Application of Fenton reagent and adsorption as advanced treatment processes for removal of Maxilon Red GRL

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Abstract
In this paper, the removal of Maxilon Red GRL by adsorption onto raw soil sample known as Niğde (Bor) grape molasses soil and Fe^2+/H_2O_2 reagent as advanced treatment processes were investigated. The effects of various experimental parameters such as initial Fe^{2+} and H_2O_2 concentrations, pH, temperature, contact time, initial adsorbent dose, and initial dye concentration on adsorption and Fenton process efficiencies were studied in a batch reactor. The adsorption experimental data were modeled by Langmuir and Freundlich isotherm models. The data fitted well with the Langmuir isotherm (R^2>0.99). The optimum conditions had been determined and found that efficiency of decolorization obtained after 20 min of reaction, was about 99.4% for Fenton process.

Keywords: adsorption, advanced oxidation process, decolorization, Fenton process, maxilon red.

1. Introduction
Wastewater effluents from different industries such as textile, food processing, cosmetics, rubber and plastics contain various kinds of synthetic dyes (Chiou et al., 2004). The textile industry is a water-consuming and heavily polluting industry (Jia et al., 2017). Therefore, potentially releasing xenobiotics from washing and rinsing procedures during finishing processes (Lofrano et al., 2016). The main sources of wastewater generated by the textile industry originate from dyeing and finishing operations (Şahinkaya, 2013; Kumari et al., 2017). The dyeing wastewater is potentially toxic and poorly biodegradable, which results in a tremendous threat to aquatic organisms (Jia et al., 2017). In this context, wastewater management receives a great deal of attention with various methods being proposed for discharge hazard estimation via ecotoxicological results (Libralato et al., 2010). Textile wastewater is characterized by high chemical oxygen demand (COD), biological oxygen demand (BOD), alkalinity and total dissolved solids (TDS) (Hayat et al., 2015). Among the different classes of textile dyes the basic ones have high brilliance and intensity of colors and are highly visible even in a very low concentration (Aksu, 2005). Disposal of dyes in precious water resources must be avoided and for that various treatment, technologies are in use (Gupta and Suhas, 2009). Various physical, chemical and biological methods such as adsorption, photolysis, chemical coagulation, chemical oxidation and reduction, electrochemical precipitation have also been used for the removal of dyes from wastewater in the past few decades (Hayat et al., 2015; Jia et al., 2017). However, all of these methods suffered with one or another limitation, and none of these were successful in removing color from the wastewater completely (Doğan et al., 2009).

Adsorption occupies a prominent place and the use of inexpensive and one of the most efficient methods of removing pollutants from wastewater and provides an attractive alternative treatment, especially if the adsorbent is inexpensive and readily available (Gupta and Suhas, 2009; Karaoglu et al., 2009). Adsorption on solid surface is being growing interest in this field because of its lower price (Nandi et al, 2009). To lower the cost of wastewater treatment, many researchers have focused on finding the low-cost alternative adsorbents (Weng et al., 2009). Recently, AOPs have received considerable attention because it is possible to degrade organic compounds and color from wastewaters (Schrank et al., 2007; Ertugay and Acar, 2017). The Fenton process is an extremely attractive option to degrade organic compounds present in wastewater due to the low-operating cost and low toxicity of the reagents. The oxidation process involves the reaction of Fe^{2+} salts with H_2O_2 to produce theradical \(^*\)OH, considered the second strongest oxidizing agent after fluorine, with a standard reduction potential of 2.8 V. In this case, Fe^{2+} ions act as a catalyst for the formation of the radical \(^*\)OH. Fenton’s reaction occurs at pH values between 2 and 4, and the reaction products are carbon dioxide, water, and inorganic salts (Almazán-Sánchez et al., 2014).

In the present study, the removal of Maxilon Red GRL (MR GRL) from aqueous solutions by adsorption onto raw soil sample known as Niğde (Bor) grape molasses soil and Fe^{2+}/H_2O_2 reagent as advanced treatment processes were studied. The effect of different parameters such as initial Fe^{2+} and H_2O_2 concentrations, pH, temperature, contact time, initial adsorbent dose, and initial dye concentration on adsorption and Fenton process efficiencies were investigated.

2. Materials and Methods

2.1. Materials

All chemicals used were of analytical grade and were used without further purifications. The basic dye was supplied from carpet industry in Niğde, Turkey. Stock solution of MR GRL dye contains 500 mg dye/L and was prepared by ultrapure distilled water. This stock solution was diluted in accurate proportions to produce solutions of different initial concentrations. The structure of the dye studied is shown in Fig. 1.

The raw soil sample known as Niğde (Bor) grape molasses soil was used for the adsorption of basic dye from aqueous solutions. This soil which consists of mostly calcite was used for sedimentation to make grape molasses. The solid sample was ground and sieved to 0.075-2 mm particle size.

![Figure 1. Structure of MR GRL dye.](image)

2.2. Analytical Methods

A calibration curve was achieved using the standard solutions of MR GRL basic dye with different concentrations in the range of 0.5–50 mg/L. The efficiency of the proposed process was evaluated by monitoring MR GRL removal after measuring absorbance at ʎ max=530 nm. After sampling the reaction was continued. The dye removal efficiency of MR GRL was calculated as follows:

\[
\% \text{ Dye removal} = \left(1 - \frac{C_e}{C_0}\right) \times 100
\]

where \(C_0\) is the initial concentration of the MR GRL basic dye, and \(C_e\) is its concentration at reaction time \(t\) (min).

2.3. Fenton Experiments

The Fenton process experiments were conducted by using jar test method. The operating parameters included dosages of \(\text{H}_2\text{O}_2\) and \(\text{Fe}^{2+}\), pH value, temperature, initial dye concentration and oxidation time. Every beaker was first filled with 200 ml of dye wastewater sample, and the pH of the reaction mixture was adjusted by adding either 1 N \(\text{H}_2\text{SO}_4\) or 1 N NaOH solution and was measured by a pH-meter (InoLab-IDS Multi 940). The dye oxidation was achieved by Fenton’s reagent which was composed of a mixture of \(\text{FeSO}_4.7\text{H}_2\text{O}\) and \(\text{H}_2\text{O}_2\) 36.5%, w/w. The necessary quantities of \(\text{Fe}^{2+}\) and \(\text{H}_2\text{O}_2\) were added simultaneously to the dye solution and the process was proceeded with rapid mixing of wastewater sample at 150 rpm for 2 min, slow mixing at 30 rpm for 20 min, and then maintaining standstill for 30 min. The supernatant samples were filtered 0.45 μm filter paper and analysed for colour analyses.

2.4. Batch Adsorption Experiments

Batch adsorption experiments were carried out 100 ml capacity of erlenmeyer with 50 ml of dye solution for different contact times by using shaker. The effects of pH (2 to 10), initial dye concentration (5-100 mg/L), amount of grape molasses soil (1-15 g/L), contact time (0-1440 min) were performed onto adsorption efficiency. The suspensions were mixed at 150 min at constant temperature (30 °C) in a shaker at 150 rpm until equilibrium was reached. The pH of the solutions was adjusted to the required value by adding either 1 N \(\text{H}_2\text{SO}_4\) or 1 N NaOH solution.

After equilibrium, the final concentration \(C_e\) was measured using absorbance values with a Hach-Lange Model spectrophotometer and compared with the absorbance value of the initial solutions \(C_0\). The absorbance was measured at 530 nm. The amount of dye adsorption per unit mass of grape molasses soil at equilibrium \(q_e\) (mg/g) was calculated by the following equation:

\[
q_e = \frac{(C_0 - C_e)V}{W}
\]

where \(V\) is the volume of the dye solution (mL) and \(W\) is the weight of adsorbent (g) added to volume \(V\).

3. Results and Discussion

3.1. Fenton Experiments

The pH of the solution is a very important parameter for Fenton process, which controls the production rate of hydroxyl radical and the concentration of \(\text{Fe}^{2+}\). Fig. 2 demonstrates the pH effect on the color removal of MR GRL by Fenton oxidation with a fixed amount of 100 mg/L MR GRL, 100 mg/L \(\text{Fe}^{2+}\) and 200 mg/L \(\text{H}_2\text{O}_2\).

![Figure 2. Effect of pH on decolorization of MR GRL during Fenton oxidation treatment.](image)

Reaction conditions: \([\text{H}_2\text{O}_2]=200 \text{ mg/L}, [\text{Fe}^{2+}]=100 \text{ mg/L}, [\text{Dye}]=100 \text{ mg/L} , t=20 \text{ min}, \text{ and temperature}=25\pm1 \text{ °C}\). To find the optimal pH of reaction mixture, a series of experiments were conducted at different pH values 2.0, 3.0, 3.5, 4.0 and 5.0. The results indicated that the degradation of MR GRL was significantly influenced by the pH of the solution and the optimal solution pH is about 3.0 giving a yield equal to 98.3%. These results agree with
several studies already carried out (Bousla et al., 2010). The color removal efficiency increases from 94.4 to 98.3% as a consequence of the pH value increase from 2.0 to 3.0 within 20 min.

To see the role of initial concentration of Fe$^{2+}$ on the yield of MR GRL decolorization by the Fenton process and to determine the most appropriate concentration of Fe$^{2+}$, a series of experiments were conducted with different concentrations of [Fe$^{2+}$]$_0$ from 50 to 150 mg/L. The results obtained are presented in Fig. 3 and the highest decolorization efficiency (98%) was recorded at [Fe$^{2+}$]$_0$ value of 150 mg/L and 20 min reaction time. Many researchers have reported that the use of a much higher concentration of Fe$^{2+}$ could lead to the self-scavenging of OH$^-$ radicals by converting it to hydroxyl ions during oxidation of Fe$^{2+}$ (Hameed and Lee, 2009).

$$\text{OH} + \text{Fe}^{2+} \rightarrow \text{Fe}^{3+} + \text{OH}^- \quad (3)$$

To determine the concentration of H$_2$O$_2$ giving the maximum MR GRL decolorization efficiency, experiments were conducted, and results obtained are represented in Fig. 4. The decolorization efficiency increased from 94.9 to 99% because of increasing H$_2$O$_2$ dosage from 100 to 500 mg/L at 20 min. This can be explained by the effect of OH$^-$ radicals produced additionally (Hameed and Lee, 2009). The study of the initial concentration effect of MR GRL dye on the efficiency of the decolorization was conducted from concentration of 50 mg/L to 500 mg/L, while maintaining pH, [Fe$^{2+}$]$_0$, [H$_2$O$_2$]$_0$, reaction time and temperature constant at 3.0, 150, 200 mg/L, 20 min and 25±1 °C, respectively. Fig. 5, shows the effect of [dye]$_0$ on its decolorization by Fenton process.

Figure 3. Effect of Fe$^{2+}$ concentration on decolorization of MR GRL during Fenton oxidation treatment. Reaction conditions: [H$_2$O$_2$]$_0$=200 mg/L, [Dye]$_0$=100 mg/L, pH=3.0, t=20 min, and temperature=25±1 °C

Oxidation of dyes by the Fenton process is carried out by OH$^-$ radicals that are directly produced from reaction between H$_2$O$_2$ that plays the role of an oxidizing agent and Fe$^{2+}$.

Figure 4. Effect of H$_2$O$_2$ concentration on decolorization of MR GRL during Fenton oxidation treatment. Reaction conditions: [Fe$^{2+}$]$_0$=150 mg/L, [Dye]$_0$=100 mg/L, pH=3.0, t=20 min, and temperature=25±1 °C

The reaction time was varied between 5 and 60 min. The initial pH, the H$_2$O$_2$ dose, the Fe$^{2+}$ dose and the initial MR GRL were kept constant at 3, 200, 150 and 100 mg/L, respectively. The optimum reaction time was determined to be 20 min. In 20 min, the supplied H$_2$O$_2$ was probably almost completely consumed as shown in Fig. 6.

Figure 5. Effect of initial dye concentration on decolorization of MR GRL during Fenton oxidation treatment. Reaction conditions: [Fe$^{2+}$]$_0$=150 mg/L, [H$_2$O$_2$]$_0$=200 mg/L, pH=3.0, t=20 min, and temperature=25±1 °C

Temperature affects the reaction between H$_2$O$_2$ and Fe$^{2+}$ and therefore, it should influence the kinetics of dyes
decolorization (Bouasla et al., 2010). Experiments were performed by varying the temperature from 25 to 50 °C. From Fig. 7, it can be seen that decolorization efficiency of MR GRL determined from 98.8 to 99% because of increasing temperature within 20 min of reaction time.

Figure 7. Effect of temperature on decolorization of MR GRL during Fenton oxidation treatment. Reaction conditions: [Fe^{2+}]=150 mg/L, [H_{2}O_{2}]=200 mg/L, pH=3.0.

3.2. Batch Adsorption Experiments

The initial pH of the aqueous solution plays an important role in the adsorption capacity (El-Sayed et al., 2013). The effect of initial pH on removal of MR GRL from aqueous solutions was investigated by varying the pH range of 2.0-10.0 at 30 °C temperature for 150 min contact time. In this study, pH did not importantly affect the dye removal and adsorption capacity (Fig. 8). Therefore, all adsorption experiments were performed at natural pH of dye solutions.

Figure 8. Effect of pH on MR GRL adsorption capacity. Experimental conditions: [Dye]=50 mg/L, [Adsorbent]=10 g/L, t=150 min.

In order to study the effect of adsorbent dose on dye removal by grape molasses soil was performed in a range of 1-15 g/L under the conditions specified and the results are given for removal efficiency (%) and the amount of removed per unit weight of adsorbent (qe) in Fig. 9. An increase in the adsorbent dose from 1 to 15 g/L resulted in a decrease of the qe from 19.4 to 3.2 (mg/g).

Figure 9. Effect of adsorbent dose on MR GRL adsorption capacity. Experimental conditions: [Dye]=50 mg/L, t=150 min.

The effect of initial dye concentration on the percentage color removal of the dye was investigated at different initial dye concentrations 5, 10, 20, 30, 40, 60, and 100 mg/L using optimum pH and adsorbent dose. Results show that increased initial MR GRL concentration led to increased MR GRL adsorption (Fig. 10).

Figure 10. Effect of dye concentration on MR GRL adsorption capacity. Experimental conditions: [Adsorbent]=10 g/L, t=150 min.

The effect of contact time ranges 0-1440 min on the adsorption of MR GRL by the grape molasses soil is shown in Fig. 11. Based on these results, the extent of removal of MR GRL by adsorbent was found to increase reach a maximum value with increase in contact time, after 120 min.

Particle size of an adsorbent played a very important role in the adsorption capacity of dye (Aljeboree et al., 2014). In order to the effect of particle size of adsorbent on the adsorption of MR GRL, adsorbent dose fixed at 10 g/L with particle size in the range 0.075-2 mm was added to dye solutions and the results are given for removal efficiency (%) in Fig. 12. It shows minimum particle size showed greater adsorption than larger size. Small size of adsorbent
increases the surface area for adsorption (Rehman et al., 2013).

According to Fig. 12, the dye removal efficiency was obtained above 90% at 10 g/L adsorbent dose, 50 mg/L dye concentration and particle size of < 0.075 mm.

Figure 11. Effect of contact time on MR GRL adsorption capacity.
Experimental conditions: [Dye]₀=50 mg/L, [Adsorbent]₀=10 g/L.

Figure 12. Effect of particle size on MR GRL Experimental conditions: [Dye]₀=50 mg/L, [Adsorbent]₀=10 g/L, T=30 °C, t=150 min.

Adsorption experiments were carried out for 50 mL dye solution containing 50 mg/L at different temperatures 25, 30, and 40 °C using 10 g/L of adsorbent in order to follow the effect of temperature on the adsorption of MR GRL (Fig. 13).

In Fig. 13, the results showed that removal efficiencies of initial dye concentration were not changed as temperature from 25 to 40 °C.

The adsorption data were analyzed using adsorption isotherm models, Langmuir and Freundlich. The linearized MR GRL Langmuir and Freundlich adsorption isotherms at different initial dye concentration (5, 10, 20, 30, 40, 50, 100 mg/L), at natural pH of dye solution, temperature 30 °C, and 10 g/L adsorbent dose were used to determine the adsorption capacity of grape molasses soil for MR GRL. The adsorption constant evaluated from the isotherms with correlation coefficients are shown in Table 1. The maximum capacity Q₀ determined from the Langmuir isotherm defines the total capacity of the grape molasses soil for the dye as 7.686 mg/g adsorbent. The fact that, compared to Freundlich isotherm, the Langmuir isotherm fits the experimental data due to the homogeneous distribution of active sites on the surface of adsorbent.

Figure 13. Effect of temperature on MR GRL. Experimental conditions: [Dye]₀=50 mg/L, [Adsorbent]₀=10 g/L, t=150 min.

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