, with consequences for climate and human health (Lelieveld et al.,  
 147 2015).  $\text{NO}_x$  emissions are mainly generated by combustion processes from anthropogenic pollution  
 148 sources including transportation, energy production and other industrial activities (Wang et al.,  
 149 2017; He et al., 2020a, 2020b).



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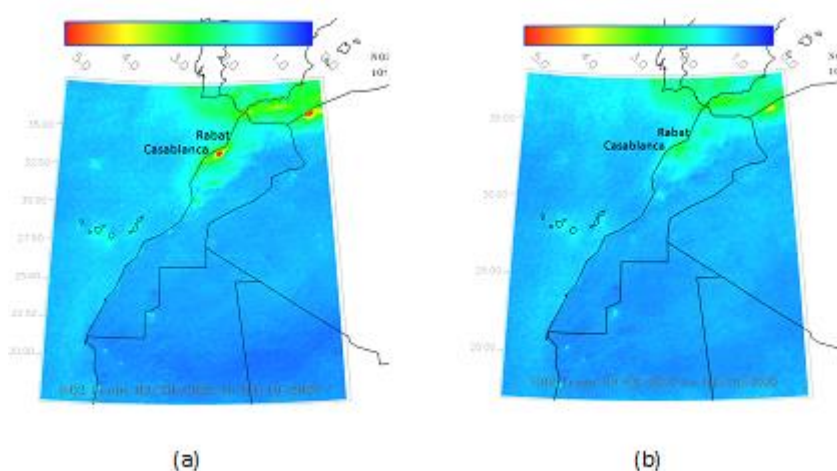
151 **Figure 1.** Daily NO and NO<sub>2</sub> mean concentrations ( $\mu\text{g m}^{-3}$ ) determined at PCMA

152 monitoring station from February 20, 2020 to April 20, 2020 in Rabat - Morocco

153 Daily NO<sub>x</sub> concentrations decrease is mainly attributed to the containment measures against the  
 154 spread of the Covid-19, which caused sharp reductions in road traffic and fossil combustion. Recent  
 155 researches have demonstrated that traffic emissions from heavy-duty diesel trucks are major sources  
 156 of nitrogen oxides (Nakada et Urban, 2020; Wang et al., 2020) and is considered to be responsible  
 157 for over half of NO<sub>x</sub> emissions and represents a higher proportion in urban areas.



158 On the other hand, NO<sub>2</sub> emissions were measured using TROPOMI instruments through  
159 Copernicus Sentinel-5P satellite. The satellite image (Fig. 2) provide comparison between the  
160 average of concentrations from 20 February to 19 March 2020 (panel a) and the average of  
161 concentrations from 20 March to 20 April 2020 (panel b). The NO<sub>2</sub> emissions reduced up to 60%  
162 due to lockdown, especially across the major cities such as Casablanca and Rabat, thus confirming  
163 the measures observed at PCMA monitoring station. Several studies have used satellite observations  
164 of NO<sub>2</sub> columns to assess the impact of covid-19 (Nakada et Urban 2020; Tobias et al. 2020;  
165 Zambrano-Monserrate et al., 2020).



166

167 **Figure 2.** Average TROPOMI NO<sub>2</sub> tropospheric columns over Morocco

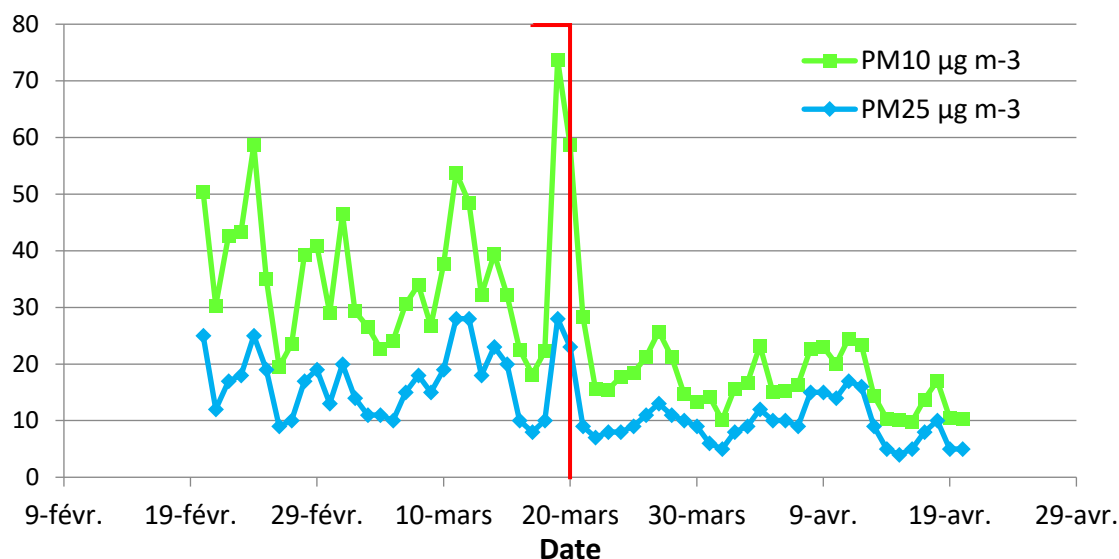
168 (E15 molecules/cm<sup>2</sup>), (a) before and (b) during the lockdown

### 169 3.2.2. Particulate matter

170 The particulate matter (PM) showed the same behavior (Fig. 3) as the nitrogen oxides during the  
171 Covid-19 control period compared with the reference period, in fact the mean concentrations  
172 decreased (from 17.1 to 9.4  $\mu\text{g m}^{-3}$ ) and from (36.4 to 17.1  $\mu\text{g m}^{-3}$ ), which corresponds to an AVC  
173 of  $-45\%$  and  $-53\%$ , respectively, for PM<sub>2.5</sub> and PM<sub>10</sub>. The higher decrease in PM<sub>10</sub> compared to  
174 PM<sub>2.5</sub> could be due to its greater contribution from anthropogenic sources (Klimont et al., 2017).  
175 The main source of PM<sub>10</sub> in the urban background is road traffic, combustion of fuels, industrial  
176 sources, construction works, dust resuspension, Saharan dust episodes (Querol et al., 2009), and  
177 regional air mass transport which could have an influence on PM to the point of reducing the effects



178 of local emission reduction. Aerosol particles  $PM_{2.5}$  are also receiving worldwide attention, as they  
179 have a potential greater adverse effect on the health of human beings, especially on that part of the  
180 population living in urban areas influenced by high traffic density or industry (Dongarrà et al.,  
181 2010).



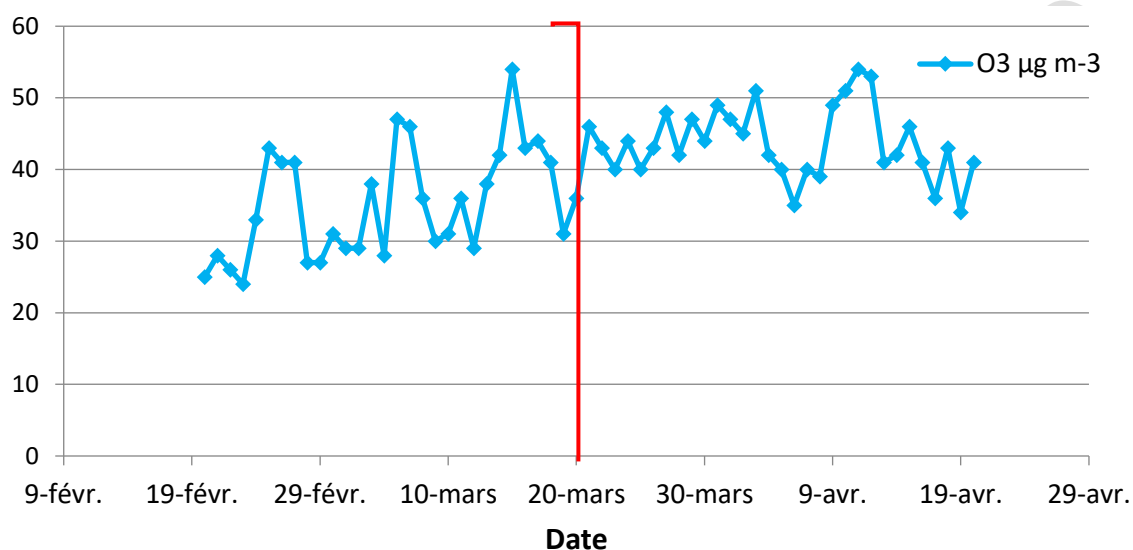
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183 **Figure 3.** Daily  $PM_{2.5}$  and  $PM_{10}$  mean concentrations ( $\mu\text{g m}^{-3}$ ) determined at PCMA monitoring  
184 station from February 20, 2020 to April 20, 2020 in Rabat - Morocco

185 The reduction of  $PM_{10}$  ( $-53\%$ ) was likely attributable to the significant reduction in vehicular  
186 traffic given by the transfers restrictions imposed by the authorities considering that transports  
187 represent the main source of  $PM_{10}$  in Rabat. The effects of reducing local air pollution on  $PM_{10}$   
188 concentrations might still be limited due to the contribution of long-range transported PM. Indeed,  
189 the back trajectory analysis portrayed that air quality in Rabat urban area is mainly affected by PM  
190 originating from the Mediterranean Basin (characterized by marine vessel emissions out of Western  
191 Europe and Northern Africa), local continental and from the near Atlantic Ocean (Othmani et al.,  
192 2020). The reduction of  $PM_{2.5}$  of  $-45\%$  can be attributed to the simultaneous reduction in the  
193 concentration of other pollutants such as  $NH_3$ , volatile organic compounds (e.g. benzene), and  $NO_x$   
194 which act as precursors to the formation of secondary  $PM_{2.5}$  (Chen et al., 2017). Similar findings  
195 were reported by Nakada and Urban (2020) and Zambrano-Monserrate et al. (2020).

### 196 3.2.3. Ozone

197 By comparing mean concentrations before and during the lockdown, we observed that higher O<sub>3</sub>  
198 concentrations occurred during the lockdown. The daily O<sub>3</sub> mean concentrations increased  
199 considerably, from 35.1 to 43.7 µg m<sup>-3</sup> which corresponds to an AVC of +20% (Fig. 4). This result  
200 is consistent with the recent literatures, for example, [Sharma et al., \(2020\)](#) reported an increase in  
201 O<sub>3</sub> in many cities covering different regions of India and [Nakada et Urban \(2020\)](#) observed an  
202 increase of approximately 30% in ozone concentrations in São Paulo.



203

204 **Figure 4.** Daily O<sub>3</sub> mean concentrations (µg m<sup>-3</sup>) determined at PCMA monitoring  
205 station from February 20, 2020 to April 20, 2020 in Rabat - Morocco  
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207 The photochemical processes related to tropospheric ozone formation have been extensively  
208 discussed in the literature. Ground-level O<sub>3</sub> is subject to in situ chemical reactions and physical  
209 processes and is directly affected by precursor emissions, temperature, solar radiation and other  
210 meteorological factors. A number of studies have examined the role of anthropogenic and  
211 meteorological factors in the formation of ozone and concluded that meteorological factors were not  
212 negligible, but anthropogenic factors were dominant ([Yu et al., 2019](#)). [Li et al. \(2019a, 2019b\)](#)  
213 found that the increasing trend cannot be simply explained by changes in anthropogenic precursor  
214 [NO<sub>x</sub> (where NO<sub>x</sub> = NO + NO<sub>2</sub>) and volatile organic compound (VOC)] emissions, this is mainly  
215 due to aerosol chemistry rather than photolysis. In the troposphere, NO<sub>x</sub> reacts with volatile organic  
216 compounds (VOCs) to form O<sub>3</sub>, and therefore, one would expect that O<sub>3</sub> would have decreased

217 along with NO<sub>x</sub> during the control period for Covid19, however, this was not the case. In fact, the  
218 daily O<sub>3</sub> mean concentrations increased considerably (+20%). The increase in O<sub>3</sub> levels in the Rabat  
219 city can be explained by different possible combined causes. Firstly, the usual increase of insolation  
220 and temperatures (~3.7°C) from February to April leads to an increase in O<sub>3</sub>. Secondly, the decrease  
221 of nitrogen oxide (-94%) reduces the O<sub>3</sub> consumption (titration, NO + O<sub>3</sub> = NO<sub>2</sub> + O<sub>2</sub>), and causes  
222 an increase of O<sub>3</sub> concentrations. Thirdly, the decrease of NO<sub>x</sub> in a VOCs-limited environment  
223 might cause urban O<sub>3</sub> to increase, as opposed to the behavior at the rural-regional background,  
224 which is mainly NO<sub>x</sub>-limited (Monks et al., 2015). Silva et al. (2017), reported that at low  
225 VOC/NO<sub>x</sub> ratios, a typical situation of polluted urban centers, the system is VOC-controlled, and  
226 reducing NO<sub>x</sub> at constant VOCs leads to an increase of ozone concentrations. Finally, the observed  
227 increases in O<sub>3</sub> during the Covid-19 controls may be related to the observed decreases in fine  
228 particles. The decrease of PM<sub>2.5</sub> (-45%) promotes ozone formation due to the role of fine particles  
229 as scavenger of hydroperoxy radicals (HO<sub>2</sub>) which would increase peroxy radical-mediated O<sub>3</sub>  
230 production (Li et al., (2019a, 2019b)). Liao and Seinfeld (2005) found that simulated surface layer  
231 O<sub>3</sub> concentrations can be reduced by 25-30% due to heterogeneous reactions on fine particulate  
232 matter using a coupled global chemistry-aerosol-climate model.

#### 233 4. Conclusions

234 As it could be expected, significant air quality improvements were observed during the lockdown in  
235 the studied area, a strong decrease in the mean concentrations of specific air pollutants was  
236 recorded, NO (up to -94%), NO<sub>2</sub> (up to -51%), PM<sub>2.5</sub> (up to -45%), and PM<sub>10</sub> (up to -53%).  
237 Overall, the significant declines in NO<sub>x</sub> and PM concentrations are mainly attributed to the  
238 containment measures against the spread of the Covid-19, which caused sharp reductions in traffic  
239 and industrial activities. However, the reduction in nitrogen oxides and particulate matter emissions  
240 did not fully eliminate air pollution, and O<sub>3</sub> actually increased (up to +20%). This result illustrates  
241 the importance of reactions that can occur between gaseous and particulate pollutants. Ozone  
242 increase can be explained by the lower titration of O<sub>3</sub> by NO due to the strong reduction in local

243 NO<sub>x</sub> emissions and by the drop in PM<sub>2.5</sub> which can lead to less trapping of hydroperoxy radicals  
244 (HO<sub>2</sub>), which would increase peroxy radical-mediated O<sub>3</sub> production.

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#### 252 **References**

- 253 Bauwens M., Compernelle S., Stavrou, T., Muller J.F., van Gent J., Eskes, H., Levelt P.F., Van  
254 Der A.R., Veefkind J.P., Vlietinck J., Yu H. and Zehner C. (2020), Impact of coronavirus  
255 outbreak on NO<sub>2</sub> pollution assessed using TROPOMI and OMI observations, *Geophys. Res.  
256 Lett.* **47**, GL087978.
- 257 Chen T.-F., Chang K.-H. and Tsai C.-Y. (2017), Modeling approach for emissions reduction of  
258 primary PM 2.5 and secondary PM 2.5 precursors to achieve the air quality target, *Atmos.  
259 Res.*, **192**, 11–18.
- 260 Cohen A.J., Brauer M., Burnett R., Anderson H.R., Frostad J., Estep K., et al. (2017), Estimates and  
261 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis  
262 of data from the Global Burden of Diseases Study 2015, *Lancet*, **389**, 1907–1918.
- 263 Collivignarelli M.C., Abba A., Bertanza G., Pedrazzani R., Ricciardi P. and Miino M.C. (2020),  
264 Lockdown for CoViD-2019 in Milan: what are the effects on air quality? *Sci. Total Environ.*  
265 **732**, 139280.
- 266 Dongarrà G., Manno E., Varrica D., Lombardo M. and Vultaggio M. (2010), Study on ambient  
267 concentrations of PM<sub>10</sub>, PM<sub>10-2.5</sub>, PM<sub>2.5</sub> and gaseous pollutants. Trace elements and  
268 chemical speciation of atmospheric particulates, *Atmospheric Environment*, **44**, 5244-5257.

269 He M.Z., Kinney P.L., Li T., Chen C., Sun Q., Ban J., Wang J., Liu S., Goldsmith J. and  
270 Kioumourtzoglou M.A. (2020a), Short- and intermediate-term exposure to NO<sub>2</sub> and  
271 mortality: a multi-county analysis in China, *Environ. Pollut.* 114165.

272 He L., Zhang S., Hu J., Li Z., Zheng X., Cao Y., Xu G., Yan M. and Wu Y. (2020b), On-road  
273 emission measurements of reactive nitrogen compounds from heavy-duty diesel trucks in  
274 China. *Environ. Pollut.*, 114280.

275 Ialongo I., Virta H., Eskes H., Hovila J. and Douros J. (2020), Comparison of TROPOMI/Sentinel-5  
276 Precursor NO<sub>2</sub> observations with ground-based measurements in Helsinki, *Atmospheric  
277 Measurement Techniques*, **13**, 205–218.

278 Klimont Z., Kupiainen K., Heyes C., Purohit P., Cofala J., Rafaj P., Borken-Kleefeld J. and  
279 Schöpp W. (2017), Global anthropogenic emissions of particulate matter including black  
280 carbon, *Atmos. Chem. Phys.*, **17**, 8681–8723.

281 Lelieveld J., Evans J.S., Fnais M., Giannadaki D. and Pozzer A. (2015), The contribution of outdoor  
282 air pollution sources to premature mortality on a global scale, *Nature*, **525**, 367–371.

283 Li K., Jacob D.J., Liao H., Shen L., Zhang Q. and Bates K.H. (2019a), Anthropogenic Drivers of  
284 2013-2017 Trends in Summer Surface Ozone in China, *Proc. Natl. Acad. Sci. U.S.A.*, **116**,  
285 422-427.

286 Li K., Jacob D. J., Liao H., Zhu J., Shah V., Shen L., Bates K. H., Zhang Q. and Zhai S. (2019b), A  
287 Two-Pollutant Strategy for Improving Ozone and Particulate Air Quality in China, *Nat.  
288 Geosci.*, **12**, 906-910.

289 Liao H. and Seinfeld J.H. (2005), Global impacts of gas-phase chemistry/aerosol interactions on  
290 direct radiative forcing by anthropogenic aerosols and ozone, *J. Geophys. Res.*, **110**, D18208

291 Menut L., Bessagnet B., Siour G., Mailler S., Pennel R. and Cholakian A. (2020), Impact of  
292 lockdown measures to combat Covid-19 on air quality over western Europe, *Sci. Total  
293 Environ.*, **741**, 140426.

294 Monks P.S., Archibald A.T., Colette A., Cooper O., Coyle M., Derwent R., Fowler D., Granier C.,

295 Kathy S. Law, Mills G. E., Stevenson D.S., Tarasova O., Thouret V., von Schneidemesser E.,  
296 Sommariva R., Wild O. and Williams M.L. (2015), Tropospheric ozone and its precursors  
297 from the urban to the global scale from air quality to short-lived climate forcer, *Atmos. Chem.*  
298 *Phys.*, **15**, 8889–8973.

299 Zhu N., Zhang D., Wang W., Li X., Yang B., Song J., Zhao X., Huang B., Shi W., Lu R., Niu P.,  
300 Zhan F., Ma X., Wang D., Xu W., Wu G., Gao G.F., and Tan W. (2020), A Novel  
301 Coronavirus from Patients with Pneumonia in China, 2019, *New England Journal of*  
302 *Medicine*, **382**, 727-733.

303 Nakada L.Y.K. and Urban R.C. (2020), COVID-19 pandemic: Impacts on the air quality during the  
304 partial lockdown in São Paulo state, Brazil, *Sci. Total Environ.*, **730**, 139087.

305 Otmani A., Benchrif A., Tahri M., Bounakhla M., Chakir E., El Bouch M. and Krombi M., (2020),  
306 Impact of Covid-19 lockdown on PM10, SO2 and NO2 concentrations in Salé City  
307 (Morocco), *Sci. Total Environ.*, **735**, 139541.

308 Pascal M., Corso M., Chanel O., Declercq C., Badaloni C., Cesaroni G., et al. (2013), Assessing the  
309 public health impacts of urban air pollution in 25 European cities: results of the Aphekom  
310 project, *Sci. Total Environ.*, **449**, 390–400.

311 Querol X., Alastuey A., Pey J., Pandolfi M., Cusack M., Pérez N., Viana M., Moreno T.,  
312 Mihalopoulos N., Kallos G. and Kleanthous S. (2009), African dust contributions to mean  
313 ambient PM10 mass-levels across the Mediterranean Basin, *Atmos. Environ.*, **43** (28), 4266–  
314 4277.

315 Sharma S., Zhang M., Anshika Gao, J., Zhang H. and Kota S.H. (2020), Effect of restricted  
316 emissions during COVID-19 on air quality in India, *Sci. Total Environ.* **728**,138878.

317 Sicard P., De Marco A., Agathokleous E., Feng Z., Xu X., Paoletti E., Diéguez Rodriguez J.J. and  
318 Calatayud V. (2020), Amplified ozone pollution in cities during the COVID-19 lockdown,  
319 *Sci. Total Environ.*, **735**, 139542.

320 Silva C. M., da Silva L. L., Corrêa S. M. and Arbilla G. (2017), Speciation Analysis of Ozone

321 Precursor Volatile Organic Compounds in the Air Basins of the Rio de Janeiro Metropolitan  
322 Area, *Rev. Virtual Quim.*, **9**, 1887-1909.

323 Tobias A., Carnerero C., Reche C., Massagué J., Via M., Minguillon M.C., Alastuey A. and Querol  
324 X. (2020), Changes in air quality during the lockdown in Barcelona (Spain) one month into  
325 the SARS-CoV-2 epidemic, *Sci. Total Environ.*, **726**, 138540.

326 Wang S., Yu S., Yan R., Zhang Q., Li P., Wang L., Liu W. and Zheng X. (2017), Characteristics  
327 and origins of air pollutants in Wuhan, China, based on observations and hybrid receptor  
328 models, *J. Air Waste Manage. Assoc.*, **67**, 739–753.

329 Wang Y., Yuan Y., Wang Q., Liu C., Zhi Q. and Cao J. (2020), Changes in air quality related to the  
330 control of coronavirus in China: Implications for traffic and industrial emissions, *Sci. Total  
331 Environ.*, **731**, 139133.

332 Yu Y., Wang Z., He T., Meng X., Xie S. and Yu H. (2019), Driving Factors of the Significant  
333 Increase in Surface Ozone in the Yangtze River Delta, China, During 2013–2017, *Atmos.  
334 Pollut. Res.*, **10**, 1357-1364.

335 Zambrano-Monserrate M.A., Ruano M.A. and Sanchez-Alcalde L. (2020), Indirect effects of  
336 COVID-19 on the environment. *Sci. Total Environ.*, **728**, 138813.