Acid and ultrasound assisted modification of boron enrichment process waste and using for methylene blue removal from aqueous solutions

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Abstract

In this study, raw boron enrichment waste was treated with hydrochloric acid and ultrasound at 35-kHz frequency for 60 minutes. To optimize the adsorption conditions for removal of methylene blue (MB) from synthetic wastewaters using raw boron enrichment waste (BEW), acid modified boron enrichment waste (HBEW) and ultrasound modified BEW (UBEW) by adsorption process and to compare the adsorption efficiency of chemical and ultrasonic modifications of BEW were aimed. The optimum adsorption conditions were determined economically and eco-friendly aspect and MB removal percents were found as 80%, 80% and 92% at optimum conditions for BEW, HBEW and UBEW, respectively. The maximum regression coefficient values were obtained as 0.911, 0.998 and 0.984 for BEW, HBEW and UBEW, respectively at Langmuir isotherm model. The adsorption rate was fitted well to pseudo-second order kinetics according to a good correlation coefficient. The adsorption of MB onto adsorbents studied is spontaneous in nature and feasible because of negative ΔG values. The results indicated that the boron enrichment process waste could be a suitable adsorbent for removal of MB from aqueous solution. The maximum adsorption capacities were equal to about 107,0 mg/g, 160,7 mg/g and 145,3 mg/g for BEW, HBEW and UBEW adsorbents at 298 K, respectively. The maximum dye removal percent was achieved for UBEW as 92% and ultrasound assisted modification was found more efficient method compared with acidic modification for MB removal.

Keywords: Adsorption, methylene blue, boron enrichment process waste, aqueous solution, adsorbents.

1. Introduction

The contamination of wastewater by synthetic dyes is an important problem (Anbia, Hariri *et al.,* 2010, Kaur, Rani *et al.,* 2015). Synthetic dyes are preferred because of color availability, easy to use, persistent, quick-setting compared with natural dyes (Chowdhury and Saha 2010; Kurniawan, Sutiono *et al.,* 2012). Large volumes of colored effluents including hazardous substances for human health and natural ecosystems generate from several industries that need to be effectively treated (Christie, 2007; Petrucci, Di Palma *et al.,* 2015). The release of these wastewaters in the environment is a significant source of visual pollution, eutrophication and serious environmental impacts because of highly toxic components. Nowadays, thermally stable dyes are introduced which are difficult to degrade after use. Therefore, suitable treatment of these wastewaters prior to discharge into receiving streams has drawn increasing attention (Panizza and Cerisola, 2008). To remove the dyes used different industries are difficult by conventional wastewater treatment plants since they are resistant to chemical oxidation, light and microbial digestion (Extremera, Pavlovic *et al.,* 2012). Various treatment techniques have been applied for the removal of dyes from aqueous solutions. Adsorption (Bousher, Shen *et al.,* 1997), electrolysis (Gupta, Ali *et al.,* 2012) coagulation (Bache, Hossain *et al.,* 1991), chemical oxidation with different material (Muthukumar, Sargunamani *et al.,* 2004), fenton (Karthikeyan, Gupta *et al.,* 2012), degradation (Saravanan, Karthikeyan *et al.,* 2013), photodegradation (Gupta, Jain *et al.,* 2011; Gupta, Jain *et al.,* 2012; Saleh and Gupta, 2012; Saravanan, Karthikeyan *et al.,* 2013; Rajendran, Khan *et al.,* 2016) are frequently employed methods for dye removal. Adsorption is most preferable method because of low energy consumption and cost, easy availability and design criteria for wide range of dye concentration and high efficiency (Meshko, Markovska *et al.,* 2001, Sanghi and Bhattacharya, 2002). Besides everything the success of the adsorption process relates primarily on the type of the dye and of the wastewater variety. A number of natural adsorbents were examined for the treatment of dyes, such as clays as an alternative to activated carbon to decrease the operation costs (Gupta, Srivastava *et al.,* 1997, Kannan and Sundaram, 2001). In recent years, various types of waste materials were applied as adsorbent for dye removal such as ash (Janoš, Buchtová *et al.,* 2003, Mittal, Mittal *et al.,* 2009), blast furnace dust (Jain, Gupta *et al.,* 2003), fertilizer waste (Gupta, Srivastava *et al.,* 1998, Jain, Gupta *et al.,*

2003), slag (Ramakrishna and Viraraghavan, 1997, Gupta, Ali *et al.,* 2003), agricultural residues materials (Mittal, Krishnan *et al.,* 2005; Mittal, Mittal *et al.,* 2010, Ahmaruzzaman and Gupta, 2011; Gupta and Nayak, 2012), waste tea (Gokce and Aktas, 2014), lubrication oil/palm waste (AlOthman, Habila *et al.,* 2014), potato plant waste (Gupta, Kushwaha *et al.,* 2016), rice husk (Sharma, Kaur *et al.,* 2010), cotton waste (Ertaş, Acemioğlu *et al.,* 2010), phosphate rock (Malash and El-Khaiary, 2010), citrus limetta peel waste (Shakoor and Nasar, 2016) and red mud (Gupta, Suhas *et al.,* 2004) etc.

The boron ores are important natural resources in Turkey. Because of Turkey has large scale of boron reserves, a large amount of boron enrichment process waste (also known as borate waste/boron containing waste) is discharged from the boron plants during enrichment process and can cause disposal problem environmentally. Boron enrichment process waste, the source of which is essentially the extractive activity (concentrator waste), has a rather broad phase composition (Cicek, Esposito *et al.,* 2012). The use of boron ore waste as an adsorbent is reasonable because of ulexite, calcite, dolomite, zeolite and some clays contents and low cost of these wastes (Olgun and Atar, 2009; Atar, Olgun *et al.,* 2011). The efficiency of the adsorption method depends primarily on the structure of the dye (Abo-Elela and El-Dib, 1987).

Modification of waste with ultrasound is an environmentally friendly method and reduces the chemicals usage, waste production and energy consumption compared with other chemical modification methods (Sujka, 2017). In this study, raw boron enrichment waste (BEW), modified boron enrichment waste with hydrochloric acid (HBEW) and modified boron enrichment waste with ultrasound (UBEW) were used as adsorbent for removal of MB from synthetic wastewater. The novelty bought by these adsorbent is the improvement in keeping dye pigments, thereby enhancing more dye removal efficiency. The process was examined for different initial adsorbent dosage, temperature, pH and agitation speed to determine the optimum equilibrium conditions.

2. Materials and methods

2.1 Adsorbents, Dyestuff and Reagents

Methylene blue is a basic, cationic dye (molecular formula $C_{16}H_{18}ClN_3S \cdot 3H_2O$ and molecular weight 373.90 g mol^{-1}) with CAS No. 61-73-4. The chemical structure of the dye is shown in Figure 1.
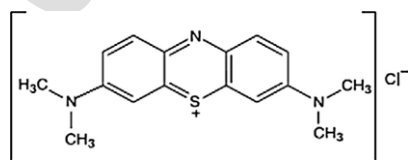

Figure 1. Chemical structure of methylene blue (Khosravi and Eftekhar 2014)

A stock methylene blue solution (1000 mg/L) was prepared and diluted with distilled water to obtain various initial concentrations for adsorption experiments. The concentration of MB in synthetic wastewater was measured on DR-2800 model UV–VIS spectrophotometer (Hach Lange, Germany) at λmax of 661 nm wavelenght.

The raw BEW, HBEW and UBEW were used as adsorbents for methylene blue removal from water in this study. BEW formed during boron enrichment process was supplied from Etibor Colemanite and Ulexite Production Plant (Eti Mine Works General Directorate Bigadiç Boron Works Management Offices, Balikesir, Turkey). BEW was dried in oven at 103 °C and then stored in desiccator until to use. Acid-modified boron enrichment waste (HBEW) was treated with hydrochloric acid for to increase the activated surface of adsorbent. The ultrasonic modification process was performed in water bath at 35-kHz ultrasonic frequency approximately 60 minutes.

Batch adsorption experiments were performed in flasks on a thermal shaker. Investigation the effect of different parameters on MB adsorption was followed according to adsorbent dose, contact time, shaking speed and pH. The dye concentrations (mg L^{-1}) at the beginning (C_0) and equilibrium (C_e) of the experiments were determined and dye removal percent was calculated according to Eq. (1):

$$\text{dye removal (\%)} = \left[\frac{(C_o - C_e)}{C_o} \right] \times 100 \tag{1}$$

The equilibrium adsorption capacity was determined by the following equation:

$$q_e = \left[\frac{(C_o - C_e)}{W} \times V \right] \tag{2}$$

where q_e is the amount of dye adsorbed at time t (mg g^{-1}adsorbent); W is the adsorbent mass (g) and V is the volume of the solution (L).

The pH measurements were performed using Hach Multi-HQ40d Instrument. A ZHWY-200B model (ZHICHENG Analytical Co., Ltd) thermal stirrer was used for the batch experiments. The solutions were filtered through 0.45-μm membrane filters (Millipore Corp., Bedford, Mass.) at the end of the experiments and then analyzed using UV-VIS spectrometer (Hach Lange DR 2800).

3. Results and discussion

3.1 Adsorbents

Chemical compositions of raw BEW, HBEW and UBEW samples were determined by using SEM spectrometer. Major chemical constituents of BEW, BEW and UBEW are given in Figure 2.

The highest porosity was observed at UBEW surface as shown in Figure 2. Surface of boron enrichment waste modified with ultrasound has more numerous and smaller pores than surfaces of BEW and HBEW. Potato, rice, wheat and corn starches surfaces were treated with ultrasound in literature and it was found that ultrasound treatment caused significant change of specific surface area of all studied starches and new pores were formed at 1.7-300 nm range (Sujka, 2017). In other study, chitin surface was modified with ultrasound and surface area of ultrasonic

modified chitin was found higher than raw chitin (Dotto, Santos *et al,.* 2015). Under high thermodynamic conditions, adsorbent surface was modified by forces to cavities formed in the liquid, supported by the hot spot theory

(Dotto, Santos *et al.,* 2015). The surface activity and adsorption capacity will increase with the increasing of surface porosity.
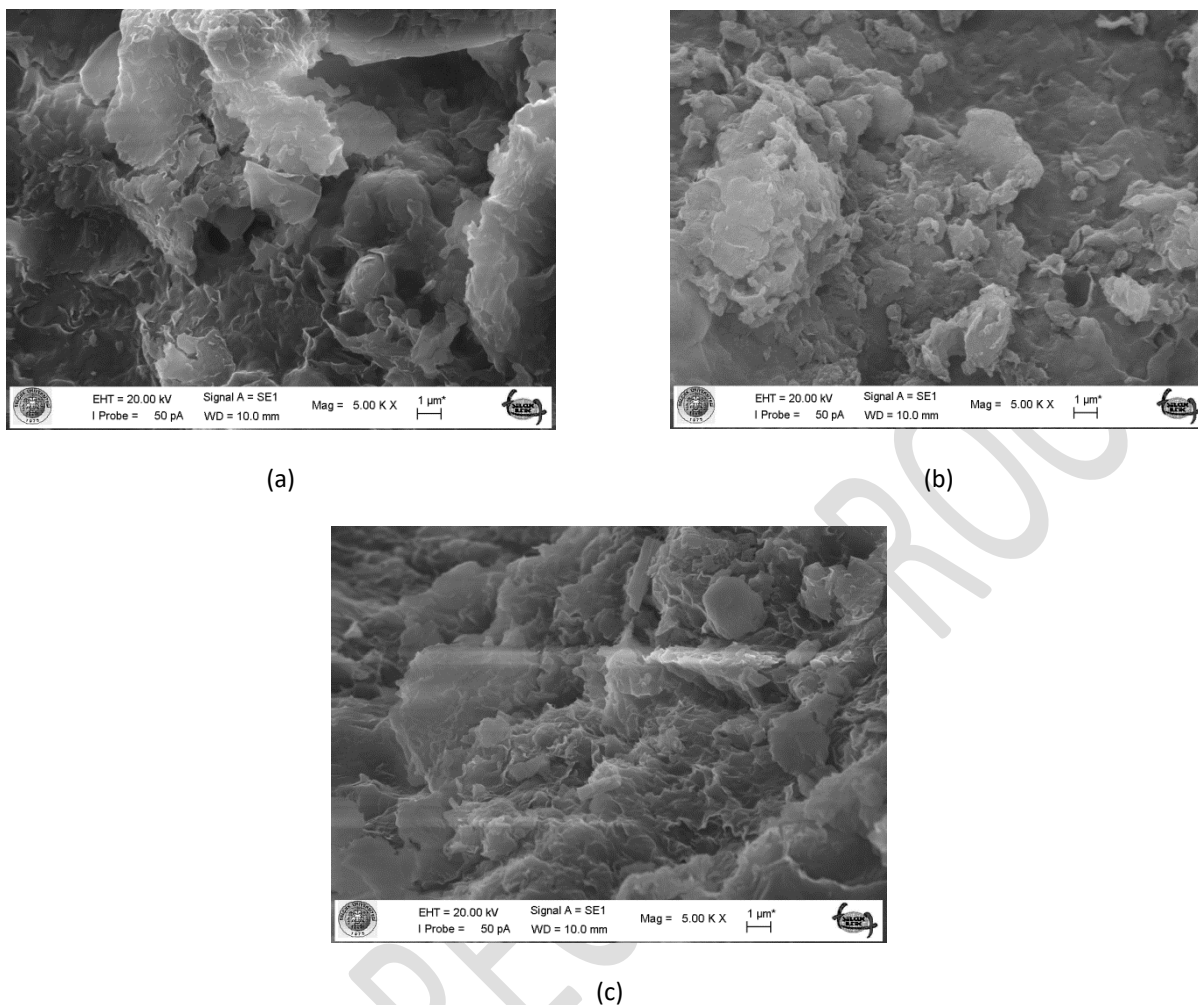

(a)

(b)

(c)

Figure 2. Scanning electron microscope images of (a) BEW, (b) HBEW and (c) UBEW
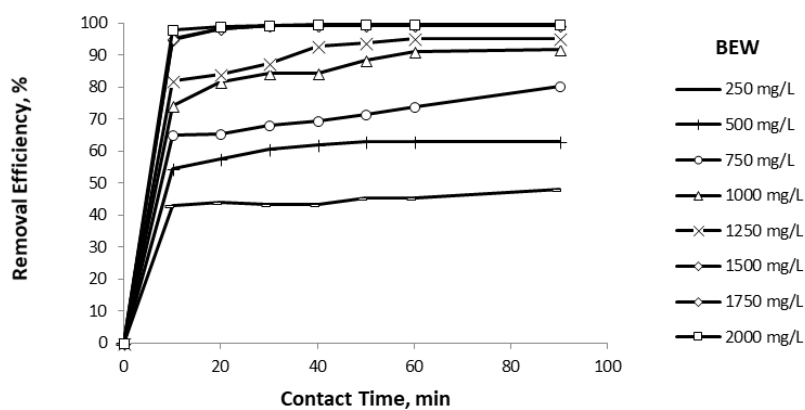
3.2 Effect of adsorbent dosage, contact time and stirring speed

MB removal was performed using BEW, HBEW and UBEW as adsorbents from synthetic wastewaters containing 100 mg/L stock MB at initial adsorbent dose 250-2000 mg/L, temperature 25 °C, original pH value 5-6.5 range, shaking speed 200 rpm and contact time 0-150 minutes. The results of experimental studies for to determine the optimum adsorption conditions at different contact time and adsorbent dose were given in Figure 3.

The optimum conditions were selected as economically aspect and applicability in field. It was obtained that color removal was 70% for BEW at 40 minute contact time and 750 mg/L initial adsorbent dose. MB removal was determined as 72% for HBEW at 30 minute contact time and 500 mg/L initial adsorbent dose. The optimum MB removal was achieved as 75% for UBEW at 40 minute contact time for 625 mg/L adsorbent dose. El Haddad, (2015) found that the maximum adsorption of Basic Fuchsin is observed at 60 min for all dye concentrations (El Haddad, 2015). In other study, the maximum removal

efficiency was obtained as 80.4% at equilibrium time 10 min and adsorbent dose 0.25 g/L in batch optimization studies (Muthukumaran, Sivakumar *et al.,* 2016). The removal of methylene blue with montmorillonite clay was obtained at 30 minutes by (Almeida, Debacher *et al.,* 2009) and the equilibrium was achieved at 120 min for remove Basic Green 4 with by Sea shell powder (Chowdhury and Saha, 2010).

The dye removal percent increased with the increasing dose of adsorbent. Similar results have also been reported in other studies and predicated to the more adsorbent surface area and availability of adsorption sites (Ertaş, Acemioğlu *et al.,* 2010). Fig. 3 indicates that the adsorption efficiency for the BEW, HBEW, and UBEW were increased rapidly first 30-40 min contact time and then process was reached to equilibrium. This may be according to the availability of abundant sites on adsorbent surface or to bind the MB at initial stages (Gobi, Mashitah *et al.,* 2011). At the same time, higher initial MB concentrations from 750 mg/L didn't affected to dye removal efficiency due to the absence of abundant sites of adsorbent surface.

(a)
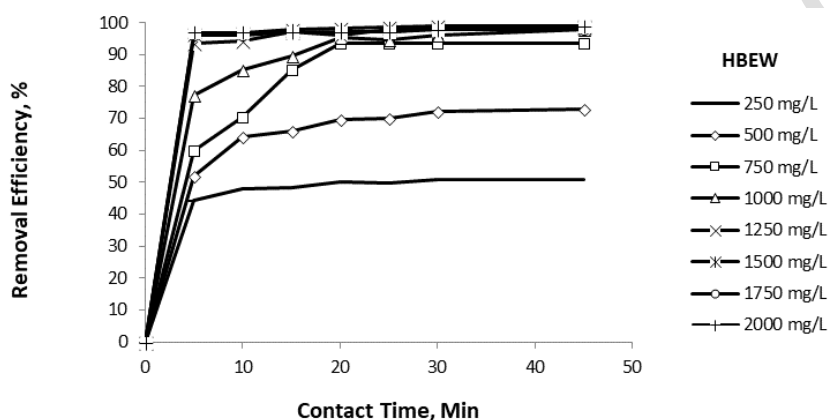
(b)
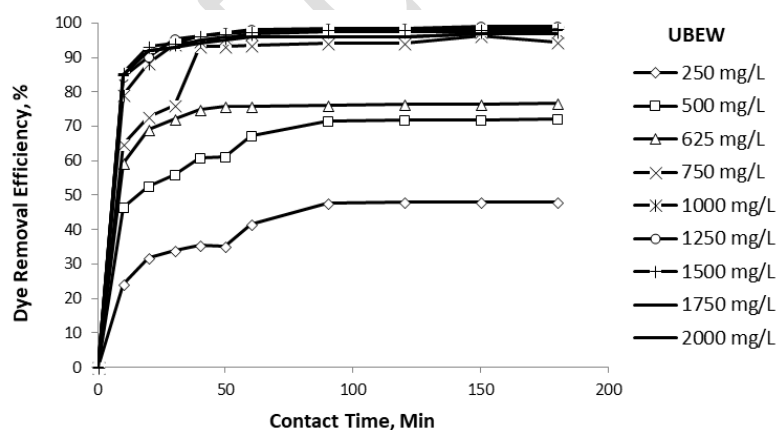
(c)

Figure 3. Effects of contact time and adsorbent dose on MB removal efficiency for a) BEW b) HBEW c) UBEW (Initial conditions = pH:5-6, temperature: 298 Kelvin, dye concentration: 100 mg/L, shaking speed: 200 rpm).

MB dye removal increased with increasing contact time and initial dye concentration in other study (AlOthman, Habila *et al.,* 2014) and it explained abundant of uncovered surface and active sites may be influenced to rapid adsorption at initial stages of batch experiments. Also the larger driving force to mass transfer resistance between liquid and dye was increased at higher initial dye concentrations (Gobi, Mashitah *et al.,* 2011).

The effects of stirring speed on the removal of MB was investigated, for this purpose batch adsorption experiments were carried out with different adsorbents using different stirring speeds in the range of 100–400 rpm at 100 mg L^{-1} initial dye concentration. The adsorption values for the BEW, HBEW, and UBEW were lowest at 100 rpm and increased when the shaking speed increased to 200 rpm, then fixed slightly as speed increased to 400 rpm.

When shaking speed was increased from 200 to 300 rpm, dye removal percent decreased at an insignificant level and the external mass transfer may be ineffective at higher stirring speed from 200 rpm. Therefore, the optimum stirring speed was accepted as 200 rpm. In literature, the flasks were agitated frequently at 150-200 rpm constant speed (Chowdhury and Saha, 2010).

3.3 Effect of pH

The initial values of synthetic wastewater for different adsorbents were adjusted in the 1-13 range. Dye removal efficiency for three adsorbents studied at different pH values were shown in Figure 4.
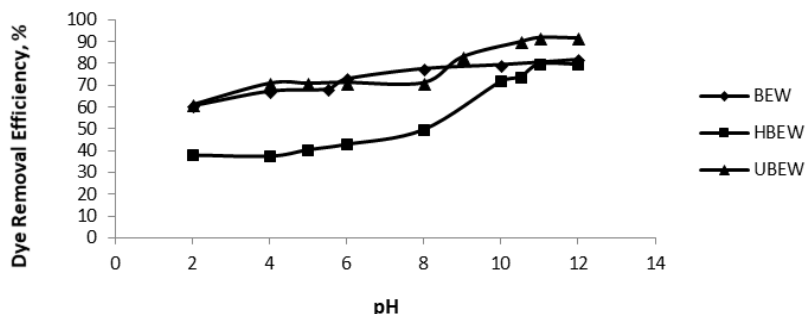

Figure 4. Effects of pH on MB removal efficiency for BEW, HBEW and UBEW (Initial Conditions; Dye Concentration: 100 mg/L, Speed:200 rpm, 298 K temperature)

pH optimization experiments were performed under the following adsorption conditions: Adsorbent dose: 750 mg/L, 500 mg/L and 625 mg/L for BEW, HBEW and UBEW respectively; shaking speed: 200 rpm, temperature: 298 K (25 °C); contact time: 40 minute, 30 minute and 40 minute for BEW, HBEW and UBEW respectively. The percents of dye removal was achieved as 77%, 79% and 91% for BEW, HBEW and UBEW adsorbents at 8, 11 and 11 pH values, respectively.

Initial pH of solution is important for MB removal because of its impact on ionization of dye molecules and active groups of adsorbent surface. MB adsorption capacity increased when the initial pH increased and it explained that formation of negatively charged groups on surface of

adsorbent at higher pH values (Sharma, Kaur *et al.,* 2010, Auta and Hameed, 2011, AlOthman, Habila *et al.,* 2014).

The optimum adsorption conditions were given in Table 1 for three adsorbents. Optimum MB removal efficiency was achieved for UBEW as 92 %. It can be seen that, BEW and HBEW show similar adsorption performance (approximately 80%). There are different approaches for acidic modification of adsorbents for MB removal performance in literature. Acidic modifications of surface of adsorbents increase oxygen content, weaken the dispersion forces and adsorption capacities. Another approach is acidic groups on adsorbent surface increase electrostatic attraction, molecular starking and locate of dye molecules vertically (Gokce and Aktas, 2014).

Table 1. Comparison of the optimum adsorption confitions of adsorbents for removal MB from aqueous media.

Adsorbent	BEW	HBEW	UBEW
Contact time, min	40	30	40
Adsorbent dose, mg/L	750	500	625
pH	8	11	11
Shaking speed, rpm	200	200	200
Temperature, °C	25	25	25
MB removal	80	80	92

3.4 Determination of adsorption isotherms

An adsorption isotherm is important to characterize the adsorbent-adsorbate relation and Langmuir, Freundlich and Tempkin isotherm models are commonly used for this purpose (Chabani, Amrane *et al.,* 2006, Serpen, Ataç *et al.,*

2007, Wang, Ying *et al.,* 2012). The adsorption of dye onto each of the three adsorbents was studied at all three temperatures, respectively 25, 35, and 45 °C. Langmuir, Freundlich and Temkin adsorption isotherm constants were calculated and given in Table 2.

Table 2. Estimated constants of Langmuir, Freundlich and Temkin isotherms for MB removal by BEW, HBEW and UBEW.

	Freundlich isotherm			Langmuir isotherm			Temkin isotherm		
	Kf	n	R^2	*b (aL)* mg/g	*KL* (L/mg)	R^2	KT	at	R^2
BEW	1,358	0,605	0,295	0,086	0,174	0,911	0,503	16,245	0,512
HBEW	39,994	0,91	0,365	0,081	20,408	0,998	44,15	1,661	0,966
UBEW	57,016	0,792	0,255	0,750	125,000	0,984	25,75	245,635	0,833

It was seen that the data fitted well to Langmuir isotherm model for the temperature and dye concentration range studied. The maximum regression coefficient values were obtained as 0.911, 0.998 and 0.984 for BEW, HBEW and UBEW, respectively at Langmuir isotherm model. This indicates that the adsorption of MB on these three adsorbents takes place as monolayer adsorption on a surface of adsorbent is homogeneous in adsorption affinity (Malash and El-Khaiary, 2010). Similar results were found in other studies (Lee, Kim *et al.,* 2011),(Tang, Zhou *et al.,* 2012). The adsorption capacity of different adsorbents is given in Table 3.

Table 3. The comparison of experimental sorption capacities of BEW, HBEW and UBEW to some sorbents for methylene blue sorption

Adsorbents	Dyes	q, mg/g	References
Chemical Modified Apricot Stones, Microwave Modified Apricot Stones	Methylene Blue	95,24 126,58	(Namal and Kalipci, 2017)
Cotton Stalk, Cotton Waste, Cotton Dust	Methylene Blue	5,95 14,04 15,78	(Ertaş, Acemioğlu *et al.,* 2010)
Palm Stem Waste/Lubricating Oil Waste	Methylene Blue	128,89	(AlOthman, Habila *et al.,* 2014)
Potato Leaves Powder, PLP	Methylene Blue, Malachite Green	52,6 33,3	(Gupta, Kushwaha *et al.,* 2016)
Potato Stem Powder, PSP	Methylene Blue, Malachite Green	41,6 27,0	(Gupta, Kushwaha *et al.,* 2016)
Palm Oil Mill Effluent Waste Activated Sludge	Methylene Blue	66,23	(Gobi, Mashitah *et al.,* 2011)
Waste Tea Activated Carbon	Methylene Blue	554,30	(Auta and Hameed, 2011)
Modified Chitin	Methylene Blue	26,69	(Dotto, Santos *et al.,* 2015)
Citrus Limetta Peel Waste	Methylene Blue	227,3	(Shakoor and Nasar, 2016)
Boron Enrichment Waste (BEW), Acid Modified BEW (HBEW), Ultrasound Modified BEW (UBEW)	Methylene Blue	107,0 160,7 145,3	This Study

3.5 Adsorption thermodynamics

For to determine the adsorption thermodynamic; ln K_c and 1/T relations of equilibrium data at different temperature (25 °C, 35 °C and 45 °C) were investigated. The results of the experiments were given in Table 4. It can be seen that the adsorption process can occur spontaneously at the normal and high temperatures, because of enthalpy change, $\Delta H° > 0$, entropy change, $\Delta S° > 0$ and Gibb's energy change ($\Delta G°$) is negative.

Table 4. Thermodynamic parameters for the adsorption of MB onto BEW,HBEW and UBEW.

	T (K)	lnb	ΔG (kJ mol^{-1})	ΔH (kJ mol^{-1})	ΔS (kJ mol^{-1} K^{-1})
BEW	298	1,417	-3,963	3,589	0,023
	308	1,437	-3,725		
	318	1,509	-3,963		
HBEW	298	1,380	-3,414	1,735	0,017
	308	1,398	-3,587		
	318	1,424	-3,760		
UBEW	298	2,070	-5,159	10,575	0,052
	308	2,242	-5,687		
	318	2,338	-6,215		

3.6 Adsorption kinetics

Pseudo-first order, Pseudo-second order and intraparticle diffusion coefficient values for MB removal from wastewaters using BEW, HBEW and UBEW were given in Table 5. Because of high correlation coefficients (R^2= 0.996 and 0.998), the data fitted well to the model of pseudo second order kinetic. qe$_{hes}$ values calculated for Pseudo-second order kinetic are similar with experimental q_e values. Pseudo-second order kinetic is the best model for cationic dye removals and this model provided a more accurate and comprehensive reflection of the adsorption mechanism, because of explains the external liquid film diffusion, surface adsorption and intra-particle diffusion processes (Fan, Tang *et al.,* 2016). This result is in agreement with many other studies of kinetics of adsorption of dyes (Özcan, Erdem *et al.,* 2004, Kumar, 2007, El Boujaady, El Rhilassi *et al.,* 2011).

Table 5. Kinetic parameters for the adsorption of MB onto BEW, HBEW and UBEW.

25 °C (298 Kelvin)	Pseudo-first-order			Pseudo-second-order			Intra-particle diffusion		
	K_1 (1/min)	R^2	qe_{hes} (mg g^{-1})	K_2 (min.g mg^{-1})	R^2	qe_{hes}	K_{int} (mg g^{-1} min$^{-0.5}$)	R^2	C
BEW	0,009	0,210	21,428	0,002	0,996	111,111	6,515	0,748	54,73
HBEW	0,006	0,228	27,797	0,001	0,996	166,666	9,277	0,687	85,12
UBEW	0,004	0,469	29,040	0,002	0,998	166,667	6,628	0,661	91,68

4. Conclusions

This study focuses on the ability of raw and different modifications of boron enrichment waste as an adsorbent for removal of methylene blue dye from aqueous solutions. The operating criterions for the optimum color removal were dye solution concentration (100 mg/L), BEW, HBEW and UBEW sorbent dosage (750 mg/L; 500 mg/L; 625 mg/L), contact time (40 min, 30 min, 40 min) respectively, and temperature (298 K). Removal of methylene blue dye is depends on pH and the maximum removal was achieved at pH 8, 11 and 11 for BEW, HBEW and UBEW, respectively. The maximum adsorption capacities were equal to about 107,0 mg/g, 160,7 mg/g and 145,3 mg/g for BEW, HBEW and UBEW adsorbents at 298 K, respectively. Equilibrium data were investigated in the Langmuir, Freundlich and Tempkin isotherm models which detected that the sorption is heterogeneous and occurred through physico–chemical interactions. The maximum regression coefficient values were obtained as 0.911, 0.998 and 0.984 for BEW, HBEW and UBEW, respectively at Langmuir isotherm model. The adsorption rate was found suited to pseudo-second order kinetics with high correlation coefficient. The adsorption of MB is feasible and spontaneous in nature because of negative ΔG values, according to thermodynamic study. The results indicated that the BEW, HBEW and UBEW could be a proper adsorbent for removal of MB from aqueous solution. It is believed that the results of this study will contribute to use of boron enrichment waste released from boron mineral processing plant in different disciplines. To use of industrial waste materials as secondary raw materials can be an alternative to the commercial activated carbon and would be beneficial for environmental sustainability.

Acknowledgments

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